

Response of intertidal benthic microalgal biofilms to a coupled light-temperature stress: evidence for latitudinal adaptation along the Atlantic coast of Southern Europe.

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Martin Laviale, Alexandre Barnett, João Ezequiel, Bernard Lepetit, Silja Frankenbach, et al.. Response of intertidal benthic microalgal biofilms to a coupled light-temperature stress: evidence for latitudinal adaptation along the Atlantic coast of Southern Europe.. Environmental Microbiology, 2015, pp.3662-3677. 10.1111/1462-2920.12728 . hal-01097838

HAL Id: hal-01097838 https://hal.science/hal-01097838

Submitted on 9 Jan 2015

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- Response of intertidal benthic microalgal biofilms to a coupled light-temperature stress:
 evidence for latitudinal adaptation along the Atlantic coast of Southern Europe.
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- 22 Running title: microphytobenthos response to light-temperature stress
- 23

24 Summary:

Although estuarine microphytobenthos (MPB) is frequently exposed to excessive light and 25 temperature conditions, little is known on their interactive effects on MPB primary 26 productivity. Laboratory and *in situ* experiments were combined to investigate the short-term 27 joint effects of high light (HL) and high temperature (37°C vs. 27°C) on the operating 28 efficiency of photoprotective processes (vertical migration vs. non-photochemical quenching: 29 NPQ) exhibited by natural benthic diatom communities from two intertidal flats in France 30 (FR) and Portugal (PT). A clear latitudinal pattern was observed, with PT biofilms being more 31 resistant to HL stress, regardless the effect of temperature, and displaying a lower relative 32 contribution of vertical migration to photoprotection and a stronger NPQ in situ. However, 33 higher temperature leads to comparable effects, with photoinhibition increasing to about 3-34 times (i.e. from 3 to 10 % and from 8 to 22 % in PT and FR sites, respectively). By using a 35 36 number of methodological novelties in MPB research (lipid peroxidation quantification, Lhcx proteins immunodetection), this study brings a physiological basis to the previously reported 37 38 depression of MPB photosynthetic productivity in summer. They emphasize the joint role of 39 temperature and light in limiting, at least transiently (i.e. during emersion), MPB photosynthetic activity in situ. 40

42 Introduction

43 Estuarine tidal flats are one of the most productive ecosystems on Earth, largely owing to the photosynthetic productivity of benthic microalgae communities, or microphytobenthos (MPB) 44 (Underwood and Kromkamp, 1999). MPB communities are often dominated by motile 45 pennate diatoms (Haubois et al., 2005; Méléder et al., 2007; Ribeiro et al., 2013). In fine 46 sediment habitats, they undergo rhythmic vertical migration patterns in the upper layers of the 47 sediment following tidal/dial cycles (Saburova and Polikarpov, 2003; Consalvey et al., 2004; 48 Coelho et al., 2011). During daylight emersion, the upward migration of cells results in the 49 formation of transient photosynthetic biofilms which are thus periodically exposed to variable 50 and extreme environmental conditions, due to the complex interplay of timing of tidal 51 exposure and weather conditions (Admiraal 1984; Paterson and Hagerthey, 2001). Excessive 52 light exposure generates intra-cellular oxidative stress (Roncarati et al., 2008; Waring et al., 53 54 2010) which can lower photosynthetic efficiency (i.e. photoinhibition; Wu et al., 2012; Cartaxana et al., 2013) and ultimately community-level primary productivity (Blanchard et 55 al., 1996; Guarini et al., 2006). Furthermore, benthic diatom's photosynthesis often operates 56 under the combined action of high light and other potentially stressful environmental factors 57 such as extreme temperature (Blanchard et al., 1997; Serôdio and Catarino, 1999). Although 58 59 photosynthesis is known to be highly sensitive to temperature (Mathur et al., 2014), effects of high temperature on MPB productivity, either alone or in combination with excess light have 60 received little attention (Grant, 1986; Vieira et al., 2013 and references therein). However, 61 recent results showed implications on both dial and seasonal scales (Hancke et al., 2014). On 62 a longer term perspective, the cumulative effects of light and temperature could be relevant as 63 well, as climate change is expected to influence not only average values but also the 64 frequency and the intensity of extreme events such as heat waves (Schär et al., 2004; Stott et 65 al., 2004; Solomon et al., 2007). 66

The actual photoinhibitory effects depend on the photoadaptive strategy of the diatoms of 67 MPB, i.e. the operating efficiency of a range of mechanisms which include 'behavioural' and 68 'physiological' photoprotective processes (Barnett et al., 2014). The 'behavioural 69 70 photoprotection' consists of a strong negative phototaxis resulting in downward migration under supersaturating light intensities (Kromkamp et al., 1998; Perkins et al., 2010; Cartaxana 71 et al., 2011; Serôdio et al., 2012). The vertical migration of cells within the steep light 72 gradient of the sediment photic zone (Paterson and Hagerthey, 2001; Cartaxana et al., 2011) 73 74 has been hypothesized to allow for the optimization of light harvesting and the avoidance of excess light (Admiraal, 1984). Strong evidence for a photoprotective role of vertical migration 75 was provided by the use of the diatom motility inhibitor Latrunculin A (Lat A) on undisturbed 76 biofilms (Perkins et al., 2010; Cartaxana et al., 2011; Serôdio et al., 2012). 'Physiological 77 photoprotection' mainly includes the excess energy-dissipative non-photochemical quenching 78 79 of chlorophyll (Chl) a fluorescence (NPQ). NPQ is mainly controlled by the presence of the xanthophyll pigment diatoxanthin (DT) and of Lhcx proteins in the light-harvesting 80 81 complexes of photosystem II (PSII) (Depauw et al., 2012; Lepetit et al., 2012, 2013; Lavaud and Goss, 2014). In diatoms, DT is produced by the (dark/low light) reversible light-82 dependent conversion of the xanthophyll diadinoxanthin (DD) in the so called xanthophyll 83 84 cycle (XC) (Goss and Jakob, 2010). Besides its involvement in NPQ, it seems to have another function as antioxidant (Lepetit et al., 2010). 85

Despite the crucial role of NPQ and XC in the photoprotective response of diatoms to
excessive fluctuating light regimes, especially in field situations (Brunet *et al.*, 2010; Lavaud
and Lepetit, 2013; Lavaud and Goss, 2014), they have only scarcely been studied in MPB.
They show a relation to diurnal and tidal cycles, season, latitude (Serôdio *et al.*, 2005; van
Leeuwe *et al.*, 2008; Chevalier *et al.*, 2010), to vertical position of diatoms within the
sediment (Jesus *et al.*, 2009; Cartaxana *et al.*, 2011) or along the intertidal elevation gradient

(Jesus et al., 2009), with significant differences between the main growth forms (i.e. epipelic 92 and epipsammic: Cartaxana et al., 2011; Barnett et al., 2014). As for the implication of Lhcx 93 proteins in diatom NPQ, its characterization has been limited to few model species (Bailleul et 94 95 al. 2010; Zhu et al., 2010). Lhcx1 is involved probably at least by binding DT (Bailleul et al., 2010; Büchel, 2014). It is present at low light and its transcript level/synthesis is slightly 96 enhanced under high light (HL) (Bailleul et al., 2010; Wu et al., 2012; Lepetit et al., 2013). 97 While the transcript levels/synthesis of other examined isoforms (i.e. Lhcx2, Lhcx3, Lhcx6) 98 99 strongly increase during HL (Zhu et al., 2010; Lepetit et al., 2013) and temperature stress (for Lhcx6; Wu et al., 2012), their exact physiological role remains unknown, although Lhcx3 and 100 Lhcx6 are suspected to be involved in the binding of *de novo* synthesized DT and/or the 101 sustain part of NPQ (i.e. qI) during a prolonged HL stress (Zhu et al., 2010; Lepetit et al., 102 103 2013).

104 This work intended to study the combined effects of high light (HL) and high temperature on 105 the photophysiology of intertidal diatom-dominated MPB communities. Light stress-recovery 106 experiments (LSRE) were performed to distinguish physiological photoprotection efficiency 107 from photoinhibitory effects, and to quantify behavioural photoprotection (Serôdio et al., 2012). Until now, these processes have been characterized only to the extent of chlorophyll 108 fluorescence measurements and XC pigments analysis. The present study combines 109 chlorophyll fluorescence imaging, pigment analysis and for the first time on intact biofilms, 110 lipid peroxidation quantification and immunodetection of Lhcx proteins. LSRE were 111 performed on MPB from two locations in France and Portugal with different solar exposure 112 and temperature regimes, likely to support the establishment of diatom communities with 113 contrasting photo- and thermo-adaptive strategies. Moreover, the potential role of temperature 114 115 in modulating the HL-induced stress was investigated by carrying out LSRE at two experimental temperature conditions ('optimal' vs. high temperature). In parallel, the 116

- estimated photosynthetic and protective indices were further compared with field situation at
- the two sites during a typical summer diurnal emersion.

120 **Results**

121 Environmental conditions, in situ photosynthetic activity and taxonomic composition of
122 microphytobenthos in PT and FR sites

Environmental conditions and MPB photosynthetic activity were assessed during typical 123 summer diurnal emersion at both Portuguese (PT) and French (FR) intertidal flats. The 124 Photosynthetic Active Radiation (PAR) reached its daily maximum (2000 µmol photons m⁻² s⁻ 125 ¹) at both sites (Fig. 1A-B) with a comparable light dose over the emersion (~19 to 23 mol 126 photons m⁻², Table S1). Temperature was higher at PT site (Fig. 1A-B). Average temperatures 127 at the sediment surface were 37.5°C and 28.8°C for PT and FR sites, respectively. 128 Temperature decreased with sediment depth with a stronger effect in PT site (-7°C and -6°C 129 in the first 0.5 cm for PT and FR sites, respectively: Table S1, Fig. 1A-B). 130

At the beginning of emersion, the effective PSII quantum yield (Φ PSII) was close to its 131 132 optimum at both sites (~ 0.70: Fig. 1C-D, open squares). It decreased/increased opposite to the PAR evolution over the emersion period, reaching values as low as ~ 0.40 . The maximal 133 non-photochemical quenching of Chl a fluorescence (NPQ_{max}) was about 4-5 at PT site and 134 lower than 1 at FR site at the beginning of emersion (Fig. 1C-D, closed circles). It decreased 135 continuously at PT site to reach values around 3 at the end of the emersion, while it covaried 136 137 with PAR at FR site, reaching a maximum value of about 6 (matching the sudden increase in PAR level) and values of 3-4 at the end of the emersion. 138

At both sites, MBP communities were dominated by epipelic (i.e. motile) benthic diatoms of
the genus *Navicula* (Table S2). *Navicula phyllepta* sensu lato (Vanelslander *et al.*, 2009) was
the dominant species at PT site (relative abundance: 44 %) while FR site was co-dominated
by *N. spartinetensis* (33 %) and *N. phyllepta* s.l. (28 %).

The two temperature treatments used for the laboratory experiments during the LSREs were 27.4 \pm 1.2°C and 37.3 \pm 0.7°C (thereafter named 27°C and 37°C temperature treatments, respectively), which was close to the average temperatures monitored at the sediment surface for both sites (see above and Table S1). The light exposure of 3 h of 1200 µmol photons m⁻² s⁻¹ ¹ represented a light dose of 12.96 mol photons m⁻² (Serôdio *et al.*, 2012). This light dose was reached after 2 h 53 min and 3 h 34 min in PT and FR sites, respectively (Table S1).

151

152 LSRE-induced photoinhibition and NPQ in sediment from FR site

 Φ PSII remained constant in samples maintained under LL throughout the experiment (0.66 ± 153 0.02, Fig. 2). In comparison, samples exposed to HL exhibited dramatic changes over time 154 (time: P < 0.001, see Table S3a for detailed statistical outputs). After 3h HL, Φ PSII decreased 155 down to values between 0.13 ± 0.03 (37°C +Lat A) and 0.34 ± 0.02 (27°C Ctrl). It increased 156 157 during LL recovery but did not reach back its initial level (0.46 ± 0.05 ($37^{\circ}C$ +Lat A) and 0.61 ± 0.01 (27°C Ctrl). The differences observed between temperature and chemical 158 159 treatments were significant (temperature: P < 0.001; chemical: P < 0.01; Table S3a), and they 160 both increased with time (time \times temperature: P < 0.001; time \times chemical: P < 0.05; Table S3a). Similarly, NPQ of samples maintained under LL remained constant (0.38 ± 0.04 , Fig. 3) 161 and changed over time according to light (time: P < 0.001, Table S3b): it increased under HL 162 (between 4.4 ± 0.5 and 3.4 ± 0.4 from $37^{\circ}C$ +Lat A and $27^{\circ}C$ Ctrl treatments, respectively) 163 but only partially relaxed during the subsequent 15 min LL incubation (between 1.9 ± 0.1 and 164 1.1 ± 0.1 ; Fig. 3). The effect of temperature was highly significant (temperature: P < 0.001, 165 Table S3b) and the observed differences increased with time (time \times temperature: P < 0.001; 166 Table S3b). In comparison, the effect of lat A was less significant (chemical: P < 0.05 and 167 168 time \times temperature: P = 0.11; Table S3b).

170 LSRE-induced allomerization of Chl a and xanthophyll cycle in sediment from FR site

All samples exhibited typical diatom pigment signatures, including Chl a, Chls c, 171 fucoxanthin, the carotenoid β , β carotene (data not shown) and the xanthophyll cycle pigments 172 DD and DT. The allomer form of Chl a (Chl a-allo) accumulated during the LSRE as 173 illustrated by the Chl *a* allomerization molar ratio (AR= Chl *a*-allo / total Chl $a \times 100$, Fig. 174 4A). AR increased ~ 2-fold under HL and remained stable thereafter (time: P < 0.001; Table 175 S3c). No significant changes were observed with temperature (temperature: P = 0.48; Table 176 177 S3c), whereas AR was significantly higher in +Lat A samples than in Ctrl (chemical: P < P0.01; Table S3c), with mean maximum values of $11.9 \pm 5.0\%$ and $8.4 \pm 2.1\%$, respectively 178 (Fig. S1). AR did not significantly change in samples maintained in LL during 3 h (5.9 \pm 179 2.7% on average). 180

The pool of DD+DT (in mol. 100 mol Chl a^{-1}) increased ~ 3-fold under HL and remained 181 182 stable during the LL recovery (Fig. 4B, time: P < 0.001; Table S3d). Overall, it reached higher values at 37°C than at 27°C (temperature: P < 0.01; Table S3d) but it was comparable 183 in Ctrl and +Lat A samples (P = 0.58; Table S3d). At 37°C, DD+DT increased even in the 184 samples maintained under LL for 3 h (from 15.0 \pm 4.0 to 23.8 \pm 6.6 mol. 100 mol Chl a^{-1} , 185 Tukey HSD: P < 0.05, Fig. S2). DD de-epoxidation (DES = DT / DD+DT × 100) correlated 186 well with light whatever the temperature in both Ctrl and Lat A-treated samples (P < 0.001) 187 (Fig. 4C-D). DES increased under HL and decreased during LL recovery while it remained 188 stable in samples maintained in LL for 3 h (14.0 \pm 3.2% on average) (Fig. S3). DES was 189 significantly higher in +Lat A samples than in Ctrl ones (chemical: P < 0.001; Table S3d), and 190 at 37°C than at 27°C (temperature: P < 0.001; Table S3d) with maxima of 67.2 \pm 7.5% and 191 $33.8 \pm 7.5\%$ at 37 °C and 27 °C, respectively, in +Lat A samples (Fig. S3). DES recovered 192 much faster at 27°C than at 37°C, especially for Ctrl samples. In contrast to AR and the 193 DD+DT pool, interaction between time and temperature was significant for DES (time \times 194

temperature: P < 0.001; Table S3d). Moreover, NPQ and DES were significantly positively correlated according to the chemical treatment (Ctrl/+Lat A), *i.e.* the slope of the NPQ *vs*. DES linear regression was 2.5 lower in +Lat A samples (Fig. S4) while for both treatments the origin of the regression was similar and non-nul (Ctrl: 17.6 ± 0.4 ; +Lat A: 16.4 ± 0.4), i.e. some DT was synthesized without NPQ development.

200

201 LSRE-induced Lhcx synthesis in sediment from FR site

202 The anti-FCP6 (i.e. Lhcx1) revealed a clear Lhcx isoform in FR samples (Fig. 5A). Its size of about 24 kDa was slightly higher than the one of Lhcx 3 in Phaeodactylum tricornutum and 203 204 its amount increased after 3 h HL exposure especially in +Lat A samples. Interestingly, N. phyllepta s.l., the dominant species of the MPB assemblage, showed an Lhcx pattern similar 205 to P. tricornutum so that three isoforms (Lhcx1, Lhcx2 and Lhcx3) could be identified. The 206 207 anti-Lhcx6 from Thalassiosira pseudonana revealed two Lhcx isoforms in FR samples, the size of which was different from Lhcx6 and was neither present in HL exposed P. tricornutum 208 209 cells (Fig. 5B). N. phyllepta showed two isoforms, one which size was slightly higher than 210 Lhcx6 and the other one with a similar size to the 33-50 kDa isoform from FR samples. With such size, this isoform could be a hypothetic Lhcx dimer (given that the available sizes of 211 Lhcx monomers deduced from the genomes of P. tricornutum, T. pseudonana and 212 Fragilariopsis cylindrus range between 21 and 29 kDa) but different than in P. tricornutum. 213 The 17-23 kDa isoform showed a size similar to Lhcx 1 in P. tricornutum and N. phyllepta 214 (see Fig. 5A). Interestingly, the two isoforms reacted to light, temperature and Lat A in 215 216 different ways: while the 17-23 kDa isoform appeared not to react to any of the treatments, the 33-50 kDa isoform positively reacted to HL and 37°C with an effect enhanced by Lat A at 217 218 37°C (Fig. 5C).

220 LSRE-induced lipid peroxidation in sediment from FR site

221 Common markers of membrane lipid peroxidation (TBARS-thiobarbituric acide reactive 222 substances) were quantified. In Lat A treated samples, TBARS content normalized to the 223 surface microphytobenthos biomass (I_{diat} biomass index) did not change significantly over 224 time and was not significantly affected by chemical or temperature treatments (P = 0.71, 225 0.99 and 0.19, respectively; Table S3f), exhibiting a mean value of $1.19 \pm 0.29 \ 10^{-6}$ nmol eq 226 MDA mL⁻¹ (Fig. 6).

227

228 *Comparison between FR and PT sites: NPQ, photoinhibition, lipid peroxidation and Lhcx.*

 Φ PSII and NPQ measured on PT samples at T0 (0.67 ± 0.02 and 0.40 ± 0.05, respectively) 229 were comparable to the one measured on FR ones (Fig.2 and 3: 0.66 ± 0.01 and 0.39 ± 0.01 , 230 respectively), and they exhibited the same general trend under HL stress. However, PT 231 232 samples were less affected by the HL treatment, as Φ PSII recovered faster under LL (site: P < 0.001, Table S4a), regardless of temperature or Lat A application (Fig. 7A). Moreover, NPQ 233 234 did not relax entirely, but the level of this sustained NPQ was lower than in FR samples (site: 235 P < 0.01, Table S4b), although post-hoc multiple comparison tests did not show any statistical differences between sites (Tukey HSD: P > 0.05; Fig. 7B). 236

For each site and temperature, the extent of photoinhibition and the contribution of vertical
migration to the overall photoprotection capacity of MPB were estimated as in Serôdio *et al.*(2012) (Table 1). Photoinhibition was higher in FR than PT site but it increased similarly (~3
fold) with temperature. The contribution of vertical migration to photoprotection remained
below 15 % but was always lower in PT biofilms (< 5%).

In Lat A treated samples, TBARS content normalized to surface biomass did not change significantly at PT site whatever the temperature (Fig. S5). Mean value was $3.26 \pm 0.66 \ 10^{-6}$ nmol eq MDA mL⁻¹, which was significantly higher than for FR samples (t-test: P < 0.001).

While the anti-FCP6 (i.e. Lhcx1) revealed two clear Lhcx isoforms in PT samples (Fig. 8), the 245 anti-Lhcx6 did not bring any signal probably due to its unspecificity as regards to the species 246 247 composing the microphytobenthic community of PT site. Their respective size were close with one at about 23 kDa, the size of Lhcx 3 in P. tricornutum and N. phyllepta, and the other 248 one slightly higher (about 24 kDa) which was also present in FR samples (Fig. 5A). Both 249 isoforms (although the 24 kDa reacted in a stronger way) were enhanced after 3 h of HL 250 exposure as well as by the higher temperature; this effect was increased by the addition of Lat 251 252 A. Interestingly, at 27°C, the 24 kDa isoform was not present under LL and HL exposure, it appeared with the addition of Lat A solely (LL 27°C +) and its content was increased by HL 253 exposure (HL 27°C +). Similarly, the 23 kDa isoform positively reacted to Lat A to a level 254 even similar to HL exposure without Lat A (compare 27°C LL+ and 27°C HL-). 255

257 **Discussion**

In this study, higher temperature significantly increased photoinhibition susceptibility of 258 epipelic MPB biofilm. Communities from the Portuguese (PT) site were clearly more resistant 259 260 to a coupled high light-temperature stress, in comparison to biofilms from the French (FR) site. At both site, photoprotection was based on the interplay between physiological (NPQ, 261 DT and Lhcx proteins) and behavioural (vertical motility) response, nevertheless PT biofilms 262 displayed a lower relative contribution of vertical migration and a stronger NPQ. The 263 apparent latitudinal pattern observed at the community level ('MPB biofilm') might be driven 264 by the individual response of the respective dominant diatom species/ecotypes in each site. 265 These different aspects are discussed below. 266

267

268 Evidence for a latitudinal adaptation in the response of epipelic MPB to a coupled light-269 temperature stress

Our results suggest that PT MPB communities were adapted to a higher light and/or 270 271 temperature environment than FR ones. Overall, PT samples were more resistant than FR 272 ones to HL stress, regardless the effect of temperature. At 27°C, PT control samples recovered almost entirely from HL exposure (97 %), and photoinhibition increased only from 273 3 to 10 % at 37°C. This trend was also obvious in situ: the time course and level of ΦPSII 274 275 were comparable between sites despite the potentially more stressful conditions at PT site (similar light conditions but 8-9°C higher average surface temperature). One of the reasons 276 277 for this differential latitudinal adaptation might lie in the photoprotective strategies the two MPB communities exhibited. PT MPB communities showed higher NPQ_{max} average values, 278 an indication of a strong(er) physiological photoprotective response (Serôdio et al., 2005). PT 279 280 NPQ_{max} was already high right from the beginning of the emersion although light intensity was still moderate and it remained relatively stable along the emersion. Consequently, vertical 281

migration contributed less to the overall photoprotection capacity (< 5 %) in comparison to 282 FR biofilms. FR NPQ_{max} tended to follow the fluctuations of light intensity along the 283 emersion which would be in agreement with a more dynamics MPB response, regarding both 284 motility and physiology. Only few studies have examined the link between seasonal and/or 285 latitudinal photoacclimation and the coupling between behavioural and physiological 286 photoprotection in MPB communities. A similar increase of migration contribution was 287 reported for communities acclimated to lower irradiances (i.e. seasonal acclimation: Serôdio 288 et al., 2012). Another study provided evidence for the differential role of migration vs. XC 289 along a latitudinal gradient (Ireland, UK and Portugal), although the role of sediment type (i.e. 290 291 mud vs. sand) could not be ruled out (van Leeuwe et al., 2008).

In contrast, the contribution of thermal adaptation is difficult to ascertain. Although PT 292 samples were more resistant and resilient, higher temperature leads to comparable effects: 293 294 photoinhibition increased to about 3-times (i.e. from 3 to 10 % and from 8 to 22 %, in PT and 295 FR sites, respectively) and lipid peroxidation was comparable. The pattern of the detected 296 Lhcx isoforms was also similar, although no quantification was possible. Average surface and 297 -0.5 cm temperature conditions experienced by the FR and PT communities were clearly different. The fact that this difference was lower than the observed amplitude over an 298 emersion could explain the apparent lack of temperature latitudinal pattern. This is in 299 300 agreement with other comparative studies on short-term (i.e. over few hours) temperature effects on MPB photosynthesis (Grant, 1986; Blanchard et al., 1997; Hancke and Glud, 301 302 2004). Instead, the different average temperature conditions at PT and FR sites recorded in 303 this study, which are representative of typical summer conditions in these areas (Guarini et al., 1997; Lillebo et al., 2010), might have played a role on a longer time scale. As 304 305 temperature and light adaptations rely on similar strategies (Davison, 1991; this study, see below), it is likely that the difference in average temperatures influenced the susceptibility of 306

307 MPB to photoinhibition, directly and/or via its photoacclimation status, together with e.g.
308 daily light dose, day length and/or UV exposure.

The latitudinal pattern in ecophysiological characteristics observed at the community level 309 310 coincided with a shift in the diatom species composition. Although it is difficult to separate 'community' response from 'individual species' response within the biofilm, it is possible that 311 the observed responses at each site were driven by the few dominant species (representing c.a. 312 50% of the total abundance). The two dominant species, N. phyllepta s.l. and N. 313 spartinetensis, are known to be widespread along the European Atlantic and Northwestern 314 coasts (Vanelslander et al., 2009; Witkowski et al., 2012). Epipelic diatoms generally exhibit 315 316 a low NPQ, in comparison to non-motile ones (Barnett et al., 2014). This does not preclude the existence of interspecific variability between organisms sharing the same life style, 317 considering both behavioural (Underwood et al., 2005; Du et al., 2010, 2012) and 318 319 physiological photoprotective mechanisms (Barnett et al., 2014). Furthermore, the occurrence of N. phyllepta s.l. at both sites can be explained by the pseudocryptic feature of this 320 321 taxonomic group. It comprises different closely related species morphologically 322 indistinguishable but exhibiting important ecophysiological differences and environmental preferences (Vanelslander et al., 2009). To which extent the differences observed between PT 323 and FR sites (i.e. stronger NPQ_{max} in PT assemblage regardless of the stress treatment; 324 different Lhcx patterns) can be explained by the differential relative abundances of N. 325 phyllepta s.l. vs. N. spartinetensis and/or pseudopcryptic species or ecotypes of N. phyllepta 326 s.l is a question that remains to be answered. 327

328

329 Higher temperature increases photoinhibition susceptibility in epipelic MPB

330 To our knowledge, this is the first study which shows that on intact migratory MPB biofilms

high temperature and HL acts synergistically in decreasing photosynthesis, a feature so far

reported only for individual species (Salleh and McMinn, 2011). Other works have shown that 332 temperature alone impacts photosynthesis in benthic diatoms, in monospecific cultures 333 (Morris and Kromkamp, 2003; Yun et al., 2010), MPB suspensions (Blanchard et al., 1997), 334 335 and intact MPB biofilms (Vieira et al., 2013). In conditions where photosynthesis is lightsaturated, benthic diatom photosynthesis can respond to a transient temperature increase if it 336 remains within a 25-35°C range. The high temperature used here $(37.3 \pm 0.7$ °C) falls slightly 337 338 outside this range, a situation likely to occur regularly in situ in summer, as confirmed by our 339 field data.

340 In situ measurements showed that Φ PSII correlated well with irradiance during the emersion period. During LSRE, a strong decrease was also observed after 3 h of HL exposure. 341 Although the applied light dose was lower than *in situ*, our results show that it was strong 342 enough to induce photoinhibition at the community level. The incomplete recovery of $\Phi PSII$ 343 after 15 min LL coincided with a sustained phase of NPQ which was induced during the 344 345 prolonged HL exposure. This can be partly attributed to photoinhibition and illustrates the physiological response to harsh stress (Zhu et al., 2010; Wu et al., 2012; Lavaud and Lepetit, 346 2013). At optimal temperature (27°C; see Blanchard et al., 1997 and Experimental procedures 347 section), photoinhibition remained below 20 %, which is consistent with a previous study 348 using a similar experimental approach (Serôdio et al., 2012). Photoinhibitory effects may be 349 attributed to the progressive accumulation of reactive oxygen species (ROS, Roncarati et al., 350 2008; Waring et al., 2010), as evidenced by the increase of AR at the end of LSRE. Indeed, 351 ROS have been shown to be directly involved in Chl a allomerization (Hynninen et al., 2010 352 and references therein). This is also consistent with the accumulation of Chl a allomers in 353 response to oxidative stress in the diatom P. tricornutum (Cid et al., 1995). When 354 superimposed to HL stress, significantly increased photoinhibition (c.a. 3-times increase in 355 the controls from 27 to 37°C) which is not surprising as PSII is the most thermosensitive 356

component of the photosynthetic apparatus (Mathur *et al.*, 2014). Nevertheless, the amount of
MDA and other aldehydes (TBARS) were not affected, suggesting that antioxidant activities
were sufficient to prevent from substantial ROS-mediated membrane damage during the 3 h
stress exposure, even when migration was inhibited. In such short-term experiments (i.e. < 24
h), it seems that only stronger oxidative stress such as UV- or herbicides-induced leads to
noticeable lipoperoxidation (Rijstenbil, 2005; Wang *et al.*, 2011).

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New evidences on the interplay between behavioural and physiological response to lighttemperature stress in epipelic microphytobenthos

In response to HL alone, diatoms which were free to move (i.e. in the controls) could reach 366 deeper layers of sediment with lower light conditions, potentially closer to their 367 photosynthetic optimum (Kromkamp et al., 1998). This downward migration contributed to a 368 369 reduction of the deleterious effects of HL observed in Lat A treated samples (i.e. immobilized cells at the surface of sediment) such as an enhanced accumulation of Chl a allomers and a 370 371 slower Φ PSII recovery. Our results are in agreement with previous works using Lat A on 372 natural biofilms (Perkins et al., 2010; Serôdio et al., 2012). Similarly, the motile diatoms could reach a depth with lower and more stable temperature, as shown by the in situ 373 374 temperature depth profiles monitored here. Nevertheless, our measurements were not refined enough to conclude about the steepness of the vertical temperature gradient in muddy 375 sediment, as for light (Paterson and Hagerthey, 2001; Cartaxana et al., 2011). Additionally, 376 high temperature alone is known to directly affect the motility of MPB diatoms, i.e. over 35°C 377 cell motility significantly drops-down (Cohn et al., 2003; Du et al., 2012). Such feature might 378 partly explain the stronger photoinhibition observed at 37°C in Lat A free samples, together 379 380 with a direct impact of temperature on physiological mechanisms.

NPQ developed in the same range of mean values (i.e. 3-4) in situ than during LSRE with HL 381 stress alone. Concomitantly, the DD+DT pool size as well as DES increased. DT synthesis 382 was based not only on DD de-epoxidation (as illustrated by DES) but also by additional DT 383 de novo synthesis (Lavaud and Lepetit, 2013; Lepetit et al., 2013) which enlightened the high 384 level of stress the cells were exposed to. Whatever the source, additional DT provides higher 385 physiological photoprotection via higher NPQ and/or via higher prevention of oxidative lipid 386 peroxidation (Lepetit et al., 2010; Lavaud and Lepetit, 2013). DT synthesis (DD+DT and 387 DES) is enhanced by the high temperature stress alone (Salleh and McMinn, 2011), and this 388 effect was even stronger in conditions of a coupled HL and high temperature stress. 389 390 Specifically, DD+DT and DES showed a slower recovery which could represent a feed-back reaction (i.e. related to HL/high temperature impairment of the Calvin cycle activity, Mathur 391 et al., 2014) for the maintenance of a photoprotection capacity to cope with prolonged 392 393 stressful temperature conditions, even under LL (Lavaud and Lepetit, 2013; Lepetit et al., 2013). When the diatom motility was inhibited by Lat A, DES significantly increased 394 395 independently of the temperature, and its recovery was slower. This was apparently not due to 396 a higher DT de novo synthesis as illustrated by the similar DD+DT pool size with and without Lat A. Higher DT synthesis in presence of Lat A had no significant effect on NPQ, except for 397 398 its increased sustainability when in conjunction with high temperature. Higher sustained NPQ 399 (i.e. qI) is known to be related with both sustained DT and photoinhibition in conditions of a strong stress (Lavaud and Lepetit, 2013; Lepetit et al., 2013). As a consequence, the DT 400 quenching efficiency (i.e. the slope of the NPQ-DT relationship) was lower in presence of Lat 401 402 A, likely due to the differential involvement of DT molecules in NPQ (Lavaud and Lepetit, 2013): in presence of Lat A, the additional DT molecules instead most probably acted as ROS 403 scavengers in prevention of lipid peroxidation (Lepetit et al., 2010) arguing for a pronounced 404 stress (i.e. HL + hight temperature + inhibition of motility) level. 405

Although, as expected, the Lhcx pattern in sediment samples was not as clear as the one of 406 monocultures of N. phyllepta, one of the dominant species common to the FR and PT MPB 407 communities, the confirmation of the presence of some of the known isoforms was possible. 408 409 While an isoform with a similar size than Lhcx1 apparently did not react to any of the treatments, we found that the synthesis of two isoforms were enhanced by HL, high 410 temperature and Lat A: one isoform with a size similar to Lhcx3, and one isoform with the 411 412 size of a dimer which especially positively reacted to high temperature and Lat A treatment. Interestingly, these two isoforms were also present in N. phyllepta. A third isoform (of about 413 24 kDa), which was not detected in N. phyllepta, was synthetized in all treatments. Based on 414 415 our knowledge of the synthesis and roles of Lhcx proteins during prolonged stress, it is reasonable to argue that these isoforms participated, at least in part, to the above described 416 NPQ and XC patterns during the coupled HL and high temperature stress. Interestingly, the 417 418 Lhcx3 and 24 kDa isoforms strongly reacted to Lat A out of any stress and to a level comparable to HL stress. It suggests that the inhibition of motility might generate a strong 419 420 preventive physiological answer in diatom cells blocked at the surface of sediment (which can 421 occur in situ during sediment desiccation events for instance), an additional hint that physiological and behavioural photoprotection are closely related in epipelic diatoms. This 422 compensation, together with a stronger NPQ induction and XC activity, was nevertheless not 423 strong enough to prevent immobilized cells from additional photoinhibition, as **PSII** 424 recovery was always lower in Lat A treated samples than in migratory biofilms. 425

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427

428 *Conclusions and perspectives*

By using a number of methodological novelties in MPB research (a new cryo-sampling
method, lipid peroxidation markers for photooxidative stress-TBARS, photosynthetic protein

markers for physiological protection-Lhcx), the present study clearly strengthens the previous 431 statement of a coupled and complementary action of physiological and behavioural processes 432 in protecting the photosynthetic activity of MPB diatoms in stressful environmental 433 434 conditions (Perkins et al., 2010, Cartaxana et al., 2011; Serôdio et al., 2012). Moreover, our results bring a physiological basis to the previously reported depression of MPB 435 photosynthetic productivity in summer (Guarini et al., 1997). They emphasize the joint role of 436 temperature and light in limiting, at least transiently (i.e. during emersion), the photosynthetic 437 activity of MPB biofilm in situ. They confirm that high temperature, together with high light, 438 is an important environmental driver which supports seasonal, spatial and potentially 439 latitudinal MPB biodiversity and biomass distribution at the scale of intertidal mudflats along 440 the Atlantic coast of Southern Europe (Guarini et al., 1997; Brito et al., 2013). 441

Moreover, temperature being expected to increase due to global change (Solomon et al., 442 443 2007), further longer-term experiments are required in order to question its importance in modulating MPB photosynthetic activity. Our results support previous works suggesting that 444 445 MPB gross photosynthesis will most probably not be affected by a small increase (even if 446 significant, i.e. +1°C to +4°C for the most extreme previsions; Solomon et al., 2007) of average temperature (Vieira et al., 2013; Hancke et al., 2014). However, it does not preclude 447 a potential change in the community structure, as observed for associated invertebrate 448 assemblages (Hicks et al., 2011), which might be of importance in driving the whole 449 community response as reported here. Further experiments should rather focus on the 450 cumulative effects of successive extreme events such as heat waves experienced during 451 spring/summer (Schär et al., 2004; Stott et al., 2004), with an emphasis on the 452 'species/individual' vs. 'community/biofilm' response. 453

455 Experimental procedures

456 Study areas and experimental design

Two intertidal mudflats were sampled along the Atlantic Coast of Southern Europe, in France and Portugal (FR and PT sites, respectively): a macrotidal one in the Baie de l'Aiguillon near Esnandes (FR, 46°15.36'N, 1°8,55'W) and a mesotidal estuary in the Ria de Aveiro near Vista Alegre (PT, 40°35' N, 8°41' W). Both sites are composed of fine muddy sediments (dominant particle size < 63 μ m) and are known to be colonized by epipelic diatom-dominated MPB throughout the year (Haubois *et al.*, 2005; Serôdio *et al.*, 2012, respectively).

All experiments were performed consecutively in early summer (weeks 26-27, June-July 2012). For each site, natural migratory MPB was studied *in situ* over one diurnal emersion period. In parallel, light stress-recovery experiments (LSRE) were carried out in controlled conditions in the laboratory at two temperatures using freshly collected samples (see section below). During LSRE, samples were sequentially exposed to (i) low light (LL) level to determine pre-stress reference state; (ii) high light (HL) intensity to induce photoinhibitory effects; (iii) LL to monitor recovery from HL stress (Serôdio *et al.*, 2012).

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471 In situ *measurements*

472 For each site, the day for carrying out the experiment was selected so that the emersion period matched with maximum light and temperature exposure, i.e. noon/early afternoon low tides. 473 Photosynthetic Active Radiation (PAR) and temperature at the surface of the sediment were 474 assessed every 30 s with a universal light-meter and data logger (ULM-500, Walz Effeltrich 475 476 Germany) equipped with a plane light/temperature sensor (accessory of the ULM-500) and a plane cosine quantum sensor (Li-COR). Depth temperature profiles were measured with Hobo 477 sensors (Hobo Pro V2, Massachussets, USA) fixed on a home-made stick that was vertically 478 inserted into the sediment to position the sensors at four (-0.5 cm, -2 cm, -5 cm, -10 cm). The 479

effective photosystem II quantum yield (Φ PSII) and the non-photochemical quenching of 480 chlorophyll a (Chl a) fluorescence (NPQ) of MPB was assessed with a Water-PAM (micro-481 fiber version, Walz GmbH, Effeltrich, Germany) as described in Lefebvre et al. (2011). **PSII** 482 was calculated as $\Delta F / F_m' = (F_m - F) / F_m'$. Rapid light curves (RLCs) with 30 s light steps 483 were used to measure NPQ vs irradiance (NPQ-E) curves (as recommended in Lefebvre et al., 484 2011). NPQ-E curves were fitted with the model by Serôdio and Lavaud (2011) in order to 485 estimate NPQ_{max} where NPQ = $(F_m - F_m) / F_m$. F_m and F_m ' are the maximum Chl a 486 fluorescence levels measured in dark-adapted and illuminated (at the end of each RLCs light 487 steps) MPB, respectively, while F is the steady-state Chl a fluorescence level measured at the 488 end of each RLC light step, just before F_m' was assessed. F was measured with the application 489 of a non-actinic modulated beam of 455 nm, and F_m and F_m' with the application of over-490 saturating (about 4000 μ mol photons m⁻² s⁻¹) light pulses (800 ms). Because in *in situ* as well 491 492 as in laboratory (see below) conditions (i.e. light-responsive migratory biofilm) it is virtually impossible to measure a true F_m level, for the calculation of NPQ, F_m was in reality the 493 494 maximum F_m ' of the respective RLC (i.e. F_m '_{max} ~ F_m , see Lefebvre *et al.*, 2011).

495

496 *Light stress-recovery experiments (LSRE)*

LSRE were carried out following a protocol adapted from Serôdio et al. (2012). The top 1 cm 497 of sediment was collected at the beginning of low tide and immediately transported to the 498 laboratory where it was sieved (0.5 mm mesh) to remove debris and macrofauna. Sediment 499 was thoroughly mixed and spread in trays, forming a 2-4 cm thick slurry, and it was covered 500 501 with seawater from the sampling site and left undisturbed overnight in the dark at room temperature (~20-25°C). The next morning, the slurry was homogenized and transferred to 502 24-well plates using a 100 mL syringe, filling the wells completely (ca. 3 mL). The well 503 plates were then exposed to constant LL of 50 μ mol photons m⁻² s⁻¹, provided by three slide 504

projectors (Reflecta Diamator AF, Reflecta GmbH, Rottenburg, Germany) containing halogen 505 lamps (Quartzline DDL 150W, General Electric, USA) to induce the upward migration of 506 MPB. Once the MPB biofilm was formed (typically after 3 h), the plates were transferred to a 507 water bath of constant temperature of 25°C or 40°C and exposed to constant LL for 30 min. 508 These two experimental temperatures were defined based on a previous study (Blanchard et 509 al., 1997) which defined 25°C as an optimum and 35-40°C as an extreme maximum for the 510 photosynthetic productivity of MPB. A fixed volume (200 µL) of Lat A solution (10 µM final 511 512 concentration) or of filtered sea water (Control samples: Ctrl) was added to each well. Lat A solution was freshly prepared daily from a 1 mM stock solution of purified Lat A (Sigma-513 Aldrich) dissolved in dimethylsulfoxide and stored at -20°C (Cartaxana and Serôdio, 2008). 514 After 15 min waiting for the inhibitor to diffuse, the plates were exposed to HL (1200 µmol 515 photons m⁻² s⁻¹, 3 h). This light dose (12.96 mol photons m⁻²) was previously shown to be 516 517 excessive for MPB photosynthesis (Serôdio et al., 2012). After 3 h, samples were placed under LL for 15 min. In parallel, additional samples (for both Ctrl and +Lat A treatments) 518 519 were maintained during 3 h under continuous LL. Temperature of the sediment in the top 5 mm was regularly checked using the same sensor previously described. The duration of each 520 step described above was always the same, but the timing was adjusted so that HL period 521 522 always started at the low tide maximum expected in situ.

For each chemical and light treatment (Ctrl and +Lat A; HL or continuous LL), several endpoints were assessed on 3-6 independent samples at three times: just before the HL stress (T_0) , at the end of the HL stress and after 15 min of LL recovery (Recov). For the experiments with FR samples, Chl *a* fluorescence was measured using an Imaging-PAM fluorometer (Maxi-PAM M-series, Walz GmbH, Effeltrich, Germany) and samples were collected for further quantification of photosynthetic pigments, lipid peroxidation, and Lhcx proteins. For PT samples, Chl *a* fluorescence was measured with a FluorCAM 800MF, open version (PSI, Brno, Czech Republic, see Serôdio *et al.*, 2013). Samples were collected for the same
analyses than for FR site, except that pigments were not quantified. Samples were collected
for the same analyses than for FR site, except that pigments were not quantified. ΦPSII and
NPQ were calculated as previously described (Serôdio *et al.*, 2012).

For the sediment sampling in wells, a contact-corer specially designed was used to collect rapidly (< 30 s) the uppermost layer of sediment (0.5 mm thick, 1.5 cm² area) in each well. It is based on the cryo-sampling of the biofilm, through the contact of the sediment surface with a metal cylinder of known length previously cooled in liquid nitrogen (Laviale *et al.*, submitted). Each core was immediately frozen in liquid nitrogen and stored at -80°C until further analysis.

540

541 *Pigment analysis*

542 Pigment extraction and HPLC analysis were carried out following Lepetit et al. (2013), with the exception of the first extraction step which was performed after 24 h lyophilisation of the 543 544 sediment cores (3 mL of 4°C mixture extraction buffer: 90% methanol/0.2 M ammonium 545 acetate (90/10 vol/vol) + 10% ethyl acetate). Extracted samples were analyzed with a Hitachi LaChrom Elite HPLC system equipped with a 10°C-cooled autosampler and a Nucleosil 120-546 5 C18 column (Macherey-Nagel). The allomerization of Chl a was calculated as the ratio AR 547 = Chl *a*-allo / tot Chl $a \times 100$ where tot Chl a = (Chl a-allo + Chl a' + Chl a) and Chl *a*-allo 548 and Chl a' are the allomer and epimer forms of Chl a, respectively. The pigment molar 549 concentrations (expressed as mol. 100 mol Chl a^{-1}) for the xanthophyll cycle (XC) pigments 550 diadinoxanthin (DD) and diatoxanthin (DT) were used to calculate the DD de-epoxidation 551 state (DES) as follows: DES = $[DT / (DD + DT)] \times 100$. 552

553

554 *Lipid peroxidation*

Lipid peroxidation generated by oxidative stress was estimated by quantifying thiobarbituric 555 acide reactive substances (TBARS) according to Kwon and Watts (1964). Each sample was 556 resuspended in 0.8 mL of 10% trichloroacetic acid (TCA, Sigma-Aldrich) and 500 mg L⁻¹ 557 butylated hydroxytoluene (BHT, Sigma-Aldrich) and sonicated (U200S control; IKA 558 Labortechnik, Ika-Werke GmbH, Staufen, Germany) on ice 5 times during 30 s (cycle: 1, 559 amplitude: 30%) with intervals of 30 s to prevent overheating. After extraction, 0.8 mL of 560 0.5% thiobarbituric acid (TBA, Sigma-Aldrich) in 10% TCA was added. The mixture was 561 heated for 30 min at 90 °C and immediately put on ice to stop the reaction. Samples were then 562 centrifuged (15 min, 10000 g) and 250 µL of supernatant were transferred in a 96 well-plate 563 (Sarsted). TBARS were measured by subtracting the nonspecific turbidity (at 600 nm) from 564 the TBARS absorption maximum (532 nm) using a Synergy HT absorbance microplate reader 565 (BioTek Instruments Inc., Vermont, USA). Blanks consisted of the same mixture without 566 567 sediment sample. TBARS concentration was expressed in malondialdehyde (MDA) equivalent (nmol mL⁻¹) using an extinction coefficient of 155 mM⁻¹ cm⁻¹. As it was not 568 569 possible to obtain paired measurements of Chl a and MDA, values were normalized to the 570 microphytobenthos biomass present at the sediment surface (measured by spectral reflectance: see below). As migration was susceptible to occur in Ctrl samples, only Lat A-treated samples 571 could be compared. 572

573

574 MPB surface biomass

575 MPB biomass present in the surface layers of the sediment was estimated by spectral 576 reflectance analysis with a fiber optic spectrometer (USB2000-VIS-NIR, grating #3, Ocean 577 Optics, Duiven, The Netherlands) according to Serôdio *et al.* (2009). In particular, 578 measurements were carried out under LL (50 μ mol photons m⁻² s⁻¹) supplied with an halogen 579 lamp (120W, Plusline ES, Philips Lighting) positioned with a 45° angle to the sediment

surface while the sensor was at a fixed distance (~2 cm) perpendicularly to it. Reflectance 580 measurements at 550, 600 and 675 nm (ρ_{550} , ρ_{600} and ρ_{675} , respectively) were used to 581 calculate the normalized index I_{diat} as follows: I_{diat} = 2 × ρ_{600} / (ρ_{550} + ρ_{675}) (modified from 582 Kazemipour et al 2012). It is based on the spectral peak centered at 600 nm between the 583 absorbance bands due to the fucoxanthin at 550 nm and the Chl a at 675 nm. In comparison to 584 other MPB spectral indices available (Méléder et al., 2010), it was shown to be less affected 585 by water or detritical organic matter contents, thus allowing a better comparison between sites 586 587 (data not shown).

588

589 *Lhcx protein analysis*

590 Protein extraction from sediment cores, SDS-PAGE, Western-blot and ECL immunodetection 591 were carried out following the protocol by Coesel et al. (2009) modified by Lepetit et al. (2013): In order to improve the proteins extraction, sediment samples were exposed to eight 592 593 freeze thaw cycles in the extraction buffer (500 μ L lysis buffer + 500 μ L protease inhibitor, Sigma-Aldrich, USA) and then incubated at RT (total extraction was 90 min). Chl a 594 concentration of extracted samples was determined according to Jeffrey and Humphrey 595 (1975) and samples corresponding to an amount of 0.5 μ g Chl *a* were loaded on the gel. 596 597 Correct blotting was verified by correct transfer of pre-stained protein markers on the PVDF 598 membrane (Amersham Hybond-P, GE Healthcare life Sciences, USA) and by staining gels with Coomassie R-250 (Pierce Imperial Protein stain, Thermo scientific, Rockford, USA). 599 Anti-FCP6 (Lhcx1) from Cyclotella cryptica (Westermann and Rhiel, 2005) and anti-Lhcx6 600 601 from T. pseudonana (Zhu and Green, 2010) were applied in a 1:1000 dilution for LR samples overnight (respective incubations with the secondary antibody were 2 h and 1 h). For PT 602 603 samples, the anti-Lhcx6 antibody did not yield a good signal; anti-FCP6 was used at a dilution of 1:1000 (incubation with the secondary antibody, 1 h). Note that because of the species-604

specific immunodetection pattern no quantification per se was possible. However, the 605 comparison with the planktonic models P. tricornutum (Bailleul et al., 2010; Lepetit et al., 606 2013) and T. pseudonana (Zhu et al., 2010) enabled to possibly distinguish different isoforms 607 608 including some for which behaviour under HL and temperature stress and their role is known or suspected (see Introduction; Bailleul et al., 2010; Zhu et al., 2010; Lepetit et al., 2013; 609 Büchel, 2014). For P. tricornutum, T. pseudonana and N. phyllepta samples, the procedure 610 was as in Lepetit et al. (2013): exponentially growing cells were exposed to HL stress (2 h at 611 2000 µmol photons. m⁻². s⁻¹) to induce protein expression of several Lhcx isoforms which 612 were separated on the LDS-gel. 613

614

615 *Taxonomic composition*

For each site, the first cm of sediment was collected and stored at -20 °C until analysis. 616 617 Extraction from sediment of the organic fraction, including microalgae cells, was done following Méléder et al. (2007). Preliminary observations of the organic fraction using optical 618 619 microscopy (Olympus AX70 using 200-fold magnification) were performed to confirm the 620 absence of other classes of microalgae than diatoms. Permanent slides of cleaned diatom frustules (cremation: 2 h at 450 °C) were then prepared in a high refractive index medium 621 (Naphrax; Northern Biological Supplies Ltd, Ipswich, UK). The taxonomic composition was 622 determined to the species level by examining 200-300 frustule valves at 500 fold 623 magnification on the basis of morphological criteria using reference works (Ribeiro et al., 624 2013, and references therein). Scanning electron microscopy was used to confirm the analysis 625 (JEOL JSM 7600F reaching 50 000-fold magnification). 626

627

628 *Statistics*

Data presented are mean ± one standard deviation (SD). Each LSRE experiment was 629 performed twice. Analyses of variance (ANOVA), Tukey's honestly significant difference 630 (HSD) tests, Student's t-tests and linear regressions were performed using the R statistical 631 computing environment (v 2.15.1, Ihaka and Gentleman, 1996). In particular, potential 632 interactions between factors were tested considering data set from each site individually, with 633 time (T₀, after 3 h of HL, after 15 min recovery under LL), temperature (27°C and 37°C) and 634 chemical (Ctrl or +Lat A) as fixed factors. The fact that Chl a fluorescence measurements 635 were made on the same samples over the experiment was also taken into account using the 636 sample name as random effect in a linear mixed effect model (lme function of the nlme 637 package: Pinheiro and Bates, 2000). Comparisons between sites (FR, PT) were performed on 638 data gathered at the end of the LSRE (after recovery), expressed as % of T₀. Site, temperature 639 and chemical were considered as fixed factors. Data normality (Shapiro-Wilk test) and 640 641 homoscedasticity (Bartlett test) were checked using the residuals.

643 Acknowledgements

The present study was supported by the Egide/Campus France-PHC Pessoa exchange 644 program (n°27377TB, JL and JS), the Centre National de la Recherche Scientifique-CNRS 645 (JL), the University of La Rochelle-ULR and the Contrat Plan Etat Région-CPER 'Littoral' 646 (JL, BL post-doc grant), the Region Poitou-Charentes (AB Ph.D. grant), the Deutscher 647 648 Akademischer Austausch Dienst-DAAD (BL post-doc grant), European Funds through 649 COMPETE, the Portuguese Science Foundation (FCT) through projects PEST (PEst-C/MAR/LA0017/2013) and MigROS (PTDC/MAR/112473/2009: JS, ML post-doc grant and 650 SF research assistant grant), and through grants SFRH/BSAB/962/2009 (JS), SFRH/BD/ 651 44860/2008 (JE Ph.D. grant). The authors thank Dr. E. Rhiel (Univ. Oldenburg, Germany) 652 653 and Prof. B. Green (Univ. British Columbia, Canada) for the kind gifts of the anti-FCP6 and of the anti-Lhcx6, respectively, Prof. B. Green for her help analyzing the Lhcx6 Western-blot, 654 C. Pais for her technical assistance with the TBARS quantification, Dr L. Ribeiro (Univ. 655 656 Lisboa, Portugal) for his help in confirming the identification of N. phyllepta s.l., and Sr Armindo do Evaristo for his cheerful support. We thank the two anonymous reviewers for 657 their constructive comments on the manuscript. 658

660 **References**

- Admiraal, W. (1984) The ecology of estuarine sediment inhabiting diatoms. *Prog Phycol Res*3: 269-322.
- Bailleul, B., Rogato, A., de Martino, A., Coesel, S., Cardol, P., Bowler, C. et al. (2010) An
- atypical member of the light-harvesting complex stress-related protein family modulates
 diatom responses to light. *Proc Natl Acad Sci USA* **107**: 18214-18219.
- Barnett, A., Méléder, M., Blommaert, L., Lepetit, B., Gaudin, P., Vyverman, W. et al. (2014)
 Growth form defines physiological photoprotective capacity in intertidal benthic diatoms.
- 668 *ISME J* doi:10.1038/ismej.2014.105.
- Blanchard, G.F., Guarini, J.M., Gros, P., and Richard, P. (1997) Seasonal effect on the
 relationship between the photosynthetic capacity of intertidal microphytobenthos and
 temperature. *J Phycol* 33: 723-728.
- Blanchard, G.F., Guarini, J.M., Richard, P., Gros, P., and Mornet, F. (1996) Quantifying the
 short-term temperature effect on light-saturated photosynthesis of intertidal
 microphytobenthos. *Mar Ecol Prog Ser* 134: 309-313.
- Brito, A.C., Benyoucef, I., Jesus, B., Brotas, V., Gernez, P., Mendes, C.R. et al. (2013)
 Seasonality of microphytobenthos revealed by remote-sensing in a South European
- 677 estuary. *Cont Shelf Res* **66**: 83-91.
- Brunet, C., and Lavaud, J. (2010) Can the xanthophyll cycle help extract the essence of the
 microalgal functional response to a variable light environment? J *Plankton Res* 32: 16091617.
- Büchel, C. (2014) Fucoxanthin-chlorophyll-proteins and non-photochemical fluorescence
 quenching in diatoms. In *Non-Photochemical Fluorescence Quenching and Energy*
- 683 Dissipation In Plants, Algae and Cyanobacteria. Demmig-Adams, B., Garab, G., Adams,
- 684 W.W.III, and Govindjee (eds). Advances in Photosynthesis and Respiration, Volume 40.,

- 685 pp 259-275. Dordretch, The Netherlands: Springer.
- Cartaxana, P., and Serôdio, J. (2008) Inhibiting diatom motility: a new tool for the study of
 the photophysiology of intertidal microphytobenthic biofilms. *Limnol Oceanogr-Meth.* 6:
 466-476.
- Cartaxana, P., Ruivo, M., Hubas, C., Davidson, I., Serôdio, J., and Jesus, B. (2011)
 Physiological versus behavioral photoprotection in intertidal epipelic and epipsammic
 benthic diatom communities. *J Exp Mar Biol Ecol* 405: 120-127.
- 692 Cartaxana, P., Domingues, N., Cruz, S., Jesus, B., Laviale, M., Serôdio, J., and da Silva, J.M.
- 693 (2013) Photoinhibition in benthic diatom assemblages under light stress. *Aquat Microb*694 *Ecol* **70**: 87-92.
- Chevalier, E.M., Gévaert, F., and Créach, A. (2010) In situ photosynthetic activity and
 xanthophylls cycle development of undisturbed microphytobenthos in an intertidal
 mudflat. *J Exp Mar Biol Ecol* 385: 44-49.
- Cid, A., Herrero, C., Torres, E., and Abalde, J. (1995) Copper toxicity on the marine
 microalga *Phaeodactylum tricornutum* Effects on photosynthesis and related parameters.
 Aquat Toxicol 31: 165-174.
- Coelho, H., Vieira, S., and Serôdio, J. (2011) Endogenous versus environmental control of
 vertical migration by intertidal benthic microalgae. *Eur J Phycol* 46: 271-281.
- Coesel, S., Mangogna, M., Ishikawa, T., Heijde, M., Rogato, A., Finazzi, G., Todo, T.,
 Bowler, C., Falciatore, A. (2009) Diatom PtCPF1 is a new cryptochrome/photolyase
 family member with DNA repair and transcription regulation activity. *EMBO Rep* 10:
 655–661.
- Cohn, S.A., Farrell, J.F., Munro, J.D., Ragland, R.L., Weitzell, R.E., and Wibisono, B.L.
 (2003) The effect of temperature and mixed species composition on diatom motility and
 adhesion. *Diatom Res* 18: 225-243.

- Consalvey, M., Paterson, D.M., and Underwood, G.J.C. (2004) The ups and downs of life in a
 benthic biofilm: Migration of benthic diatoms. *Diatom Res* 19: 181-202.
- 712 Davison, I.R. (1991) Environmental effects on algal photosynthesis Temperature. *J Phycol*713 27: 2-8.
- Depauw, F.A., Rogato, A., d'Alcala, M.R., and Falciatore, A. (2012) Exploring the molecular
 basis of responses to light in marine diatoms. *J Exp Bot* 63: 1575-1591.
- Du, G.Y., J.H., O., Li, H., and Chung, I.K. (2010) Effect of light and sediment grain size on
 the vertical migration of benthic diatoms. *Algae* 25: 133-140.
- 718 Du, G.Y., Li, W.T., Li, H.B., and Chung, I.K. (2012) Migratory responses of benthic diatoms
- to light and temperature monitored by chlorophyll fluorescence. *J Plant Biol* **55**: 159-164.
- Goss, R., and Jakob, T. (2010) Regulation and function of xanthophyll cycle-dependent
 photoprotection in algae. *Photosynth Res* 106: 103-122.
- Goss, R., Pinto, E.A., Wilhelm, C., and Richter, M. (2006) The importance of a highly active
 and delta pH-regulated diatoxanthin epoxidase for the regulation of the PSII antenna
 function in diadinoxanthin cycle containing algae. *J Plant Physiol* 163: 1008-1021.
- Grant, J. (1986) Sensitivity of benthic community respiration and primary production to
 changes in temperature and light. *Mar Biol* **90**: 299-306.
- Guarini, J.M., Blanchard, G.F., and Richard, P. (2006) Modelling the dynamic of the
 microphytobenthic biomass and primary production in European intertidal mudflats. In
 Functioning of microphytobenthos in estuaries. Kromkamp, J.C., de Brouwer, J.F.C.,
- Blanchard, G.F., Forster, R.M., and Créach, V. (eds). Amsterdam, the Netherlands:
 KNAW, pp. 187-226.
- Guarini, J.M., Blanchard, G.F., Gros, P., and Harrison, S.J. (1997) Modelling the mud surface
 temperature on intertidal flats to investigate the spatio-temporal dynamics of the benthic
 microalgal photosynthetic capacity. *Mar Ecol Prog Ser* 153: 25-36.

- Hancke, K., and Glud, R.N. (2004) Temperature effects on respiration and photosynthesis in
 three diatom-dominated benthic communities. *Aquat Microb Ecol* 37: 265-281.
- Hancke, K., Sorell, B.K., Chresten, L., Lund-Hansen, L.C., Larsen, M., Hancke, T., and Glud,
- R.N. (2014) Effects of temperature and irradiance on a benthic microalgal community: A
- combined two-dimensional oxygen and fluorescence imaging approach. *Limnol.*
- 740 *Oceanogr.* **59**: 1599–1611.
- Haubois, A.G., Sylvestre, F., Guarini, J.M., Richard, P., and Blanchard, G.F. (2005) Spatiotemporal structure of the epipelic diatom assemblage from an intertidal mudflat in
 Marennes-Oleron Bay, France. *Estuar Coast Shelf S* 64: 385-394.
- Hicks, N. Bulling, M.T., Solan, M., Raffaelli, D., White, P.C.L., and Paterson, D.M. (2011)
- 745 Impact of biodiversity-climate futures on primary production and metabolism in a model
 746 benthic estuarine system. *BMC Ecology*. 11: 7.
- Hynninen, P.H., Kaartinen, V., and Kolehmainen, E. (2010) Horseradish peroxidase-catalyzed
 oxidation of chlorophyll a with hydrogen peroxide Characterization of the products and
 mechanism of the reaction. *BBA-Bioenergetics* 1797: 531-542.
- Hwang, Y.-S., Jung, G., and Jin, E. (2008) Transcriptome analysis of acclimatory responses to
 thermal stress in Antarctic algae. *Biochem Bioph Res Co* 367: 635-641
- 752 Ihaka, R., and Gentleman, R. (1996) R: a language for data analysis and graphics. *J Comput*753 *Graph Stat* 5: 299-314.
- Jeffrey, S.W., and Humphrey, G.F. (1975) New spectrophotometric equations for determining
 chlorophylls a, b c1 and c2 in higher-plants, algae and natural phytoplankton. *Biochem Physiol Pfl* 167: 191-194.
- Jesus, B., Brotas, V., Ribeiro, L., Mendes, C.R., Cartaxana, P., and Paterson, D.M. (2009)
 Adaptations of microphytobenthos assemblages to sediment type and tidal position. *Cont Shelf Res* 29: 1624-1634.

- Kazemipour, F., Launeau, P., and Meleder, V. (2012) Microphytobenthos biomass mapping
 using the optical model of diatom biofilms: Application to hyperspectral images of
 Bourgneuf Bay. *Remote Sens Environ* 127: 1-13.
- Kromkamp, J., Barranguet, C., and Peene, J. (1998) Determination of microphytobenthos PSII
 quantum efficiency and photosynthetic activity by means of variable chlorophyll
 fluorescence. *Mar Ecol Prog Ser* 162: 45-55.
- Kwon, T., and Watts, B.M. (1964) Malonaldehyde in aqueous solution and its role as measure
 of lipid oxidation in foods. *J Food Sci* 29: 294-302.
- Lavaud, J., and Lepetit, B. (2013) An explanation for the inter-species variability of the
 photoprotective non-photochemical chlorophyll fluorescence quenching in diatoms. *BBA- Bioenergetics* 1827: 294-302.
- The peculiar features of non-photochemical fluorescence
 quenching in diatoms and macrophytic brown algae. In *Non-Photochemical Fluorescence*

773 *Quenching and Energy Dissipation In Plants, Algae and Cyanobacteria.* Demmig-Adams,

- B., Garab, G., Adams, W.W.III, and Govindjee (eds). Advances in Photosynthesis and
 Respiration, Volume 40, pp 421-443. Dordretch, The Netherlands: Springer.
- Lavaud, J., Materna, A.C., Sturm, S., Vugrinec, S., and Kroth, P.G. (2012) Silencing of the
 violaxanthin de-epoxidase gene in the diatom *Phaeodactylum tricornutum* reduces
 diatoxanthin synthesis and non-photochemical quenching. *Plos One* 7: e36806.
- Laviale, M., Ezequiel, J., Pais, C., Cartaxana P. and Serôdio, J. (submitted) A new highresolution method for the fast vertical sampling of intertidal fine sediments.
- Lefebvre, S., Mouget, J.L., and Lavaud, J. (2011) Duration of rapid light curves for
 determining the photosynthetic activity of microphytobenthos biofilm in situ. *Aquat Bot* **95**: 1-8.
- 784 Lepetit, B., Volke, D., Gilbert, M., Wilhelm, C., and Goss, R. (2010) Evidence for the

- existence of one antenna-associated, lipid-dissolved and two protein-bound pools of
 diadinoxanthin cycle pigments in diatoms. *Plant Physiol* 154: 1905-1920.
- Lepetit, B., Goss, R., Jakob, T., and Wilhelm, C. (2012) Molecular dynamics of the diatom
 thylakoid membrane under different light conditions. *Photosynth Res* 111: 245-257.
- 789 Lepetit, B., Sturm, S., Rogato, A., Gruber, A., Sachse, M., Falciatore, A., Kroth, P.G., and
- 790 Lavaud, J. (2013) High light Acclimation in the secondary plastids containing diatom
- *Phaeodactylum tricornutum* is triggered by the redox state of the plastoquinone pool. *Plant Physiol* 161: 853-865.
- Lillebo, A.I., Valega, M., Otero, M., Pardal, M.A., Pereira, E., and Duarte, A.C. (2010) Daily
 and inter-tidal variations of Fe, Mn and Hg in the water column of a contaminated salt
 marsh: Halophytes effect. *Estuar Coast Shelf S* 88: 91-98.
- Mathur, S., Agrawal, D., and Jajoo, A. (2014) Photosynthesis: Response to high temperature
 stress. *J Photoch Photobio B* 137: 116-126
- Méléder, V., Rincé, Y., Barillé, L., Gaudin, P., and Rosa, P. (2007) Spatiotemporal changes in
 microphytobenthos assemblages in a macrotidal flat (Bourgneuf bay, France). *J Phycol*43: 1177-1190.
- 801 Méléder, V., Launeau, P., Barillé, L., Combe, J.-P., Carrère V., Jesus, B., and Verpoorter, C.
- 802 (2010) Hyperspectral imaging for mapping microphytobenthos in coastal areas. In:
- *Geomatic solutions for coastal environments*. Maanan, M., and Robin, M. (eds). Nova
 Science Publishers.
- Morris, E.P., and Kromkamp, J.C. (2003) Influence of temperature on the relationship
 between oxygen- and fluorescence-based estimates of photosynthetic parameters in a
 marine benthic diatom (*Cylindrotheca closterium*). *Eur J Phycol* 38: 133-142.
- Paterson, D.M., and Hagerthey, S.E. (2001) Microphytobenthos in contrasting coastal
 ecosystems: Biology and dynamics. In *Ecological Comparison of Sedimentary Shores*.

- 810 Reise, K. (ed). Berlin Heildelberg: Springer-Verlag, pp. 105-125.
- Perkins, R.G., Lavaud, J., Serôdio, J., Mouget, J.L., Cartaxana, P., Rosa, P. et al. (2010)
 Vertical cell movement is a primary response of intertidal benthic biofilms to increasing
 light dose. *Mar Ecol Prog Ser* 416: 93-103.
- Pinheiro, J.C., and Bates, D.M. (2000) *Mixed-models in S and S-PLUS*. New York, USA:
 Springer-Verlag.
- Ribeiro, L., Brotas, V., Rince, Y., and Jesus, B. (2013) Structure and diversity of intertidal
 benthic diatom assemblages in contrasting shores: A case study from the Tagus estuary. *J Phycol* 49: 258-270.
- Rijstenbil, J.W. (2005) UV- and salinity-induced oxidative effects in the marine diatom *Cylindrotheca closterium* during simulated emersion. *Mar Biol* 147: 1063-1073.
- Roncarati, F., Rijstenbil, J.W., and Pistocchi, R. (2008) Photosynthetic performance,
 oxidative damage and antioxidants in *Cylindrotheca closterium* in response to high
 irradiance, UVB radiation and salinity. *Mar Biol* 153: 965-973.
- Saburova, M.A., and Polikarpov, I.G. (2003) Diatom activity within soft sediments:
 behavioural and physiological processes. *Mar Ecol Prog Ser* 251: 115-126.
- Salleh, S., and McMinn, A. (2011) The effects of temperature on the photosynthetic
 parameters and recovery of two temperate benthic microalgae, *Amphora cf. coffeaeformis*and *Cocconeis* cf. *sublittoralis* (Bacillariophyceae). *J Phycol* 47: 1413-1424.
- Schär, C., Vidale P.L., Lüthi, D., Frei, C., Häberli, C., Liniger, M.A., and Appenzeller, C.
 (2004) The role of increasing temperature variability in European summer heatwaves. *Nature* 427: 332-336.
- Serôdio, J., and Catarino, F. (1999) Fortnightly light and temperature variability in estuarine
 intertidal sediments and implications for microphytobenthos primary productivity. *Aquat. Ecol.* 33: 235-241.

- Serôdio, J., Cartaxana, P., Coelho, H., and Vieira, S. (2009) Effects of chlorophyll
 fluorescence on the estimation of microphytobenthos biomass using spectral reflectance
 indices. *Remote Sens Environ* 113: 1760-1768.
- 838 Serôdio, J., and Lavaud, J. (2011) A model for describing the light response of the
 839 nonphotochemical quenching of chlorophyll fluorescence. *Photosynth Res* 108: 61-76.
- Serôdio, J., Cruz, S., Vieira, S., and Brotas, V. (2005) Non-photochemical quenching of
 chlorophyll fluorescence and operation of the xanthophyll cycle in estuarine
 microphytobenthos. *J Exp Mar Biol Ecol* 326: 157-169.
- Serôdio, J., Ezequiel, J., Frömmlet, J., Laviale, M., and Lavaud, J. (2013) A method for the
 rapid generation of non-sequential light-response curves of chlorophyll fluorescence. *Plant Physiol* 163: 1089-1102.
- Serôdio, J., Ezequiel, J., Barnett, A., Mouget, J.L., Méléder, V., Laviale, M., and Lavaud, J.
 (2012) Efficiency of photoprotection in microphytobenthos: role of vertical migration and
 the xanthophyll cycle against photoinhibition. *Aquat Microb Ecol* 67: 161-175.
- 849 Solomon, S., Qin, D., Manning, M., Chen, Z., Marquis, M., Averyt, K.B., Tignor, M., and

850 Miller, H.L. (eds.) (2007) Contribution of Working Group I to the Fourth Assessment

- 851 Report of the Intergovernmental Panel on Climate Change. In *Climate Change 2007:*
- 852 *Impacts, Adaptation and Vulnerability.* 996 pp., <u>Cambridge University Press</u>, Cambridge,
- United Kingdom and New York, NY, USA.
- Stott, P.A., Stone, D.A., and Allen, M.R. (2004) Human contribution to the European
 heatwave of 2003. *Nature* 432: 610-614.
- Underwood, G.J.C., and Kromkamp, J. (1999) Primary production by phytoplankton and
 microphytobenthos in estuaries. In *Advances in Ecological Research, Vol 29: Estuaries*,
 pp. 93-153.
- van Leeuwe, M.A., Brotas, V., Consalvey, M., Forster, R.M., Gillespie, D., Jesus, B. et al.

860 (2008) Photoacclimation in microphytobenthos and the role of xanthophyll pigments. *Eur*861 *J Phycol* 43: 123-132.

Vanelslander, B., Creach, V., Vanormelingen, P., Ernst, A., Chepurnov, V.A., Sahan, E. et al.

- 863 (2009) Ecological differentation between sympatric pseudocryptic species in the estuarine
 864 benthic diatom *Navicula phyllepta* (Bacillariophyceae). *J Phycol* 45: 1278-1289.
- 865 Vieira, S., Ribeiro, L., da Silva, J.M., and Cartaxana, P. (2013) Effects of short-term changes
- in sediment temperature on the photosynthesis of two intertidal microphytobenthos
 communities. *Estuar Coast Shelf S* 119: 112-118.
- Wang, Z.-H., Nie, X.-P., Yue, W.-J., and Li, X. (2012) Physiological responses of three
 marine microalgae exposed to cypermethrin. *Environ Toxicol* 27: 563-572.
- Waring, J., Klenell, M., Bechtold, U., Underwood, G.J.C., and Baker, N.R. (2010) Lightinduced responses of oxygen photoreduction, reactive oxygen species production and
 scavenging in two diatom species. *J Phycol* 46: 1206-1217.
- Westermann, M., and Rhiel, E. (2005) Localisation of fucoxanthin chlorophyll a/c-binding
 polypeptides of the centric diatom *Cyclotella cryptica* by immuno-electron microscopy. *Protoplasma* 225: 217-223.
- Witkowski, A., Sullivan, M.J., Bogaczewicz-Adamczak, B., Bak, M., Rhiel, E., Ribeiro, L.,
 and Richard, P. (2012) Morphology and distribution of a little known but widespread
 diatom (Bacillariophyceae), *Navicula spartinetensis* Sullivan et Reimer. *Diatom Res* 27:
 43-51.
- Wu, H.Y., Roy, S., Alami, M., Green, B.R., and Campbell, D.A. (2012) Photosystem II
 photoinactivation, repair, and protection in marine centric diatoms. *Plant Physiol* 160:
 464-476.
- Yun, M.S., Lee, S.H., and Chung, I.K. (2010) Photosynthetic activity of benthic diatoms in
 response to different temperatures. *J Appl Phycol* 22: 559-562.

- 885 Zhu, S.H., and Green, B.R. (2010) Photoprotection in the diatom *Thalassiosira pseudonana*:
- Role of LI818-like proteins in response to high light stress. *BBA-Bioenergetics* 1797:
 1449-1457.

889 Figure legends

890 Figure 1

In situ environmental conditions and photosynthetic activity of microphytobenthos 891 892 during a summer diurnal emersion in Portuguese (Ria de Aveiro, Vista Alegre, A- and C-) and French (Baie de l'Aiguillon, Esnandes, B- and D-) intertidal flats. A- and B-893 Photosynthetic Active Radiation (PAR, μ mol photons m⁻² s⁻¹, black line) at the surface of the 894 sediment and temperature (°C) at five depths (surface, -0.5 cm, -2 cm, -5 cm, -10 cm, grey 895 lines): C- and D- effective PSII quantum yield (Φ PSII: open squares) and maximum non-896 photochemical quenching of Chl a fluorescence (NPQ_{max}: closed circles). Local time (FR: 897 UTC+1, PT: UTC). 898

899

900 **Figure 2**

Index of photoinhibition (decrease in Φ PSII) of microphytobenthos harvested at the French site (FR) during light stress recovery experiments performed on sediment exposed at two temperatures (27°C, A- or 37°C, B-) and with an inhibitor of cell motility (+ Lat A) or without (Ctrl). T₀, beginning of the experiment under low light (50 µmol photons m⁻² s⁻¹); HL, after 3 h high light (1200 µmol photons m⁻² s⁻¹); Recov, after 15 min recovery under low light (50 µmol photons m⁻² s⁻¹) following the 3 h high light; LL, after 3 h of low light (50 µmol photons m⁻² s⁻¹). Values are mean ± SD (n=6).

908

909 Figure 3

910 Index of photoprotection (NPQ) of microphytobenthos harvested at the French site (FR) 911 during light stress recovery experiments performed on sediment exposed at two 912 temperatures (27°C, A- or 37°C, B-) and with an inhibitor of cell motility (+Lat A) or 913 without (Ctrl). Legend as in Fig. 2. Values are mean \pm SD (n=6).

914

915 Figure 4

Pigment content makers of photooxidative stress (Chl a allomerization ratio: AR) and of 916 photoprotection (DD+DT pool; DD de-epoxidation state: DES) of microphytobenthos 917 harvested at the French site (FR) during light stress recovery experiments performed on 918 sediment exposed at two temperatures (27°C or 37°C) and with an inhibitor of cell 919 motility (+Lat A) or without (Ctrl). A- AR, data from 27 and 37°C experiments were 920 pooled, * significant difference compared to Ctrl at the same time (P < 0.05, Tukey HSD); B-921 DD+DT pool (in mol.100 mol Chl a^{-1}), data from Ctrl and Lat A-treated samples were pooled, 922 ** significant difference compared to 27 °C at the same time (P < 0.01, Tukey HSD); C- and 923 D- DES, *** significant difference compared to Ctrl at the same time (P < 0.001, Tukey 924 HSD). Labels as in Fig. 2. Values are mean \pm SD (n=12: A,B or n=6: C,D) expressed as $\%T_0$ 925

926

927 **Figure 5**

Western blot of the Lhcx proteins of microphytobenthos harvested at the French site during light stress recovery experiments performed on sediment exposed at two temperatures (27°C, A- or 37°C, B-) and with an inhibitor of cell motility (+) or without (-). HL, after 3 h HL (1200 µmol photons $m^{-2} s^{-1}$); LL, after 3 h of LL (50 µmol photons $m^{-2} s^{-1}$). All samples were normalized to the same Chl *a* concentration (15 µg Chl *a* mL⁻¹). A-

Lhcx proteins were detected using a monoclonal anti-FCP6 (Lhcx1). This antibody detects, 933 similarly as the anti-LHCSR of C. reinhardtii (Lepetit et al., 2013), also Lhcx2 and Lhcx3 in 934 P. tricornutum (strain 'P.t.1') used here as a control. N. phyllepta s.l. (N.p.), the dominant 935 species of the MPB community, was also tested. B- and C- Lhcx proteins were detected using 936 a monoclonal anti-Lhcx6 from T. pseudonana (T.p.) (Zhu and Green, 2010) which also 937 potentially detect Lhcx3 and an Lhcx dimer in P. tricornutum; C- is a special focus on the 938 band between 33-50 kDa in 37°C exposed samples. Revelation times were: A- 30 s; B- 15 s 939 for the bands between 17-23 kDa, 45 s for the band between 33-50 kDa; C- 30 s;. Note that 940 for A- all samples exposed at 37°C did not yield satisfying signal; and for B-, the sample HL+ 941 27°C was lost. 942

943

944 Figure 6

Amount of lipid peroxidation marker (TBARS content in 10^{-6} nmol MDA equivalent mL⁻¹) in microphytobenthos harvested at the French site (FR) during light stress recovery experiments performed on sediment exposed at two temperatures (27°C, A- or 37°C, B-) with an inhibitor of cell motility (+Lat A). Labels as in Fig. 2. Values are mean ± SD (n=6) normalized to surface biomass (I_{diat} biomass index).

950

951 **Figure 7**

952 Comparison of photoinhibition level (decrease in Φ PSII) and photoprotection (NPQ) in 953 microphytobenthos harvested at the Portuguese (PT) and the French (FR) sites during 954 light stress recovery experiments performed on sediment exposed at two temperatures 955 (27°C or 37°C) and with an inhibitor of cell motility (+Lat A) or without (Ctrl). Φ PSII 956 recovery (A) and sustained NPQ (B) after 15 min incubation under LL (50 μ mol photons m⁻² 957 s⁻¹) following 3 h of HL exposure (1200 μ mol photons m⁻² s⁻¹), *** significant difference 958 between sites (P < 0.001, Tukey HSD). Values are mean ± SD (n=6).

959

960 Figure 8

Western-blot of the Lhcx proteins of microphytobenthos harvested at the Portuguese 961 site during light stress recovery experiments performed on sediment exposed at two 962 963 temperatures (27°C or 37°C) and with an inhibitor of cell motility (+) or without (-). HL, after 3 h HL (1200 µmol photons. m⁻². s⁻¹) treatment; LL, after 3 h of LL exposure. All 964 samples were normalized to the same Chl *a* concentration (15 μ g Chl *a* mL⁻¹). Lhcx proteins 965 were detected using a monoclonal anti-FCP6 (Lhcx1). This antibody detects, similarly as the 966 anti-LHCSR of C. reinhardtii (Lepetit et al., 2013), also Lhcx2 and Lhcx3 in P. tricornutum 967 968 (strain 'P.t.1') used here as a control. Note that i) the 27°C and 37°C samples were not loaded on the same gel but the revelation time was the same (30 s), ii) the use of the monoclonal anti-969 Lhcx6 from T. pseudonana did not yield any signal with these samples. 970

Figure 1_Laviale *et al.*

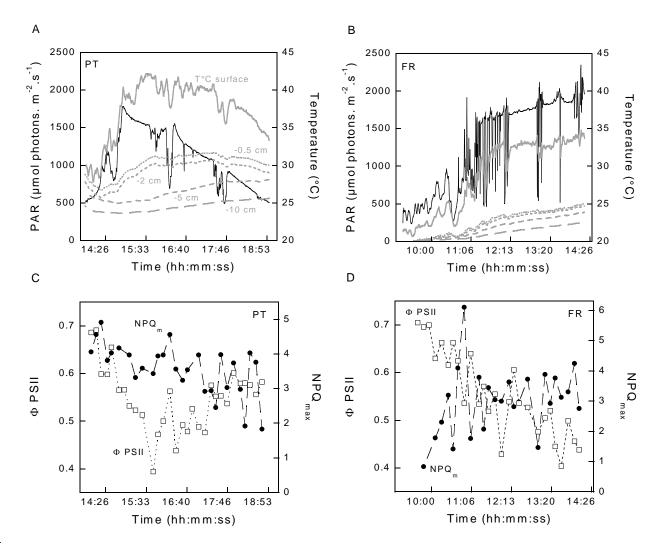
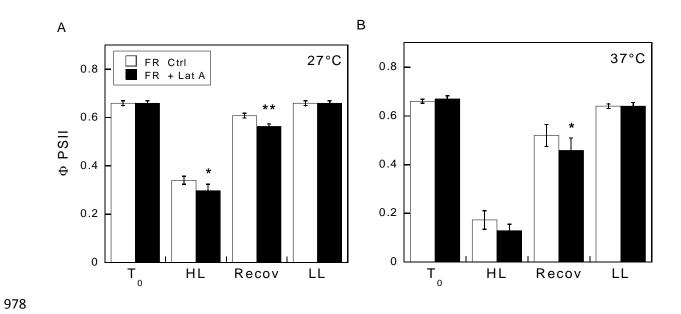


Figure 2_Laviale *et al.*



981 Figure 3_Laviale *et al.*

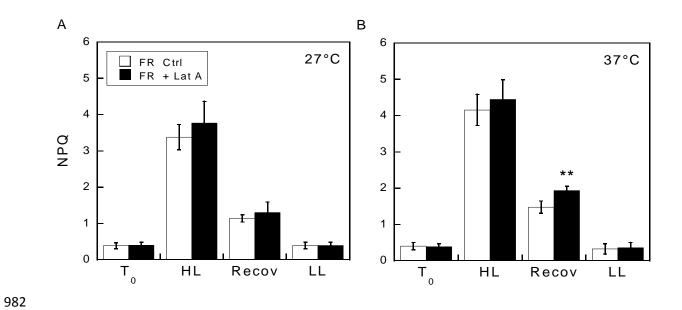




Figure 4_Laviale *et al.*

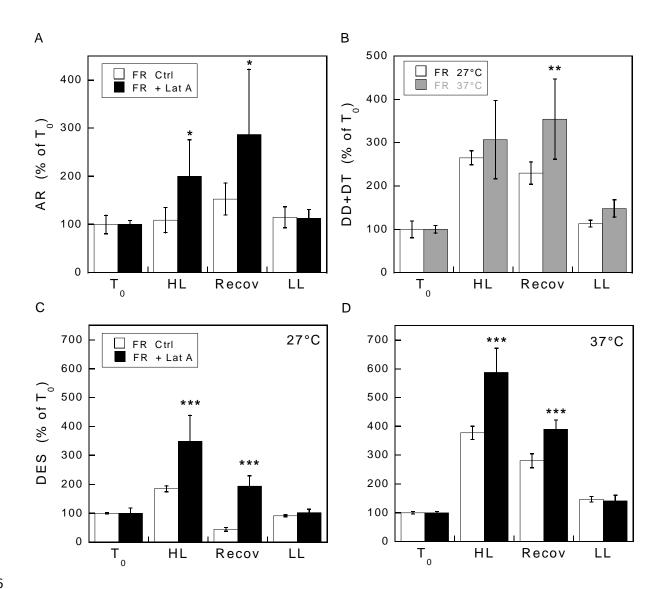


Figure 5_Laviale *et al.*

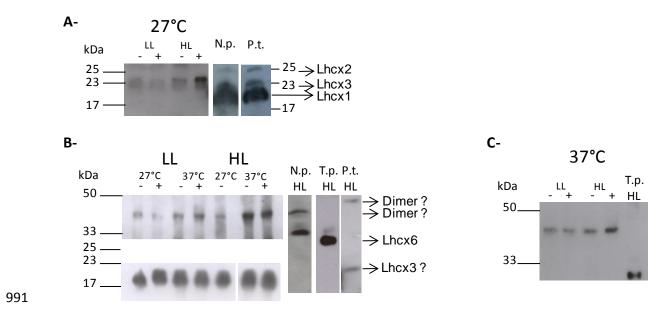
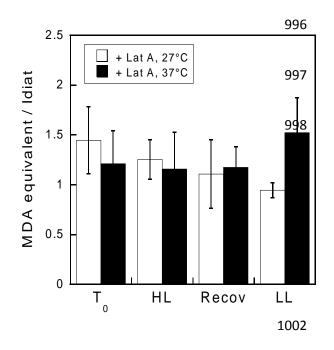


Figure 6_Laviale *et al.*



1006 Figure 7_Laviale *et al.*

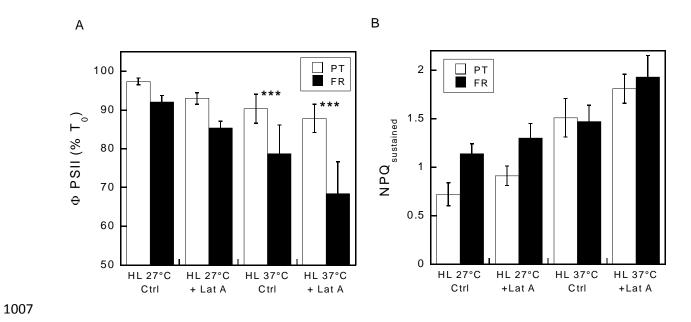


Figure 8_Laviale *et al.*

