

Comprehensive monitoring of MEA degradation in a post-combustion CO 2 capture pilot plant with identification of novel degradation products in gaseous effluents.

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Comprehensive monitoring of MEA degradation in a postcombustion CO₂ capture pilot plant with identification of novel degradation products in gaseous effluents.

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- 23

24 Abstract

25 A pilot plant campaign was performed to study MEA degradation in CO_2 capture conditions and anticipate potential degradation products emissions to the atmosphere in industrial case. Aqueous 26 27 30% wt MEA was cycled between absorption and regeneration steps during 1700 hours in the 28 presence of a synthetic flue gas containing 81% of N₂, 14% of CO₂, 5% of O₂, 97 ppm of NO, 9 ppm of 29 SO₂ and 5 ppm of NO₂. Specific methods (sampling, sample conditioning and analysis) were 30 developed to improve degradation products identification and to provide a quantification of targeted 31 compounds in liquid phase and above all in absorber and stripper gas effluents. Especially, trace 32 elements were detected in liquid phase thanks to Head Space-Solid Phase MicroExtraction (HS-SPME) 33 and liquid-liquid extraction with ChemElut cartridges and in gas phase thanks to adsorbent tubes of different type: Sep-Pak[®] for aldehydes and ketones, Orbo 60 for N-nitrosodimethylamine, Tenax and 34 35 active charcoal for non-specific adsorption. 32 degradation products were identified in liquid phase 36 and 38 in gas phase, 17 of which for the first time, especially derivatives of pyridine and oxazolidine, 37 1H-pyrrole and a new nitrosamine, the N-nitroso-2-methyl-oxazolidine.

38 Keywords

Ethanolamine degradation products in gas effluents; pilot plant monitoring; novel analytical
 methods; solid phase extraction; nitrosamines; stripper gas effluents analysis.

41 Introduction

42 Most mature post-combustion CO₂ capture technology is based on CO₂ absorption by aqueous amine

43 solutions. 2-ethanolamine (MEA) remains the benchmark solvent. When a new formulation is

- 44 developed, it is always compared to MEA in terms of energy requirement of the process, in terms of
- 45 corrosion and also in terms of degradation of the solvent. Indeed, a main problem associated with
- 46 chemical absorption using amines is degradation through irreversible side reactions mainly with CO_2 47 and O_2 but also with NO_x (Fostas et al., 2011) and SO_x (Zhou et al., 2012). These reactions can lead to
- 48 different consequences for the process: solvent loss causing cost (Carrette et al., 2013; Rao and

Rubin, 2012), formation of volatile compounds potentially unfriendly for environment (Thitakamol et al., 2007), corrosion (Lepaumier et al., 2010, Martin et al., 2012), foaming and fouling (Islam et al., 2011). It is important to list formed compounds and their concentration because some of them could be emitted to atmosphere and could be potentially a concern for human health like nitrosamines for instance (IARC, 1978; NTP, 2012; Thitakamol et al., 2007). Above all, the knowledge of their presence enables the development of adapted counter measures to avoid any risk for environment or human health.

56 Amine degradation is mainly due to oxidative degradation (Gouédard et al., 2012; Lepaumier et al., 57 2009b; Rooney et al., 1998; Sexton, 2008) but also to thermal degradation, which occurs at high 58 temperature and high CO₂ partial pressure in the stripper (Davis, 2009; Holub et al., 1998; Lepaumier 59 et al., 2008, 2009a). Oxidative degradation occurs because of a large amount of O_2 in flue gases. 60 These two mechanisms of degradation have been studied mainly under laboratory conditions, 61 sometimes on pilot plant (Reynolds et al., 2015; da Silva et al., 2012; Lepaumier et al., 2011; Strazisar 62 et al., 2003) and rarely on real CO₂ capture process. Most of the studies were focused on solvent 63 analysis. This approach is certainly the most adapted and the simplest method to study the 64 degradation of the solvent. Nevertheless, a lot of degradation products are volatile and it is 65 necessary to consider also gas effluents. Indeed, gas pollutions may have negative impacts on 66 operator but also on the environment (nature or closed population). Environmental acceptance 67 cannot be achieved without a thorough knowledge of all process effluents, gas or liquid, and 68 eventually, adapted cleaning processes. Thus, in this work a complete study of all the effluents of a 69 pilot plant was done during a dedicated campaign. This work exposes a global analytical 70 methodology which can be adapted to the monitoring of any process with any solvent. Especially, a 71 new sampling method based on solid sorbents was developed to identify trace compounds present in 72 gaseous effluents.

73 Material and Methods

74 Chemicals

75 N,N'-(Bishydroxyethyl)oxalamide (BHEOX), N-(2-hydroxyethyl)acetamide (HEA) (90 %), N-(2-76 hydroxyethyl)formamide (HEF) (97 %), , 2-methyloxazoline (≥ 99 %) and 3-methypyridine (99 %) were 77 purchased from Alfa Aesar (Schiltigheim, France). Glycolic acid, potassium acetate, sodium formate, 78 sodium propionate, sodium nitrite, magnesium sulphate, oxalic acid, lithium nitrate, potassium 79 hydroxide, formic acid, formaldehyde (37%), dimethylamine solution (40 wt%), acetamide (99%), 80 acetic acid (>99.7%), ethyleneglycol (EG) (99.8%), 1H pyrrole (98%), N,N-dimethylformamide (DMF) 81 (99.8%), pyrazine (>99%), 2-oxazolidinone (OZD) (98%), 2-methylpyrazine (2MP) (>99%), 82 diethyleneglycol (DEG) (>99%), triethyleneglycol (TEG) (>99%), 2-ethylpyrazine (>98%), 2,3-83 dimethylpyrazine (>95%), N-hydroxyethylimidazole (HEI) (97%), bicine, N-(2-84 hydroxyethyl)ethylenediamine (HEEDA), N,N-bis-2-(hydroxyethyl)ethylenediamine (BHEEDA) (97%), 85 N-(2-hydroxyethyl)imidazolidin-2-one (HEIA), diethanolamine (DEA) (>98%), 2,5-dimethylpyrazine (>98%), ethanolamine (MEA) (98 %), N-nitrosodimethylamine (NDMA), N-nitrosodiethanolamine 86 87 (NDELA), N-Nitroso-N,N-diethylamine (NDEA) and N-nitrosomorpholine (NMOR) were purchased 88 from Sigma-Aldrich (Saint Quentin Fallavier, France). Oxazoline (96 %) was purchased from Interchim 89 (France). 4-(2-hydroxyethyl)piperazin-2-one (4HEPO) and N-(2-hydroxyethyl)pyrrole (HEPyr) (99 %) 90 were purchased from Tygersci (Hopkinsville, USA). N-(2-hydroxyethyl)glycine (HEGly) (95 %) was

- 91 purchased from Enamine (Kiev, Ukraine). Glycine (99.7 %) was purchased from Merck (Lyon, France).
- 92 Oxazolidine was purchased from Selectlab (Münster, Germany). 2,6-dimethylpyrazine, 2-ethyl-6-
- 93 methylpyrazine, 2-ethyl-5-methylpyrazine and 2-ethyl-3-methylpyrazine were purchased from SAFC
- 94 (Sigma-Aldrich, Saint Quentin Fallavier, France). N,N'-Bis-(2-hydroxyethyl)urea (BHEU) was purchased
- 95 from Chemos GMBH (Regenstauf, Germany). 1-(2-Hydroxyethyl)piperazinone hydrochloride (**1HEPO**)
- 96 was purchased from Ukrorgsyntez (Kiev, Ukraine). Methanol, acetonitrile and ethylacetate were
- 97 purchased from Carlo-Erba (Val de Reuil, France).
- 98 Ultra-pure water was produced using a Direct-Q UV 3 system (18.2 MΩ.cm) from Millipore
 99 (Molsheim, France).
- 100 N-nitroso-2-methyl-oxazolidine was synthesized according to the protocol of Saavedra (Saavedra,
- 101 1981). N,N'-bis(2-hydroxyethyl)piperazin-2,5-dione (**2,5-BHEPDO**) and 2-methyloxazolidine were
- synthesized according to Gouedard *et Al.* (Gouedard, 2014).

103 Pilot plant description

- 104 IFPEN pilot plant uses two columns of around 1 meter each with a diameter of 5 cm. The first one is
- used as an absorber (C10) where the solvent (30 % wt of MEA in water) is loaded in CO_2 (Figure 1).
- 106 The second column is used as a stripper (C20) where the solvent is regenerated and CO_2 is released.
- 107 Each column is equipped with high performance packings for ensuring good gas/liquid mass transfer:
- 108 The absorber contains DX packing and the stripper contains BX packing.



109 110

Figure 1. Scheme of IFPEN pilot plant

111 The campaign lasted around 1700 hours. Synthetic flue gas contained 80.7% of N_2 , 14.1% of CO_2 , 5.1

112 % of O_2 , 97 ppm of NO, 9 ppm of SO_2 and 5 ppm of NO_2 . Total absolute pressure of the absorber was

113 1,2 bar. The gas flow rate was 1000 NL/h at 40°C and the solvent flow rate was 2 L/h. The total

volume of solvent was 20 liters. Absorber outlet gas was directed to a heat exchanger (E10) to 114 115 condensate water. Condensates was separated from gas phase in separator V10. Heat exchanger E10 followed by separator V10 may be considered as a low efficiency water wash section: E10 cooling 116 liquid temperature is 4°C and E10 outlet gas temperature is 15°C enabling large amount of water 117 condensation and therefore partial MEA and degradation products dissolution. V10 condensates was 118 119 mixed with CO_2 lean amine in tank T-01. Once loaded with CO_2 , CO_2 rich amine was pumped with PO2 120 and heated at 100°C in charge/effluent exchanger E21 before entering at the top of the stripper. The 121 stripper was heated at 125°C which enabled the regeneration of the amine and the release of CO₂. 122 Released CO_2 was directed to a heat exchanger (E20) to condensate water. The condensate was separated in separator V20 and was reintroduced at the top of the stripper with the CO_2 rich amine. 123 124 At the bottom of the stripper, the CO_2 lean amine was cooled in charge/effluent exchanger E21 then 125 in exchanger E40 to be introduced in tank T-01 at a temperature close to room temperature. CO_2 126 lean amine is pumped from T-01 with P01 and heated at 40°C to be introduced at the top of 127 absorber. MEA concentration was maintained between 30 and 35 % wt during the campaign by 128 water addition in tank T-01. At the end of the campaign, concentrations of Iron, Nickel and 129 Chromium, measured by Inductively Coupled Plasma with Optical Emission Spectroscopy (ICP-OES) 130 were less than 5 mg/kg. Thus, corrosion was assumed negligible.

131 Analysis

132 Ionic chromatography

133 Ion chromatography was commonly used to quantify acids in their anionic form: glycolate, formate, 134 acetate, oxalate, sulfate, nitrite, and nitrate ions. The concentration of CO₂-species (carbamate, 135 carbonate and bicarbonate ions and carbonic acid) was also determined by Ionic Chromatography to 136 determine CO_2 loading of amine given as mol CO_2 /mol MEA. Samples were diluted in ultrapure 137 water. 25 µL of solution were injected twice on a Dionex ICS-3000 ion chromatograph (ThermoFisher Scientific, Villebon-sur-Yvette, France). The column was an AS-15 (4 × 250 mm). Eluent was aqueous 138 139 KOH solution with a concentration gradient (from 8 to 60 mM) at a flow rate of 1 mL.min⁻¹. Detector 140 was a conductimeter. A run of 60 minutes enabled an optimal separation. Some samples were pre-141 treated with HNO₃ 69 % to reach a pH below 4 to eliminate MEA carbamate peak which overlaps 142 acetate and glycolate peaks. Quantification of anions was obtained with an uncertainty of \pm 10%. 143 Quantification of CO_2 loading was obtained with an uncertainty of ± 15%.

144 LC-MS/MS

145 Analyses were performed on a LC Thermo Scientific Dionex Ultimate 3000 (Analytical Autosampler 146 WPS-3000SL, Quaternary Analytical Pump LPG-3400SD) coupled with a MS Thermo Scientific TSQ 147 Quantum Access MAX with HESI-II source (ThermoFisher Scientific, Villebon-sur-Yvette, France). It 148 was used in positive mode, probe in position C, electrospray voltage of 2500V and capillary 149 temperature of 200°C. The sheath gas was nitrogen at a flow rate of 40mL/min and the auxiliary gas 150 at 8 mL/min. Chromatographic separations were conducted on a Thermo HyperCarb column (PGC) 151 150 mm x 3 mm, 5 µm-particles (ThermoFisher Scientific, Villebon-sur-Yvette, France). The mobile phase was a mixture of (A) water with 0.1% formic acid and (B) methanol with 0.1% formic acid at a 152 flow rate of 350μ L/min. 5 μ L of sample were injected. A pre-run rinse of 100% A for 8 min was 153 performed, then the solvent gradient started at 100% of A for 10 min before reaching a ratio of 80:20 154 155 (A:B v:v) in 8 min. This ratio was maintained for 12 min. Samples had to be at least 1000-fold diluted 156 before injection to prevent the mass spectrometer from being polluted by MEA. The MRM transitions of the targeted compounds used are given in supplementary material. They are reported
 from the literature or determined with standards. Quantification was obtained with an uncertainty of
 ±20%. Two internal standards were used: Diethanolamine-d8 for DEA, HEGly and BHEU, and Glycine d5 for Glygly and Gly. Compounds quantification method and performances are detailed in our
 previous work (Cuzuel et al., 2014).

162 **GC-FID**

163 MEA concentration was determined by GC-FID performed on Agilent HP6890 chromatograph (Agilent 164 Technologies, Massy, France) with a carbowax amines column (15 m x 0.53 mm, 1 μ m) made by 165 Supelco (Sigma-Aldrich, Saint Quentin Fallavier, France) and with triethyleneglycol as internal 166 standard. The oven program started at 100 °C then the temperature was raised at 8°C/min to 200 °C 167 and held for 5 min. 1 μ L of the sample was injected in the inlet at 250°C and the split flow was 100 168 mL/min. The Helium gas flow was 8mL/min with a top column pressure of 4.2 psi. Quantification was 169 obtained with an uncertainty of ±5%.

170 **GC-MS**

171 Analyses were performed on an Agilent 7890A gas chromatograph coupled with an Agilent 5975C 172 inert XL MSD mass spectrometer (Agilent Technologies, Massy, France). The device was equipped 173 with a MPS (MultiPurpose Sampler) auto sampler from Gerstel (RIC, Saint-Priest, France) that 174 enabled fully automated HS-SPME analyses. Two columns (Chromoptic, Villejust, France) were used 175 to separate all the target compounds, a non-polar fused silica capillary column CP-SIL8 CB-MS (30 m x 176 0.25 mm, 1µm)) and a polar fused silica capillary column DB-WAX (30 m x 0.25 mm, 0.5µm). For the non-polar column, initial temperature was 40°C held for 2 min then raised to 130°C at 7°C/min, 177 178 increased to 280°C at 13°C/min and held for 10 min. For the polar column, oven temperature 179 program started at 40°C, held for 2 min then raised to 130°C at 7°C/min, then increased to 200°C at 180 10°C/min and held for 7 min. In both cases, helium was used as carrier gas in constant flow mode at 1 mL/min. The transfer line temperature to the MS detector was set at 280°C. For liquid injection 181 182 procedures, real samples were diluted 10 times in methanol before injecting 1μ L in split mode (1:5) 183 at 250°C. Quantification of a selection of products was obtained with an uncertainty of \pm 15%.

For thermodesorption of tubes, gas flow rate of helium was 40 mL/min in splitless mode. Initial temperature of desorption was 35°C held for 2 min then raised to 300°C at 120°C/min and held for 6 min. Desorbed molecules were cryofocused in the injector at -40°C with liquid CO_2 . Then temperature increased from - 40°C to 300°C at 12°C/s and the molecules were injected in the column in splitless mode. The same GC/MS method as for liquid samples was used.

Detection was performed with a mass spectrometer using electronic ionization (EI) or chemical ionization (CI) source. The latter was both used in positive (PCI) and negative (NCI) mode. The EI source (70 eV) was heated to 250°C, the scan range was 25 to 250 amu. As for the CI source, it was heated to 300°C for positive mode and to 150°C for negative mode, CH₄ was used as reactant gas and the scan range was 50 to 250 amu.

Some complementary analyses were also performed on a Thermo Finnigan Tempus (GC-TOF (Time of flight)/MS) (Thermo Finnigan, Villebon-sur-Yvette, France), which was used with CP-SIL8 CB-MS (30 m x 0.25 mm, 1 μm) (Chromoptic, Villejust, France). Initial temperature was 35°C then raised to 230°C at 3°C/min and held for 50 min. Helium was used as carrier gas in constant flow mode at 0.9 mL/min.

198 The transfer line temperature to the MS detector was set at 250°C. Mass spectrometer was used

199 with the El source (70 eV) heated at 200°C. The scan range was 10 to 350 amu.

200 GC-HRMS

201 Analyses were performed on an Agilent 7890A gas chromatograph (Agilent Technologies, Massy, France) coupled with Autospec Premier mass spectrometer from Waters (Saint-Quentin-en-Yvelines, 202 203 France). A capillary column DB5-MS (30m x 0.25 mm, 0.50 μm) (Chromoptic, Villejust, France) was 204 used to separate target compounds. Initial temperature was 35°C held for 4 min then raised to 150°C 205 at 10°C/min, increased to 250°C at 20°C/min. Helium was used as carrier gas in constant flow mode 206 at 1.5 mL/min. Inlet temperature was 230°C, the split ratio was 1: 5 and the volume injected was 0.8 207 µL. The transfer line temperature to the MS detector was set at 260°C. Detection was performed 208 with a magnetic sector mass spectrometer using electronic ionization (EI) source in positive mode. 209 The EI source (35 eV with trap 450µA) was heated to 250°C. The mass spectrometer was used in a 210 Single Ion Recording mode (SIR) with a scan rate of 0.5 s, an interscan of 0.2 s and a resolution 211 comprised between 5000 and 6000. The recording ion mass were m/z = 74.0480 Th for NDMA and 212 m/z = 116.0590 Th for NMOR.

213 Head Space – SPME

For Head Space – SPME (HS-SPME) procedures, the volume of sample introduced in the 20 mL HS vial 214 215 was 5 mL. The fully automated HS-SPME procedure was as follows. First, the vial was equilibrated at 216 70°C during 5 min then the 75 µm Carboxen/PDMS fibre SPME, obtained from Supelco (Sigma-217 Aldrich, Saint Quentin Fallavier, France) was placed into the head-space of the sample for the extraction, still maintained at 70°C for 30 min. At the end of the extraction, the fibre was desorbed 218 219 directly in the injector set of the GC/MS at 250°C in split mode (1:5). This method was initially 220 developed for identification and quantification of pyrazines (Rey et al., 2013). It is adapted here to 221 the identification of other degradation products present in liquid phase.

222 **FT-IR**

A FT-IR disposal provided by GASMET (SISTEC, Vienne, France) was used to monitor different gas effluents (see the computer symbols on Figure 1). It enabled the monitoring of several products in gas phase, especially ammonia, CO₂, NO, NO₂ and SO₂. Limit of Detection (LOD) and uncertainties were determined by the equipment provider. The uncertainty is 2 % of the full scale. Used range and LOD are given in Table 1.

228

Table 1: Range and LOD values of FTIR analysis

Gas	Range (molar)	LOD
NH_3	0-4000 ppm	0,3 ppm
CO ₂	0-100 %	200 ppm
SO ₂	0-10000 ppm	1 ppm
NO	0-2500 ppm	4 ppm
NO ₂	0-200 ppm	1 ppm

229 Gas sampling

230 Gas sampling was performed to analyze degradation products which were not observed by FT-IR.

231 Solid phase extraction (SPE) of gaseous compounds was performed during 1 or 19 hours in function

of solid adsorbent type. During gas sampling, the gas flow in the absorber was reduced to 500 NL/h

and the solvent flow was reduced to 1 L/h. Gas samplings have been done after the separator V10

every two weeks and once after the separator V20 at the end of the campaign.

235 Assisted Liquid-Liquid Extraction (ALLE)

An original sample handling approach using assisted liquid-liquid extraction on diatomaceous earth 236 237 cartridges (ChemElut) (Agilent Technologies, Massy, France) allowed to extract selectively 238 nitrosamines from samples. This extraction on solid matrix allowed to eliminate a large amount of 239 MEA and thus enhanced the extraction of nitrosamines (improvement in sensitivity of an order 5 to 240 10). 2 mL of phosphate buffer (pH = 7.4) were mixed with 1 mL of sample. This mixture was adsorbed 241 during 5 min on ChemElut (3mL). Then, elution was performed with 9 mL of ethyl acetate. Solution 242 was smoothly evaporated with the help of a nitrogen flux gas to 0.25 mL and redissolved in 100 µL of 243 ultrapure water. Sample were then directly analyzed with LC-MS/MS.

244 Solid phase extractions (SPE)

A device enabling simultaneous adsorption on four different solid phase extraction tubes was implemented on outlet gas lines after V10 and V20. After V10 and V20, gas temperature (15°C) was below room temperature avoiding any water condensation in tubes. Gas flow rate was regulated with a flow meter in order to respect the maximum flow rate specified by tubes suppliers. For each adsorbent type, three tubes were used in series to enable quantification of the retained products in case of saturation or piercing of the two first tubes. The sampling conditions are given in Table 2.

251

Table 2: Sampling	conditions and	SPE tubes	according to	target cor	npounds

Target Compounds	Tubos Poforoncos	Gas flow	Sampling time
raiget compounds	Tubes References	(NL/h)	(h)
Aldehydes and ketones	Sep-Pak [®] DNPH	6	1
Nitrosamines	ORBO 60	10	19
Volatile compounds	Tenax TA TDU	6	1
Volatile compounds	Activated charcoal	6	1

252

Sep-Pak[®] tubes Saint Quentin 253 (Sigma-Aldrich, Fallavier, France) containing 2,4-254 Dinitrophénylhydrazine (DNPH) were analyzed by the French institute INERIS (Verneuil-en-Halatte, France). Sep-Pak[®] tubes were chemically desorbed with 2 mL of acetonitrile and products were 255 analyzed by High Performance Liquid Chromtography (HPLC) with diode array detector (DAD). 256 Aldehydes were quantified by external calibration according to the method INRS Métropol 001. 257 Excepted for formaldehyde, all aldehydes were absorbed at more than 99% in the first tube. 258 259 Quantification was obtained with an uncertainty of ±30%.

260 Thermosorb/N are the best sorbents to analyze nitrosamines because they are free of artifacts 261 formation (Roundbehler et al., 1980). However, these sorbents were not available for our study. 262 Therefore ORBO 60 tubes were used. They contain a Florisil phase (Sigma-Aldrich, Saint Quentin 263 Fallavier, France) which is particularly efficient to adsorb nitrosamines and especially NDMA which is 264 one of the most volatile and one of the most carcinogenic nitrosamine potentially present in CCS gas 265 effluents. Florisil sorbent are known to form slightly NDMA artifacts (Roundbehler et al., 1980). In this study, only NDMA was quantified. Orbo 60 tubes were chemically desorbed twice with 2 mL of 266 267 ethylacetate. Results showed that 2mL were sufficient to elute more than 95% of NDMA. NDMA was quantified by GC/HRMS. NDMA was adsorbed at more than 90% in the first tube. 268

Tenax TA TDU tubes (Gerstel, Saint-Priest, France) were thermally desorbed. Tenax is a porouspolymer widely used as an adsorbent in both air collection and purge and trap applications. Its

- structure provides alternate adsorption/desorption of a wide variety of compounds. Quantification
 of a selection of products was obtained with an uncertainty of ±30%.
- 273 Activated charcoal tubes (Gerstel, Saint-Priest, France) were used as non-specific adsorbent such as Tenax TA TDU tubes. The large surface area of this carbon form enables excellent 274 275 adsorption/desorption of a wide variety of volatile compounds. The activated charcoal tubes were 276 also thermally desorbed. Tenax TA TDU tubes enabled to adsorb larger amounts of products but 277 activated charcoal tubes were useful to confirm (or infirm) punctually the presence of some volatile 278 compounds observed with Tenax tubes. Indeed, some products detected on Tenax might be due only 279 to the degradation of the Tenax phase in gas effluent conditions and it was necessary to be able to 280 cross the results between the two different adsorbent phases.

281 Results and Discussion

282 Pilot plant monitoring

283 Gas online FTIR analyses enabled to study the composition of the gas, in terms of major compounds,

- at the different point of the pilot plantTable 3 gives CO₂, SO₂, NO and NO₂ concentrations at the inlet
- and outlet of absorber and stripper. NH_3 concentration was between 10 and 20 ppm vol at C10 exit
- 286 during the whole campaign.
- 287

Table 3: Gas composition in C10 and C20 outlet gases after condensation step (LOD: Limit Of Detection).

Gas	Inlet C10	Exit C10 dry gas	Exit C20 dry gas
CO ₂	14,1 %	6,5 %	99,99 %
SO ₂	9 ppm	2 ppm	< LOD
NO ₂	5 ppm	< LOD	< LOD
NO	97 ppm	83 ppm	50 ppm
O ₂	5,1 %	5,1 %	Not analyzed
N ₂	80,7%	88 % (calculated)	-

288

289 Based on the total inlet flow rate and assuming that nitrogen did not absorb and that O_2 290 consumption was negligible, CO_2 capture yield was around 60 % which corresponds to 282 kg of CO_2 treated in around 1700 hours. Solvent sampling was done three times a week at the bottom of the 291 292 absorber (C10) for the CO₂ rich amine and in the tank T-01 for the CO₂ lean amine. All samples were 293 stored in brown flaks between 4 and 8°C. The campaign monitoring consisted in analyzing the 294 concentrations of MEA, CO₂-species and anions. The concentration of MEA was determined by GC-295 FID enabling also an estimation of water loss (Figure 2). The pilot lost approximately 1,5 L of water per week in V10 and V20 outlet gases. Water loss was compensated by water addition (Figure 2). 296 297 14,4 kg of water were added during the campaign. Some MEA was lost also because of sampling. 298 Initially 20 kg of 30% wt MEA in water was introduced in the pilot plant. 1,3 kg of 30 % wt MEA 299 solution were therefore added during the campaign. MEA concentration was then maintained between 30% and 35% wt during the whole campaign. The concentration of CO₂-species (carbamate, 300 301 carbonate and bicarbonate ions and carbonic acid) was determined by Ionic Chromatography to 302 determine CO₂ loading of amine given as mol CO₂/ mol MEA. CO₂ loading was close to 0,3 for CO₂ 303 lean amine and 0,5 for CO₂ rich amine in stationary conditions (Figure 3). The concentration of formate, acetate, oxalate, glycolate, propionate, nitrite, nitrate and sulfate (heat stable salts) wasdetermined by ionic chromatography.





306

Figure 2. Evolution of MEA concentration in T01 (CO₂ lean amine) and added water quantities

Figure 3 shows that CO₂ loading in the solvent at the exit of absorber and stripper was quite stable at roughly 0,5 and 0,3 mol/mol respectively.







Two kinds of heat stable salts (HSS) were formed: the ones due to flue gas impurities such as NO_2 or SO₂ and the organic ones due to solvent oxidative degradation. Figure 4 shows the evolution of

nitrates, nitrites and sulfates during the campaign.

2500 Concentration (mg/kg) 2000 1500 Sulfate J. Nitrate 1000 × Nitrite 500 0 200 600 800 1000 0 400 1200 1400 1600 1800

315 316

314

Figure 4. Evolution of nitrates, nitrites and sulfates

Time (h)

317 The evolution of sulfates was almost linear from the beginning to the end of the campaign as 318 expected. The concentration of sulfates reached 2416 mg/kg in 1701 hours. It means that 94 % of the 319 total amount of injected SO₂ gave sulfates if SO₂ concentration at absorber exit is taken equal to 2 320 ppm mol which is the detection limit of FTIR. Nitrites concentration was quite constant at around 19 321 ppm from 69 to 1701 hours which may indicate that nitrites are some kind of intermediates. They 322 were formed and consumed throughout the campaign, mainly due to their oxidation into nitrates 323 which reached 383 mg/kg at 1701 hours. Nitrates and nitrites are mainly due to the presence of NO_2 324 and NO in flue gas but they may also be a resultant of MEA degradation. Table 4 gives the mass 325 balance between NO₂/NO absorbed in solvent based on FTIR results and nitrates/nitrites 326 accumulation in the solvent at the middle and at the end of the campaign.

327 Table 4: Comparison of concentration of products formed from introduced NOx with nitrates and nitrites concentration

Time (h)	Calculated concentration of products formed from introduced NOx (mol/kg)	Total concentration of analyzed nitrates and nitrites (mol/kg)	Ratio
836	4,12E-02	4,23 ± 0,4 E-02	1,0
1701	8,39E-02	1,45 ± 0,15 E-01	1,7

328

In the middle of the campaign measured nitrites/nitrates are close to NO_2/NO absorption. But at the end of the campaign, the amount of nitrites/nitrates exceeds NO_2/NO absorption because the proportion due to MEA degradation increased. On the other hand, they may be also consumed to form other degradation products like nitrosamines but this contribution is less important.

A quantitative monitoring of organic HSS was focused on formates, glycolates and oxalates (Figure 5).



334 335

Figure 5. Evolution of formates, glycolates and oxalates

336 Evolution of formates was almost linear and the concentration reached 195 mg/kg at 1701 hours. It is 337 the main organic acid formed by MEA degradation as expected (Vevelstad et al., 2013). Glycolates 338 were already present in supplied MEA and their concentration fluctuated between 60 and 100 mg/kg 339 during the campaign without significant evolution. Organic acids are known to be starting material of 340 many degradation products (Gouedard et al., 2012) which may explain such a constant concentration. On the other hand, oxalates concentration was lower than the limit of quantification 341 342 of the analytical method until 812 hours and reached 39 mg/kg at 1701 hours. Expected acetates and 343 propionates (Cuzuel et al., 2015) have not been detected at all during this campaign. Acetates are 344 assumed to have mainly reacted with MEA to form N-(2-hydroxyethylacetamid (HEA) (see table 4). 345 On the other hand propionic acid was not present neither as propionate nor as N-(2-346 hydroxyethylpropanamide). In our previous work (Cuzuel et al., 2015) both were observed but MEA was more concentrated than in the present work: 40% wt instead of 30% wt. Therefore, in present 347 348 work, degradation is slower and they might be detectable after a longer running time.

349

350 Degradation products in the solvent

351 Organic products identified in the solvent throughout the campaign are given in Table 5. The main 352 degradation products described in our previous paper (Cuzuel et al., 2015) were observed here. 353 Nevertheless, our previous pilot plant campaign was done with 40% wt MEA enabling higher 354 concentrations of degradation products. That is why less degradation products were observed in this 355 work but all should be observed after a longer running time. Each degradation product identification 356 was confirmed by corresponding standard analysis and a mechanism of formation was proposed for 357 each one (Gouedard et al., 2014a; Gouedard, 2014b; Rey et al., 2013). For each identified molecule, 358 the corresponding standard was purchased or synthesized. Identification was confirmed by obtaining the same mass or NMR spectrum. 359

Table 5: Degradation products detected in the solvent

			Analysis methods				
		GC- MS TOF	G	C-MS	LC-MS -MS	GC- HRMS	
Specific sample treatment		Ø	Ø	HS- SPME	Ø	Extrac -tion	
Products	Mw g/mol	CAS					
Ethyleneglycol (EG)	62	107-21-1		Х			
1H-Pyrrole	67	109-97-7			Х		
Oxazoline	71	504-77-8			Х		
Oxazolidine	73	504-76-7		Х	Х	Х	
N-nitrosodimethylamine (NDMA)	74	62-75-9					Х
Glycine	75	56-40-6				Х	
Pyrazine	80	290-37-9	Х		X (*)		
2-Methyloxazoline	85	1120-64-5	Х				
2-Oxazolidinone (OZD)	87	497-25-6	Х	Х	X (*)		
N-(2-hydroxyethyl)formamide (HEF)	89	693-06-1	Х	Х			
2-Methylpyrazine (2MP)	94	109-08-0			X (*)		
N-(2-hydroxyethyl) acetamide (HEA)	103	142-26-7		Х			
N-(2-hydroxyethyl) ethylenediamine (HEEDA)	104	111-41-1		Х		Х	
Diethanolamine (DEA)	105	111-42-2				Х	
2,5-dimethylpyrazine	108	123-32-0			Х		
2-ethylpyrazine	108	13925-00-3			Х		
N-(2-hydroxyethyl)pyrrole (HEPyr)	111	6719-02-4			Х		
N-(2-hydroxyethyl)imidazole (HEI)	112	3699-54-5		Х			
N-nitrosomorpholine	116	59-89-2					Х
N-(2-hydroxyethyl)glycine (HEGly)	119	5835-28-9				Х	
2-ethyl-6-methylpyrazine	122	36731-41-6			Х		
2-ethyl-5-methylpyrazine	122	236-416-6			Х		
2-ethyl-3-methylpyrazine	122	15707-23-0			Х		
N-(2-hydroxyethyl)	130	3699-54-5				х	
N-nitrosodiethanolamine (NDELA)	134	1116-54-7				Х	
1-(2-hydroxyethyl) piperazin-2-one (1HEPO)	144	59702-23-7		Х			
4-(2-hydroxyethyl) piperazin-2-one (4HEPO)	144	23936-04-1	х	х			
N,N'-bis(2-hydroxyethyl) ethylenediamine (BHEEDA)	148	3197-06-6				Х	
N,N'-Bis-(2-hydroxyethyl)urea (BHEU)	148	15438-70-7				Х	
Bicine	163	150-25-4				Х	
N,N'-(Bishydroxyethyl)oxalamide (BHEOX)	176	1871-89-2		Х			
N,N'-bis(2-hydroxyethyl)piperazin- 2,5-dione (2,5-BHEPDO)	202	23936-14-3		Х			

^{361 (*)} present in initial solution at lower concentration

Some of these products were present at a high enough concentration to monitor them quantitatively during the campaign. Diethanolamine (DEA), N-(2-hydroxyethyl)glycine (HEGly) and N,N'-Bis-(2budrequethyl)grap (DUEL)) were graptified by LC MS MS (Figure C). Thylepoplycel (FC) N (2)

364 hydroxyethyl)urea (BHEU) were quantified by LC-MS-MS (Figure 6). Ethyleneglycol (EG), N-(2-

hydroxyethyl)formamide (HEF), N-(2-hydroxyethyl)acetamide (HEA) and oxazolidinone (OZD) were 365 366 quantified by GC/MS (Figure 7). Their proportions were globally in agreement with state of art (da 367 Silva et al., 2012; Vevelstadt et al., 2013). Nevertheless, there were some differences. HEgly was not 368 the main degradation product in our case. The first reason is that this compound is an intermediate 369 to some other degradation products (Gouedard, 2014b) whose formation may be conditions 370 dependent. The second reason is that the main product in our case was BHEU which was never 371 quantified by other authors to the best of our knowledge. Its formation may be explained by a 372 reaction between MEA carbamate and MEA: nucleophilic attack of MEA amine function on 373 carbamate carbonyl group as in the case of imidazolidinones formation which are cyclic ureas 374 (Lepaumier et al., 2009a). The high concentration of both reagents may explain the high 375 concentration of BHEU.





376

Figure 6. LC-MS-MS monitoring of DEA, HEGly and BHEU during the campaign.





379

Figure 7. GC-MS monitoring of EG, HEF, HEA, 1HEPO, 4HEPO and OZD during the campaign

380 OZD concentration increased rapidly during the first 500 hours before reaching a kind of maximum at 0.9 g/L. OZD is an intermediate easily formed by carbamate cyclisation but it is also easily consumed 381 382 by reaction with nucleophilic compounds (Lepaumier et al., 2009c): it reacts especially with amine 383 functions of MEA and its degradation products. HEF was already present in supplied MEA at 0.2g/l and kept quite constant during the campaign: it is formed from MEA and formic acid (Gouedard et 384 al., 2012) and is also an intermediate to other degradation products as oxazoline for example 385 386 (Gouedard et al., 2014a). On the contrary, 4HEPO started relatively slowly but seemed to increase more rapidly after 1000 hours. Several mechanisms are proposed for 4HEPO formation (Gouedard, 387 388 2014b). In any case, starting material is a degradation product of MEA: HEEDA or HEGly. Therefore 389 their accumulation during the campaign may explain 4HEPO evolution. In this work 4HEPO 390 concentration was much higher than HEGly contrary to Da Silva et al. results (Da Silva et al., 2012) 391 suggesting that our conditions promoted HEGly consumption to form 4HEPO. 1HEPO was less 392 concentrated than 4HEPO because its formation is less favorable (Gouedard, 2014b).

393 Nitrosamines are the subject of a growing interest in the CCS literature due to their potential 394 carcinogenic properties. Three nitrosamines were identified in liquid phase in this work: N-395 nitrosodiethanolamine (NDELA), N-nitrosomorpholine (NMOR) and N-nitrosodimethylamine (NDMA). 396 N-nitroso-HEGly observed by Morken et al. couldn't be confirmed here because the corresponding 397 standard couldn't be provided (Morken et al., 2014). Some other nitrosamines were only identified in 398 the gas phase as shown below. NDELA could not be quantified accurately by our method but its 399 concentration was definitely superior to our limit of detection 50 µg/kg. The NMOR was detected but 400 its concentration was inferior to our limit of quantification of 4 µg/kg. Finally, the concentration of NDMA was determined in both CO_2 rich and CO_2 lean amine (Table 6). 401

Table 6: Concentration of NDMA in CO₂ lean and CO₂ rich amine determined by GC-HRMS

Concentration of NDMA in µg/kg	400 hours	1400 hours
lean amine (T01)	317 ± 63	541 ± 108
rich amine(C10)	265 ± 53	446 ± 89

403

404 Concentrations of NDMA in CO₂ rich and CO₂ lean amine are equivalent if higher loading of CO₂ in 405 rich amine compared to lean amine (dilution of nitrosamine) and uncertainty of the method are 406 taken into account. These concentrations are quite high comparing to other results in the literature 407 (Einbu, 2013) and that could be due to the IFPEN pilot specifically. However, NDMA is known to be 408 carcinogenic and volatile. Thus, the monitoring of the degradation products in the gas effluent in CCS 409 plant is definitely needed.

410 Degradation products in gas effluents

The analysis of degradation products in gas effluent was based on different sampling methods using 411 solid phase extraction. Tenax and activated charcoal adsorbents were used as non-selective 412 adsorbents. Orbo 60 adsorbents were dedicated to the analysis of nitrosamines. Sep-Pak[®] tubes 413 containing 2,4-Dinitrophénylhydrazine (DNPH) were used to adsorb aldehydes and ketones. These 414 415 analytical campaigns were achieved on gas effluents after the condenser V10 of the absorber and 416 after the condenser V20 of the stripper. Table 7 shows degradation products identified with Tenax, activated charcoal and Sep-Pak[®] tubes. Compounds written in bold are novel compounds which have 417 never been detected in CCS literature so far to the best of our knowledge. Some of them were also 418 419 identified in our previous work dedicated to liquid phase analysis (Cuzuel et al., 2015).

402

Table 7: Degradation products detected on Tenax, activated charcoal and Sep-Pak® tubes in gas effluents

			Absorber			Stripper	
Products	CAS	Mw (g/mol)	Tenax	Activated charcoal	Sep- Pak [®]	Tenax	Sep- Pak [®]
Formaldehyde	50-00-0	30			Х		Х
Acetaldehyde	75-07-7	44	Х		Х		Х
Dimethylamine	124-40-3	45	Х	Х			
Propionaldehyde	123-38-6	58			Х		Х
Acetamide	60-35-5	59				Х	
Acetic acid	64-19-7	60	Х				
Ethyleneglycol (EG)	107-21-1	62	Х	Х		Х	
1H-pyrrole	109-97-7	67	Х	Х		Х	
2-butenal	4170-30-3	70	Х			Х	
Oxazoline	504-77-8	71	Х	Х			
Butyraldehyde	123-72-8	72					Х
2-butanone	78-93-3	72	Х	Х		Х	Х
N,N-dimethylformamide (DMF)	68-12-2	73	Х	Х			
Oxazolidine	504-76-7	73	Х	Х		Х	
Propionic acid	79-09-4	74	Х				
N-nitrosodimethylamine (NDMA)	62-75-9	74	Х	Х			
Pyrazine	290-37-9	80	Х	Х		Х	
2-methyloxazolidine	16250-70-7	87	Х			Х	
2-oxazolidinone (OZD)	497-26-6	87	Х			Х	
2-methylpyridine	109-06-8	93	Х			Х	
3-methylpyridine	108-99-6	93	Х			Х	
2-methylpyrazine (2MP)	109-08-0	94	Х	Х		Х	
2-ethyloxazolidine	16250-71-8	101	Х			Х	
N-Nitrosodiethylamine	62-75-9	102	Х	Х			
Diethyleneglycol (DEG)	111-46-6	106	Х				
2-ethylpyridine	100-71-0	107				Х	
2,3-dimethylpyridine	583-61-9	107				Х	
3-ethylpyridine	536-78-7	107				Х	
2,6-dimethylpyrazine	108-50-9	108	Х	Х			
2-ethylpyrazine	13925-00-3	108	Х				
2,3-dimethylpyrazine	5910-89-4	108	Х			Х	
N-(2-hydroxyethyl)pyrrole (HEPyr)	6719-02-4	111	Х				
N-(2-hydroxyethylimidazole) (HEI)	3699-54-5	112	Х				
2-ethyl-2-methyloxazolidine	17026-89-0	115				Х	
2-propyloxazolidine	53019-51-5	115				Х	
N-nitroso-2-methyl-oxazolidine	39884-53-2	116	Х				
5-ethyl-2-methylpyridine	104-90-5	121				Х	
2-ethyl-3-methylpyrazine	15707-23-0	122				Х	

422 Some volatile products observed in liquid phase (some pyrazines for example) were not detected in gas phase because their concentration was likely too low. Nevertheless they must be considered as 423 424 potentially present in gas effluent of the pilot plant. On the other hand, several products were only observed in gas phase: 2-butanone, DMF, oxazolidine, 2-methyloxazolidine, 2-methylpyridine, 3-425 426 methylpyridine, N-nitrosodiethylamine, N-nitroso-2-methyl-oxazolidine and also acetamide, 427 butyraldehyde, 2-ethylpyridine, 2,3-dimethylpyridine, 3-ethylpyridine, 5-ethyl-2-methylpyridine, 2-428 ethyl-2-methyloxazolidine and 2-propyloxazolidine only observed in stripper outlet gas. Two 429 explanations may be considered: matrix effect can prevent their observation in liquid phase and/or 430 they may be formed in gaseous phase. The latter may explain specific products formed in specific conditions of stripper outlet gas (high CO₂ concentration and high temperature) unless their stripping 431 432 would be promoted by the higher temperature of stripper. Therefore, 38 degradation products were 433 identified in gas phase, 17 of which for the first time, especially derivatives of pyridine and 434 oxazolidine, 1H-pyrrole and a new nitrosamine, the N-nitroso-2-methyl-oxazolidine.

435 11 degradation products present in gas effluents were quantified in this work. Some aldehydes were
436 quantified in gas effluents with the help of DNPH Sep-Pak[®] tubes. 2 tubes were used in series.
437 Concentration profiles of these compounds in absorber outlet gas are given in Figure 8.







Figure 8. Quantification of aldehydes in absorber outlet gas.

The second tube contained some formaldehyde. Therefore, two consecutive tubes of Sep-Pak[®] with 440 441 DNPH were not sufficient to retain all the formaldehyde with a flow of 6 NL/h during 1 hour. Thus, 442 the quantification gave a minimum concentration of formaldehyde in gas effluent. Acetaldehyde was 443 the main aldehyde detected maybe because of the even higher reactivity of formaldehyde. 444 Propionaldehyde, likely formed from formaldehyde and acetaldehyde (Gouedard, 2014b) increased slowly and remained at a low level. In stripper outlet gas, formaldehyde, acetaldehyde and 445 446 propionaldehyde were roughly one order of magnitude more concentrated than in absorber outlet 447 gas (Table 8). Two explanations may be considered: their formation is promoted by stripper 448 conditions and/or their stripping is promoted by the higher temperature of stripper.

Table 8: Concentrations of some aldehydes after 1600 hours in absorber and stripper outlet gases

Aldehydes concentrations in g/L at 1600 h	Absorber	Stripper		
Formaldehyde	8	38		
Acetaldehyde	4970	48000		
Propionaldehyde	88	567		

450

451 With the help of the Sep-Pak[®] DNPH tubes, it was also possible to detect the presence of butanone 452 and butyraldehyde in the stripper outlet gas exclusively. The concentrations of butanone and 453 butyraldehyde were both 40 μ g/m³. However, neither traces of benzaldehyde nor linear carbonyl 454 compound (from five to eight carbons) were detected.

455 At the end of the campaign, three consecutive Tenax tubes were used to make a quantitative analysis

456 of ethyleneglycol (EG), pyrazine, 1H-pyrrole, dimethylformamide (DMF), 2-methylpyrazine (2MP), N-

457 nitrosodimethylamine (NDMA) and diethylenglycol (DEG) in absorber outlet gas (Table 9).

458

B Table 9: Concentration of a selection of degradation products in absorber gas effluent measured on Tenax tubes

Adsorbed quantity (ng)	Molecular weight (g/mol)	Tube 1	Tube 2	Tube 3	Total	Concentration in gas (µg/m3)
EG	62	1469	851	193	2513	419
Pyrrole	67	60	16	8	84	14
DMF	73	69	17	23	109	18
NDMA	74	1000	397	187	1584	264
Pyrazine	80	850	203	53	1106	184
2MP	94	149	0	0	149	25
DEG	106	114	57	0	171	29

⁴⁵⁹

For EG, pyrazine, pyrrole, NDMA and DMF, the third tube contained also these products which 460 461 means that three consecutive tubes of Tenax were not sufficient to adsorb all compounds with a flow 462 of 6 NL/h during 1 hour. However, the amount adsorbed on the third tube never exceeded 12% of the total amount except in the case of DMF. The cases of DMF and NDMA are suspicious because 463 both of them are formed from dimethylamine. It is likely that dimethylamine reactions were still 464 465 going on in the gas phase. Dettmer and Engewald reported the formation of NDMA from 466 dimethylamine in the presence of nitrogen oxides on such tubes (Dettmer and Engewald, 2002). 467 Therefore concentrations of DMF and NDMA might be over evaluated, especially NDMA which 468 concentration was determined at much lower values by other techniques during the same pilot plant 469 campaign.

470 Orbo 60 tubes were used to quantify NDMA. At the end of the campaign, the analysis of the gas 471 effluent was done with two consecutive Orbo 60 tubes in absorber and stripper outlet gases: 75 472 μ g/Nm³ and 60 μ g/Nm³ respectively. On one hand the value for absorber outlet gas confirms that 473 value obtained with Tenax Tubes was over evaluated. On the other hand these values are one order 474 of magnitude higher than expected values considering NDMA in equilibrium with the liquid phase 475 (Table 10). Indeed, the concentration of NDMA measured in the solvent (Table 5) and the Henry's law

449

476 constant H of NDMA in water at 25°C (2.67 X 10^{-7} atm m³/mol (EPA, 2008)) enable to calculate 477 approximately NDMA concentration in gas phase (Table 9) by using the following equation:

$$P_i = P_t y_i = H_i C_i$$

478

Table 10: Estimation of NDMA concentration in gas phase in equilibrium with liquid phase

Total pressure Pt (bara)	1,2	400 hours	1400 hours
N.	absorber	9,38E-10	1,60E-09
Y NDMA	stripper	7,84E-10	1,32E-09
C _{NDMA} in gas phase	absorber	3	5
(µg/m³)	stripper	2	4

479

Therefore, NDMA concentration values in gas phase should be considered with caution. The Tenax tubes enabled to detect in gas phase a novel nitrosamine: the N-nitroso-2-methyl-oxazolidine (NMOx). NMOx was detected by GC/MS on three different Tenax tubes and was certainly formed by nitrosation of 2-methyloxazolidine also detected on Tenax tubes. NMOx has been synthesized and

used as a standard to confirm its presence in gas effluent (Gouedard, 2014b).

485 **Conclusion**

A pilot plant campaign was performed to study MEA degradation in CO₂ capture conditions. Specific 486 487 methods were developed for sampling, sample conditioning and analysis of degradation products. 488 This analytical methodology can be adapted to the monitoring of any process with any solvent, 489 especially the new gaseous sampling method for trace compounds identification. The analysis of the 490 liquid phase confirmed degradation products identified in our previous works. On the other hand, 491 gas phase analysis went further than state of art enabling identification of 17 novel degradation 492 products. In addition to pyrazines, highlighted in our previous work, several derivatives of pyridine 493 and oxazolidine, 1H-pyrrole and a new nitrosamine, the N-nitroso-2-methyl-oxazolidine were 494 identified here among some others. This is a very important point because these products may be 495 emitted to the atmosphere. Nevertheless, the knowledge of their presence enables the development 496 of adapted counter measures to avoid any risk for environment or human health.

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