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Reactivity of 3-hydroxy-3-methyl-2-butanone: Photolysis
and OH reaction kinetics
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25 Abstract

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27 Hydroxycarbonyl compounds are important secondary reaction products in the oxidation of Volatile Organic Compounds (VOCs) in the atmosphere. The 28 atmospheric fate of these oxygenated VOCs is however poorly understood, 29 30 especially the relevance of the photolytic pathway. In this work, a combined investigation of the photolysis and temperature-dependent OH radical reaction of 3-31 32 hydroxy-3-methyl-2-butanone (3H3M2B) is presented. A photolysis lifetime of about 4-5 days was estimated with a global quantum yield of 0.10. The OH reaction rate 33 coefficient follows the Arrhenius trend (298 - 356 K) and could be modelled through 34 the following expression: $k_{3H3M2B}(T) = (5.12\pm0.07) \times 10^{-12} \exp(-563\pm119/T)$ in cm³ 35 molecule⁻¹ s⁻¹. A 3H3M2B atmospheric lifetime of 15 days towards the OH radical 36 37 was evaluated. Our results showed that the photolysis pathway is the major degradation channel for 3H3M2B. Photolysis products were identified and quantified 38 39 in the present work with a carbon balance of around 80% enabling a reaction 40 mechanism to be proposed. The present work underlines the need for further studies 41 on the atmospheric chemistry of oxygenated VOCs.

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53 Keywords: 3-hydroxy-3-methyl-2-butanone, photolysis, kinetics, OH radicals, 54 tropospheric lifetimes

56 **1. Introduction**

57 Oxygenated Volatile Organic Compounds (OVOCs) are critical components in 58 the chemistry of the troposphere. These species constitute a large family of Volatile 59 Organic Compounds (VOCs) (Atkinson et al., 2003) emitted from various anthropogenic and biogenic sources. More importantly, they are formed in situ as 60 61 intermediates of photooxidation of several VOCs. Though scarce measurement data 62 exist on the concentrations of multi oxygenated VOCs (e.g. (Matsunaga et al., 2000; 63 Destaillats et al., 2002; Spaulding et al., 2002)), evidence has been gained in the last decade for a significant role played by such compounds in atmospheric chemistry 64 65 (Singh et al., 2001).

Chemical mechanism modelling studies indicate that more and more oxidized 66 67 species are formed in the course of VOC oxidation, especially compounds bearing 68 hydroxyl and carbonyl functions (Aumont et al., 2005; Madronich, 2006; Goldstein et 69 al., 2007). Thus, β -hydroxycarbonyls can be formed from the OH[•] radical initiated 70 reactions of diols with molar formation yields between 50% and 90% (Bethel et al., 71 2003), as well as from alcohols and alkenes (Tuazon et al., 1998; Reisen et al., 72 2003). In the atmosphere, OVOCs are supposed to react mainly with OH[•] radicals, 73 although reactions with NO₃[•], Cl and ozone cannot be excluded for some of the species in specific environments (Mellouki et al., 2003). The photolysis channel could 74 75 be significant for carbonyl species (Moortgat, 2001; Bouzidi et al., 2014) and yet, its 76 importance in the case of hydroxycarbonyl VOCs is almost unknown, while recent 77 studies suggest that photolysis could be significant (Henry et al., 2012; Messaadia et 78 al., 2012). Atmospheric chemistry of hydroxyacetone, the smallest hydroxyketone 79 compound, is reasonably well understood (Orlando et al., 1999; Dillon et al., 2006). 80 For larger hydroxycarbonyls ($\geq C_4$), degradation pathways have not been well 81 characterized. The few available studies indicate that saturated hydroxyketones are mainly removed by OH[•] radicals (Aschmann et al., 2000; Magneron et al., 2003; 82 83 Messaadia et al., 2012; El Dib et al., 2013; Messaadia et al., 2013; Sleiman et al., 2013). Reactivity with O_3 and NO_3^{\bullet} radicals is expected to be not important in the 84 85 atmosphere (Calvert et al., 2011). Hydroxyketone lifetimes with respect to OH* radicals are in the range of about 1 day (e.g. for 4-hydroxy-2-butanone) to several 86 87 days (e.g. 14 days for 3-hydroxy-3-methyl-2-butanone) (Aschmann et al., 2000; 88 Calvert et al., 2011), thus showing large variability in the reactivity that can be 89 attributed to the structure of the molecule. Recently, the determination of the absorption cross-sections for four hydroxycarbonyl (≥ C4) compounds by (Messaadia 90 91 et al., 2012) showed that all obtained spectra of these compounds were similar in 92 shape with two absorption bands, the first one below 220 nm and the second one 93 between 230 and 340 nm with a maximum in the wavelength range 265-290 nm. The 94 tropospheric lifetime of 3H3M2B due to photolysis was estimated by Messaadia et al. 95 (2012) to be at least 0.4 day using an upper limit of the photolysis rate assuming a 96 quantum yield of unity. Therefore, atmospheric fate of these compounds may also be 97 significantly affected by photolytic loss processes. As a consequence, a more 98 detailed characterization of the photochemical fate of hydroxyketone compounds is 99 clearly needed to make further progress in our understanding of the implication of 100 these species in tropospheric chemistry and allow further extension of atmospheric 101 chemistry models (especially for hydrocarbons) for improvement of air quality 102 predictions.

103 The objective of the present work was to evaluate the importance of the 104 photolysis and the OH-initiated reaction in the case of 3-hydroxy-3-methyl-2-105 butanone (3H3M2B), ((CH_3)₂C(OH)C(O)CH₃). This oxygenated VOC is an oxidation 106 product of the OH-initiated reaction of 2-methyl-2-butene (Tuazon et al., 1998). It has 107 been observed in a measurement campaign carried out at San Francisco (Destaillats 108 et al., 2002). In the atmosphere, like other carbonyl compounds, 3H3M2B is 109 expected to be removed by chemical reactions with the atmospheric photooxidants 110 and by photolysis. The atmospheric degradation of this species is not well known 111 however. In fact, only one kinetic study on the gas phase reaction of 3H3M2B is 112 found in the literature (Aschmann et al., 2000a). In this study, the degradation of 113 3H3M2B due to its reaction with OH, NO₃ and O₃ was only investigated at room 114 temperature and atmospheric pressure using a relative rate technique.

In order to elucidate the atmospheric fate of 3H3M2B, photolysis and kinetic studies as a function of temperature are necessary. In this work, photolysis of this species was studied from 270 to 390 nm and major products were analysed. Moreover, the OH-initiated reaction of 3H3M2B was studied and rate constants were determined as a function of temperature.

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121 **2. Experimental section**

Experiments have been carried out using two different experimental set-ups, a simulation chamber made of Teflon film, located at the SAGE department of Mines Douai, and a temperature controlled Pyrex simulation chamber, located at University of Reims.

126 2.1 Photolysis study (Mines Douai)

127 2.1.1 Simulation chamber description and experimental conditions

128 The experimental procedure and equipments used for the photolysis 129 experiments performed are briefly described in the following. More details on the 130 experimental setup can be found in previous publications (Turpin et al., 2006; Szabo 131 et al., 2009; Szabo et al., 2011). Experiments were carried out in a ~ 300 L FEP 132 Teflon film chamber in an air-conditioned laboratory (296 K) and at 1 atm. 133 Fluorescent tubes emitting in the 270-390 nm region with an energy peak at 312 nm 134 were used for irradiation. In a typical experiment, a known amount of 3H3M2B was 135 introduced in the chamber through a stream of purified air and the reaction mixture is allowed to stand for about 1 h in the dark. Three samples were then analysed to 136 137 determine the initial concentration of the reactant. After the third sampling, the lamps 138 are switched on and about 10 samples are taken along the photolysis experiments 139 (lasting 6-8h) to monitor the concentrations of 3H3M2B and end-products over the time. 3H3M2B initial concentration was in the range of 9 to 50 ppm. Some 140 141 experiments were conducted in the presence of excess cyclohexane (151-227 ppm), 142 cyclopentane (280 ppm), m-xylene (33-280 ppm) or carbon monoxide (1330-3750 143 ppm) in order to scavenge any OH radicals formed in the reaction mechanism.

144 **2.1.2 Test experiments**

145 Test experiments were performed to investigate possible losses of 3H3M2B 146 and products in the dark and during the photolysis experiments. First, deposition 147 rates on the chamber walls were tested for by running experiments in the dark. 148 Results showed that such losses were negligible for 3H3M2B, CH₂O, acetone, 149 biacetyl, and methanol (rates < 1% per hour). In addition, tests for possible losses of products through photolysis were carried out. Results indicated that photolysis could 150 151 be significant in the case of CH₂O, acetone and biacetyl with loss rates of 31%, 7% 152 and 4% per hour, respectively. In the following, all the reported concentrations of 153 CH₂O, acetone and biacetyl have been corrected for wall deposition and photolysis 154 according to the rates determined in the test experiments. Formaldehyde was also 155 corrected for production from acetone photolysis.

156 **2.1.3 Sampling and analytical techniques**

3H3M2B and photolysis product analysis were performed either by a 157 158 ThermoDesorption-Gas phase Chromatograph (TD-GC) coupled with Fourier Transform Infrared (FTIR) spectroscopy and Flame Ionization Detection (FID) or by 159 160 direct FTIR spectrometry using a White cell. The FTIR-White cell analytical system 161 was similar to that used in the study carried out by (Bouzidi et al., 2014). The IR 162 spectra typically result from the co-addition of 100 interferograms with a resolution of 163 2 cm⁻¹ representing a measurement time of about 4 min. In the case of the TD-GC-FTIR-FID, gas samples of 100 cm³ volume were collected every 45 min on Tenax 164 adsorbent microtrap cooled down to -30℃. Injection was achieved by rapid heating 165 166 of the microtrap to 300°C. Both analytical systems were connected to the chamber through a Teflon line heated to ~338 K. Some samples were also collected onto 167 168 Tenax solid adsorbent and analysed by TD-GC-FID-Mass Spectrometry. Reactant 169 and product concentrations were monitored using classical calibration procedures. Carbonyl products were analyzed following the same procedure as that described by 170 171 (Bouzidi et al., 2014) using 2,4-dinitrophenylhydrazine (DNPH) derivatization followed 172 by HPLC-UV analysis.

173 **2.1.4 Actinometry**

174 Acetone and acetaldehyde were used as actinometers in order to estimate the 175 tropospheric photolysis lifetime and the effective quantum yields, as they have 176 relatively similar UV absorption spectra compared to 3H3M2B (Messaadia et al., 177 2012). The photolysis frequencies of acetone and acetaldehyde in the simulation chamber in the presence of cyclopentane were (0.070 ± 0.01) h⁻¹ and (0.22 ± 0.02) 178 179 h⁻¹, respectively. Acetone photolysis frequency calculated from the recommended 180 values of cross-sections and quantum yields with a solar zenith angle of 20° within the lower troposphere for a cloudless day is about 0.00216 h^{-1} (Calvert et al., 2011). 181 182 For acetaldehyde, a photolysis frequency of 0.0104 h⁻¹ was measured at the Euphore chamber for a solar zenith angle of 20° (Calvert et al., 2011). 183

184 **2.1.5 Chemicals**

3H3M2B (> 95%), 2,3-butanedione (> 97%), cyclopentane (> 99%),
cyclohexane (> 99%), m-xylène (> 99%) and acetone (> 99%) were purchased from
Sigma-Aldrich, acetic acid (> 99%) from Acros Organics and methanol from Fluka (>
99%). Dry air was produced by a zero air generator (Claind AZ 2020). Carbon
monoxide (4939 ppmv in N₂) was obtained from Praxair.

190 **2.2 Oxidation by OH-radicals (GSMA-Reims)**

191 2.2.1 Experimental set-up

192 The kinetics of the reaction of 3H3M2B with OH radicals was studied using a 193 photo-chemical reactor coupled to a FTIR spectrometer. The set-up has been 194 presented in details in a previous publication (Messaadia et al., 2013). Therefore, it 195 will be described only briefly herein. The chamber is composed of a triple-jacket 196 Pyrex cell (length of 2 m, internal diameter of 20 cm and total volume of 63 L). Multi-197 reflection was achieved through the use of gold-plated mirrors suitable for working 198 with IR radiation, and was used to vary the path-length between 4 and 80 m. A 199 primary pump was used to suck up the chamber to 10⁻³ mbar. The temperature in the 200 chamber was regulated through the circulation of a thermostatic fluid (water or 201 ethanol) between the inner and middle jackets. The fluid temperature and circulation 202 were commanded by a Julabo FPW 90 thermostat. The working temperature range 203 was from 298 K to 356 K. The temperature and the pressure in the chamber were 204 respectively measured by a thermocouple and a MKS Baratron manometer with 0-205 1000 Torr full scale. 24 UV lamps emitting in the range of 300 to 400 nm were

206 symmetrically disposed around the chamber in order to ensure homogeneous 207 photolysis and to generate OH-radicals from nitrous acid. The extremities of the 208 chamber were sealed by inox plates which hold two optical windows made of ZnSe 209 as well as the ports used for the introduction of the reacting species into the reaction 210 medium.

211 An Equinox 55 FT-IR spectrometer provided by Bruker was used to monitor 212 the concentration of the reactants and reference compounds inside the reaction 213 medium. This FTIR spectrometer is equipped with a Globar IR source, a KBr beam-214 splitter, and two detectors (DTGS and MCT). Its operating spectral range varies between 600 and 4000 cm⁻¹, with a spectral resolution of 2 to 0.5 cm⁻¹. MCT was 215 216 used as a detector in this study. The alignment of the IR beam inside the chamber 217 was accomplished by using a He-Ne laser beam by adjusting the positions of the 218 gold mirrors.

219 **2.2.2 Experimental conditions**

220 All experiments were performed in purified air. The OH reaction of 3H3M2B 221 was studied relatively to tert-butyl alcohol (tert-C₄H₉OH) as a reference compound 222 whose homogeneous reactivity towards OH-radicals is known. The OH radicals were 223 generated by the photolysis of nitrous acid at 300-400 nm. Nitrous acid (HONO) was 224 produced via the drop-wise addition of a 10 % sulphuric acid solution to a 0.2 M 225 sodium nitrite solution. A flow of nitrogen gas was used to sweep along the 226 generated acid into the chamber in the vapour phase. The used reagents were 227 obtained from the following sources: Purified Air (99.999%) provided by Air Liquide, 228 3H3M2B (95%) provided by Sigma-Aldrich and tert-butyl alcohol (> 99.7%) provided 229 by Fluka. They were further purified by repeated freeze-pump-thaw cycles before 230 use.

231 Before starting the kinetic measurements, two tests were carried out: the first 232 one aimed at evaluating the loss of 3H3M2B and the reference compound by 233 photolysis in the absence of HONO. The second one consisted on monitoring the 234 concentration of 3H3M2B and tert-C₄H₉OH in the absence of irradiations and in the 235 presence of HONO to check for the stability of the reagents. These tests showed that 236 photolysis and wall losses were negligible for both compounds. At each temperature, 237 2 to 3 independent experiments were carried out. Table 1 summarizes the 238 experimental conditions.

The reference was chosen in such a way that at least one absorption band of the 3H3M2B does not exhibit any interference with those of the reference compound and vice versa, and that the OH-degradation rate constants of both, 3H3M2B and reference, are of the same order of magnitude. The rate constant of the reaction of tert— C_4H_9OH with OH used in this work is that reported by (Téton et al., 1996):

244 $k_{ref} = (2.66 \pm 0.48) \times 10^{-12} \exp((-2240 \pm 1080)/RT) \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$ (I)

245

246 **3. Results and discussion**

247 **3.1 Photolysis**

248 **3.1.1 Photolysis rate constants**

The photolysis of 3H3M2B was carried out either in the absence or in the presence of excess of OH[•] radical scavenger. The photolysis rate constants J_{3H3M2B} were determined by plotting the natural logarithm of the ratio [3H3M2B]₀/[3H3M2B]_t versus time:

253
$$Ln \frac{[3H3M2B]_{0}}{[3H3M2B]_{t}} = J_{3H3M2B} \times t$$
 (II)

254 where [3H3M2B]₀ and [3H3M2B]_t indicate the 3H3M2B concentrations at the initial 255 and time t, respectively.

256 The 3H3M2B concentration data were well fitted to a straight line using a linear least-square procedure (Figure 1) and the slope of the linear regression 257 258 (weighted by 1σ) represents J_{3H3M2B} . The experimental conditions and the photolysis frequencies obtained for 3H3M2B in the presence and absence of OH[•] scavenger 259 260 are summarized in Table 2.

261 3H3M2B photolysis frequencies were found to be (0.24 ± 0.02) h⁻¹ and (0.33 ± 0.02) 0.02) h⁻¹ in the presence and absence of OH[•] scavenger, respectively. The reported 262 uncertainties represent one standard deviation of the slope of the regression line. In 263 the absence of OH[•] scavenger, an increase of around 40% in J_{3H3M2B} is noted. Using 264 the average OH-radical rate constant obtained in the present work, k_{OH+3H3M2B} = 265 7.6×10⁻¹³ cm³ molecule⁻¹ s⁻¹ (see below) enables to estimate an OH[•] photostationary 266 concentration of around 3×10^7 radical cm⁻³ from the observed difference in the 267 photolysis frequencies. This suggests that OH[•] radicals were produced in the 268 chemical mechanism, which is confirmed by the observation of cyclopentanol and 269 cyclopentanone (OH[•] + cyclopentane reaction products) in the presence of excess 270 271 cyclopentane, or cyclohexanol and cyclohexanone (OH[•] + cyclohexane reaction products) in the presence of cyclohexane. 272

273 3.1.2 Effective guantum yields

274 From the obtained photolysis frequencies, the effective quantum yield of the 275 3H3M2B photolysis can be calculated using acetone and acetaldehyde actinometry 276 experiments according to the following equation (Bacher et al., 2001; Bouzidi et al., 277 2014):

 $\Phi_{_{3H3M2B}} = \frac{J_{_{3H3M2B}}/J_{_{actin}}}{J_{_{3H3M2B}}^{_{calc}}/J_{_{actin}}^{_{calc}}}$ 278

(III)

 J_{3H3M2B}^{calc} and J_{actin}^{calc} are the 3H3M2B and actinometer (acetone or acetaldehyde) 279 280 photolysis frequencies, respectively, calculated using:

281
$$J_{3H3M2B}^{calc} = \int_{\lambda} \phi_{3H3M2B}(\lambda) \times \sigma_{3H3M2B}(\lambda) \times F(\lambda) d\lambda$$
 (IV)
282 and $J_{actin}^{calc} = \int \phi_{actin}(\lambda) \times \sigma_{actin}(\lambda) \times F(\lambda) d\lambda$ (V)

282 and
$$J_{actin}^{calc} = \int_{\lambda} \phi_{actin}(\lambda) \times \sigma_{actin}(\lambda) \times F(\lambda) d\lambda$$

where $\Phi_{3H3M2B}(\lambda)$ and $\Phi_{actin}(\lambda)$ represent the 3H3M2B and actinometer wavelength-283 284 dependent quantum yields, respectively. $\sigma_{3H3M2B}(\lambda)$ and $\sigma_{actin}(\lambda)$ are the 3H3M2B and 285 actinometer absorption cross-sections, respectively and $F(\lambda)$ the relative actinic flux intensity. For 3H3M2B, cross-sections were from (Messaadia et al., 2013) and 286 287 quantum yields in Equation (IV) were set to unity. $\sigma_{acetone}(\lambda)$, $\Phi_{acetone}(\lambda)$, $\sigma_{acetaldehyde}(\lambda)$ and $\Phi_{acetaldehyde}(\lambda)$ were all from the last evaluation from IUPAC (IUPAC, 2013). Using 288 289 Equation (III), the 3H3M2B effective quantum yields are: $\Phi_{3H3M2B} = 0.85 \pm 0.30$ with acetone as actinometer and $\Phi_{3H_{3M2B}} = 0.76 \pm 0.20$ with acetaldehyde as actinometer. 290 291 An average quantum yield of $\Phi_{3H3M2B} = 0.81 \pm 0.30$ is thus recommended over the 292 wavelength range 270-390 nm.

293 The main sources of uncertainties are the 3H3M2B and actinometer 294 absorption cross-sections, the acetone and acetaldehyde quantum yields and the 295 measured 3H3M2B, acetone and acetaldehyde photolysis frequencies. Using the 296 propagation of errors yields an overall uncertainty of 35% and 28% on the 3H3M2B 297 quantum yields with acetone and acetaldehyde, respectively. The fairly large 298 effective 3H3M2B quantum yields are comparable to those obtained for hydroxyacetone (Orlando et al., 1999): $\Phi_{hydroxyacetone} = 0.65 \pm 0.25$ for 240-420 nm 299

300 and $\Phi_{hvdroxvacetone} < 0.6$ for $\lambda > 290$ nm. Note that these quantum yields values are the 301 only data for hydroxyketone species known in the literature. 302 3.1.3 Photolysis products 303 Product analysis was carried out to unravel the mechanism of 3H3M2B 304 photolysis. Three carbon-carbon bond dissociation pathways are possible: 305 $(CH_3)_2C(OH)C(O)CH_3 + h\nu \rightarrow (CH_3)_2C(OH)C(O)^{\bullet} + CH_3^{\bullet}$ (1a) $(CH_3)_2C(OH)C(O)CH_3 + hv \rightarrow (CH_3)_2C(OH)^{\bullet} + {}^{\bullet}C(O)CH_3$ 306 (1b) $(CH_3)_2C(OH)C(O)CH_3 + h\nu \rightarrow CH_3^{\bullet} + (CH_3)C^{\bullet}(OH)C(O)CH_3$ 307 (1c) In addition, the direct OH[•] release from the carbon – hydroxy group dissociation is 308 309 also possible: 310 $(CH_3)_2C(OH)C(O)CH_3 + hv \rightarrow (CH_3)_2C^{\bullet}C(O)CH_3 + OH^{\bullet}$ (1d) 311 Five major reaction products - acetone, acetic acid, formaldehyde, CO and 312 methanol - were observed. Figure 2 shows reactant and product time profiles 313 obtained in the presence of m-xylene as OH-scavenger. The shape of the profiles for 314 acetone, formaldehyde and methanol displays non-zero derivative at the origin, 315 indicating that these compounds are primary products. On the contrary, carbon 316 monoxide and acetic acid present secondary product profiles with a delay time before 317 their concentrations increase. Yield plots in the form of [products] vs. [3H3M2B]_{reacted} are shown in Figure 3 318 319 for the primary products: linear profiles were obtained for acetone and formaldehyde 320 while a slightly increasing curved profile was obtained for methanol. Table 2 321 summarizes the corresponding corrected yields obtained in the photolysis of 322 3H3M2B with and without OH[•] scavenger. Acetone (CH₃)₂CO is expected to originate from channels (1a), (1b) and/or 323 (1d) through the following reaction sequences (where RO_2^{\bullet} represents a peroxy 324 325 radical): 326 (1a) $(CH_3)_2C(OH)C(O)^{\bullet} + O_2 \rightarrow (CH_3)_2C(OH)C(O)O_2^{\bullet}$ 327 328 $(CH_3)_2C(OH)C(O)O_2^{\bullet} + RO_2^{\bullet} \rightarrow (CH_3)_2C(OH)^{\bullet} + CO_2 + RO^{\bullet} + O_2$ 329 $(CH_3)_2C(OH)^{\bullet} + O_2 \rightarrow (CH_3)_2CO + HO_2^{\bullet}$ 330 (1b) $(CH_3)_2C(OH)^{\bullet} + O_2 \rightarrow (CH_3)_2CO + HO_2^{\bullet}$ 331 332 (1d) $(CH_3)_2C^{\bullet}C(O)CH_3 + O_2 \rightarrow (CH_3)_2C(O_2)^{\bullet}C(O)CH_3$ 333 334 $(CH_3)_2C(O_2)^{\bullet}C(O)CH_3 + RO_2^{\bullet} \rightarrow (CH_3)_2CO + CH_3CO^{\bullet} + RO^{\bullet} + O_2$ 335 As an acetone yield of nearly 1 is obtained, it is concluded that the branching 336 to channel (1c) (which is not supposed to yield acetone) is rather low (< 2%). 337 Formaldehyde arises from the radical recombination of $CH_3O_2^{\bullet}$: for example, the CH₃O₂[•] self reaction leads to CH₂O: 338 $CH_3O_2^{\bullet} + CH_3O_2^{\bullet} \rightarrow 2 CH_3O^{\bullet} + O_2$ 339 (2a) $CH_{3}O_{2}^{\bullet} + CH_{3}O_{2}^{\bullet} \rightarrow CH_{3}OH + CH_{2}O + O_{2}$ 340 (2b) with a branching ratio $k_{2b}/k_2 = 0.63$ recommended by (Sander et al., 2011). 341 The CH₂O yield of 23% (corrected for wall loss, photolysis and production from 342 343 acetone photolysis) - well below 100% - indicates that a fraction of the $CH_3O_2^{\bullet}$ radicals will react with HO2[•] in a terminating channel to form essentially methyl-344 hydroperoxide (Orlando et al., 2012) and that some of the CH₃C(O)O₂[•] radicals will 345 be lost through the $CH_3C(O)O_2^{\bullet} + HO_2^{\bullet}$ reaction which gives only 0.44 $CH_3O_2^{\bullet}$ 346 (Dillon et al., 2008). Note that this last reaction is known to produce OH radicals with 347 348 branching ratios between 0.4 and 0.6 (Gross et al., 2014) and may represent the sole source of OH radicals in the 3H3M2B photolysis. Noteworthy the CH₂O yield 349

decreases from the conditions with OH-scavenger (23%) to the conditions without OH-scavenger (16%) probably due to the $OH^{\bullet} + CH_2O$ reaction.

352 Methanol mostly originates from the molecular channel of $CH_3O_2^{\bullet}$ peroxy 353 radical reactions:

354 $CH_3O_2^{\bullet} + RO_2^{\bullet} \rightarrow CH_3OH + R_{-H}O + O_2$

Results from experiments with m-xylene and CO as OH[•] scavenger showed a higher methanol yield with m-xylene (11%) than with CO (6%), revealing that the additional RO₂[•] radicals from m-xylene + OH[•] convert more CH₃O₂[•] into CH₃OH (as CO + OH[•] leads to HO₂[•] + CO₂ and no RO₂[•]).

In addition to the three major primary products, carbon monoxide and acetic 359 360 acid (AA) were also observed in the present study. The plots [CO] and [AA] vs. [3H3M2B]_{reacted} are not linear but follow roughly a second order polynomial with a 361 362 zero derivative at the origin, indicating the occurrence of only secondary sources both in presence and in absence of OH[•] scavenger (Figure 4). Direct formation of CO 363 from 3H3M2B photolysis through $(CH_3)_2C(OH)C(O)CH_3 + hv \rightarrow (CH_3)_2C(OH)^{\bullet} + CO$ 364 + CH_3^{\bullet} is thus excluded. It should be stressed that a larger amount of CO is 365 quantified in the absence of OH[•] scavenger. The secondary formation of CO may 366 367 thus be attributed essentially to formaldehyde photolysis and reaction with OH[•].

368 Similarly, the absence of acetic acid formation at the beginning of the reaction 369 (Figure 4) suggests a unique secondary origin for this product. As AA is produced 370 through CH₃C(O)O₂[•] + HO₂[•]/RO₂[•] reactions (Tomas et al., 2001; Sander et al., 2011), we conclude that the photodissociation channels (1b) and (1d) (both producing 371 $CH_3C(O)O_2^{\bullet}$) are minor channels. Assuming that AA only comes through the 372 CH₃C(O)O₂[•] + HO₂[•] reaction with a branching ratio of 0.2 (Tomas et al., 2001) and 373 374 using an AA detection limit of 50 ppb enable a rough estimation of the maximum 375 branching ratio for the (1b) + (1d) channels of 5% to be calculated.

The secondary formation of acetic acid can be primarily attributed to the photolysis of acetone (Calvert et al., 2011):

378 $(CH_3)_2CO + hv (+ O_2) \rightarrow CH_3C(O)O_2^{\bullet} + CH_3O_2^{\bullet}$

followed by peroxy radical reactions of CH₃C(O)O₂[•] (Tomas et al., 2001). The 379 380 heterogeneous formation of acetic acid in the OH[•] + acetone reaction reported by 381 (Turpin et al., 2006) is probably not significant in the present conditions, as more AA is observed when OH radicals are scavenged. We also suspect the OH[•] + m-xylene 382 383 reaction that produces methylglyoxal with yields of about 40% (Calvert et al., 2002). 384 Methylglyoxal can readily be photolysed into $CH_3CO^{\bullet} + HCO^{\bullet}$, giving $CH_3C(O)O_2^{\bullet}$, HO_2^{\bullet} and CO. This source could explain the large difference in the AA formed with 385 386 m-xylene as OH-scavenger compared to that formed in the absence of OH-387 scavenger (see Figure 4). This result underlines the role that can be played by a 388 scavenger in modifying the type and amount of peroxy radicals in the investigated 389 chemical system, and thus the products observed.

390 Minor products were also identified: 2,3-butanedione (or biacetyl, 391 $CH_3C(O)C(O)CH_3)$, hydroxyacetone and methylglyoxal. Biacetyl yields of 1.4% and 392 0.4% have been quantified in the absence and presence of OH scavengers. This α -393 diketone is expected to originate from the photolysis channel (1c). The increase in 394 the biacetyl yields between OH[•] and no OH[•] conditions may be attributed to a weak 395 biacetyl formation through OH[•] + 3H3M2B. Finally, hydroxyacetone and 396 methylglyoxal were also tentatively identified and guantified with very low yields (< 397 0.5%).

398 Overall, the carbon balance obtained in the present work represents nearly 399 80%. The lacking products are assumed to be mainly peroxides formed by RO_2^{\bullet} + 400 HO_2^{\bullet} reactions (Orlando et al., 2012).

401 **3.1.4 Mechanism interpretation**

402 Based on the reaction product observations, it is possible to propose a 403 simplified photolysis mechanism of 3H3M2B as shown in the scheme displayed in 404 Figure 5. The high yields of acetone (98%) coupled with negligible amounts of CO 405 and acetic acid at the beginning of the experiments in the presence of OH[•] 406 scavenger indicates that the photolysis of 3H3M2B proceeds mainly through the 407 rupture of the C1-C2 bond of the molecule (channel (1a)). Photolysis channels (1b) 408 and (1d), both producing acetic acid, and channel (1c) leading to 2,3-butanedione, 409 respectively, are of minor importance.

410 **3.2 OH reaction kinetics**

The kinetics of the reaction of OH radicals with 3H3M2B was determined relatively to that of a reference compound using the following equation:

(VI)

413
$$\operatorname{In}\left(\frac{[\operatorname{3H3M2B}]_{t_0}}{[\operatorname{3H3M2B}]_t}\right) = \operatorname{R} \times \operatorname{In}\left(\frac{[\operatorname{ref}]_{t_0}}{[\operatorname{ref}]_t}\right)$$

where [3H3M2B]_{t0} and [ref]_{t0} are the initial concentrations of 3H3M2B and tert-butyl 414 415 alcohol at time t₀, respectively; [3H3M2B]_t and [ref]_t, are the concentrations of 416 3H3M2B and tert-butyl alcohol at time t, respectively and R is the ratio of rate 417 constants (R = k_{3H3M2B}/k_{ref}) where k_{ref} is the rate constant of the reference reaction. A plot of (ln[3H3M2B]_{t0}/[3H3M2B]_t) as a function of (ln[ref]_{t0}/[ref]_t) results in a straight 418 419 line with the slope equal to R. Knowing k_{ref} , it is possible to determine k_{3H3M2B} , the 420 rate constant of the reaction of OH-radicals with 3H3M2B. During an experiment, IR 421 spectra were recorded every 5 minutes. Each spectrum constitutes the average of 20 422 accumulated acquisitions leading to a signal to noise ratio of ~15. The obtained data 423 are presented in the form of ln([3H3M2B]_{t0}/[3H3M2B]_t) vs. ln([ref]_{t0}/[ref]_t). As can be 424 seen on Figure 6, good linearity is observed with correlation coefficients greater than 425 0.91. The rate constants obtained in the present study are reported in Table 3. The 426 given rates are derived from the average of 2 to 4 independent measurements. At 427 298 K, the average rate constant for the reaction of 3H3M2B with OH radicals is: $k_{3H3M2B} = (7.6 \pm 0.5) \times 10^{-13} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}.$ 428

429 For each experiment, the overall error in the rate constant values reported in 430 this work is calculated using propagation of error according to the following relation:

431
$$\Delta k_{3H3M2B} = k_{3H3M2B} \sqrt{(\Delta k_{ref} / k_{ref})^2 + (\Delta R / R)^2}$$

(i)

432 where k_{ref} , Δk_{ref} are the rate constant of the reference reaction and the uncertainty on 433 this parameter, respectively and R = k_{3H3M2B}/k_{ref} is the slope of the plot according to 434 Equation (VI) with ΔR the uncertainty on this value. Errors originate mainly from:

435 436 the determination of k_{ref}: this error varies from 10% to 20% according to the temperature (Téton et al., 1996);

(VII)

(ii) (ii) the determination of the slope R: this parameter is determined directly from the experimental data. The uncertainty on this parameter is related to the areas corresponding to the absorption bands of 3H3M2B and the reference compound. The standard deviation in R values obtained for each set of experiments varies between 10% and 25% (Table 3), which indicates that the experiments are fairly reproducible.

The Arrhenius plot derived from the experimental rate constants (Table 3) is presented in Figure 7 and shows a good linearity. The corresponding Arrhenius equation is:

447 $k_{3M3H2B}(T) = (5.12\pm0.07) \times 10^{-12} exp(-563\pm119/T) cm^3$ molecule⁻¹ s⁻¹ 448 In contrast to other hydroxyketones studied in our laboratory (Messaadia et al., 449 2013), the rate constant k_{3H3M2B} exhibits a slight positive temperature dependence. 450 However, this temperature coefficient remains relatively weak (E/R around 600 K). 451 This indicates that this reaction may proceeds initially via the reversible formation of 452 unstable intermediate complexes.

Until now, no kinetic data were reported in the literature for the reaction of 3H3M2B with OH radicals as a function of temperature. Only one study regarding the kinetics of OH-oxidation of 3H3M2B exists but only at room temperature (Aschmann et al., 2000). Our value of $(7.6\pm0.5)\times10^{-13}$ cm³ molecule⁻¹ s⁻¹ is in good agreement with this determination (of $(9.4\pm3.7)\times10^{-13}$ cm³ molecule⁻¹ s⁻¹) with a difference less than 20%.

459 It is also possible to compare the reactivity of 3H3M2B to that of other 460 hydroxyketones for which the hydroxyl group is attached to the carbon atom located 461 in the α position with respect to the carbonyl group, namely 1-hydroxy-2-butanone, 3hydroxy-2-butanone and 4-hydroxy-3-hexanone. The kinetic rate constants of the 462 gas-phase reaction of OH radicals with these compounds are (cm³ molecule⁻¹ s⁻¹): (7.7 \pm 0.7) x 10⁻¹² (Aschmann et al., 2000), (9.60 \pm 0.30) x 10⁻¹² (Messaadia et al., 463 464 2013) and (15.1 ± 3.1) x 10⁻¹² (Aschmann et al., 2000) for 1-hydroxy-2-butanone, 3-465 hydroxy-2-butanone and 4-hydroxy-3-hexanone, respectively. As expected, the 466 467 reactivity of 3H3M2B towards OH radicals is lower than that of the other α -468 hydroxyketones by at least one order of magnitude. This lower reactivity can be 469 attributed to the lack of secondary or tertiary H-atoms in the structure of the 3H3M2B 470 molecule. 471

472 **4. Atmospheric implications**

The photolysis frequency of 3H3M2B determined in the present work is specific to the experimental setup used and cannot be directly applied for the estimation of the atmospheric lifetime. However, the 3H3M2B atmospheric photolysis frequency (J_{3H3M2B}^{atm}) can be estimated using the known photolysis frequencies of actinometers (acetone and acetaldehyde) in the troposphere (J_{actin}^{atm}) through the following equation:

478
$$J_{3H3M2B}^{atm} = J_{3H3M2B} \times \frac{J_{actin}^{atm}}{J_{actin}}$$

(VIII)

479 Taking into account the experimental photolysis rate constants of 3H3M2B 480 J_{3H3M2B} and actinometer J_{actin} measured in the reaction chamber allows to estimate an average photolysis frequency of 3H3M2B in the troposphere of 481 $J_{3H3M2B}^{atm} = 0.0095 \ h^{-1}$. A global uncertainty of about 70% is associated with J_{3H3M2B}^{atm} 482 483 principally due to the uncertainties on J_{actin} of at least 50% (Calvert et al., 2011). Furthermore, using the absorption cross sections of 3H3M2B (Messaadia et al., 484 2012) and the actinic flux at 40 N, 1 st July and 20° zenithal angle enables calculating 485 486 an upper limit of the 3H3M2B photolysis frequency J_{3H3M2B,max} of 0.094 h⁻¹ in the 487 atmosphere (assuming a 3H3M2B quantum yield of 1 over the whole absorption range). Finally, comparing the estimated and maximum J_{3H3M2B} values yields an 488 effective quantum yield of $\Phi = J_{3H3M2B}^{atm} / J_{3H3M2B,max} = 0.0095 / 0.094 \approx 0.10$ for 3H3M2B 489 490 in the atmosphere with around 80% combined uncertainty.

491 Data on atmospheric photolysis frequencies of hydroxyketones are very 492 scarce in the literature. The estimated photolysis frequency of 3H3M2B in the 493 troposphere is roughly a factor two higher than the J value for hydroxyacetone (HA): 494 $J_{HA}^{atm} \approx (0.0054 \pm 0.0018) h^{-1}$ while a factor two lower than the higher limit determined for 4-hydroxy-4-methyl-2-pentanone (4H4M2P) $J_{4H4M2P}^{atm} \leq 0.018 \ h^{-1}$ (Calvert et al., 495 2011). The atmospheric lifetime of 3H3M2B can then be calculated according to: 496

 $\tau^{\text{atm}}_{\text{3H3M2B}} =$ 497

1

(IX)

 $\mathsf{J}_{_{3H3M2B}}^{atm}$ 498 The calculated lifetime of 3H3M2B with respect to photolysis is about 4 - 5 499 days. An atmospheric lifetime of 15 days due to reaction with OH* radicals is estimated according to the equation $\tau = 1/k_{3H3M2B} \times [OH^{\bullet}]$ by using a 24 h daytime 500 average global tropospheric OH[•] concentration of 1×10^{6} molecule cm⁻³ (Atkinson et 501 502 al., 1997). Thus, it is clear that photolysis is the major oxidation pathway for 3H3M2B 503 in the gas-phase. It should be noted that other heterogeneous loss processes for 504 3H3M2B including physical processes (dry deposition) and uptake by clouds followed 505 by wet deposition may be important. Regional impacts should thus be expected for 506 3H3M2B: photolysis of 3H3M2B represents a source of acetone and formaldehyde, 507 affecting the atmospheric oxidative capacity and the tropospheric ozone budget.

508 The data obtained in this work suggest that the reactivity of hydroxycarbonyls may be very sensitive to their chemical structures. While the OH[•] reaction is 509 510 generally the major tropospheric loss process for hydroxycarbonyls, the absence of 511 H-atom in the β position to the carbonyl group of 3H3M2B strongly depletes its 512 reactivity towards OH[•] radicals and consequently enhances the importance of the 513 photolysis channel in the atmosphere.

514

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526 References

- 527 Aschmann S. M., Arey J., Atkinson R. (2000). Atmospheric chemistry of selected hydroxycarbonyls. Journal of Physical Chemistry A 104, 3998-4003. 528
- 529 Atkinson R., Arey J. (2003). Gas-phase tropospheric chemistry of biogenic volatile 530 organic compounds: a review. Atmospheric Environment 37, S197-S219.
- 531 Atkinson R., Baulch D. L., Cox R. A., Hampson R. F., Kerr J. A., Rossi M. J., Troe J. 532 (1997). Evaluated Kinetic and Photochemical Data for Atmospheric Chemistry: Supplement VI. IUPAC Subcommittee on Gas Kinetic Data Evaluation for 533 534 Atmospheric Chemistry. Journal of Physical and Chemical Reference Data 26, 535 1329-1499.
- Aumont B., Szopa S., Madronich S. (2005). Modelling the evolution of organic carbon 536 537 during its gas-phase tropospheric oxidation: development of an explicit model

538	based on a self generating approach. Atmospheric Chemistry and Physics 5,
539	2497-2517.
540 541	Bacher C., Tyndall G. S., Orlando J. J. (2001). The atmospheric chemistry of glycolaldehyde. Journal of Atmospheric Chemistry 39, 171-189.
541 542	Bethel H. L., Atkinson R., Arey J. (2003). Hydroxycarbonyl Products of the Reactions
542 543	of Selected Diols with the OH Radical. The Journal of Physical Chemistry A
545 544	107, 6200-6205.
545	Bouzidi H., Fittschen C., Coddeville P., Tomas A. (2014). Photolysis of 2,3-
546	pentanedione and 2,3-hexanedione: kinetics, quantum yields, and product
547	study in a simulation chamber. Atmospheric Environment 82, 250-257.
548	Calvert J. G., Atkinson R., Becker K. H., Kamens R. M., Seinfeld J. H., Wallington T.
549	J., Yarwood G. (2002). The mechanisms of atmospheric oxidation of aromatic
550	hydrocarbons. New York, Oxford University Press.
551	Calvert J. G., Mellouki A., Orlando J. J., Pilling M. J., Wallington T. J. (2011). The
552	mechanisms of atmospheric oxidation of the oxygenates. New York, Oxford
553	University Press.
554	Destaillats H., Spaulding R. S., Charles M. J. (2002). Ambient Air Measurement of
555	Acrolein and Other Carbonyls at the Oakland-San Francisco Bay Bridge Toll
556	Plaza. Environmental Science & Technology 36, 2227-2235.
557	Dillon T. J., Crowley J. N. (2008). Direct detection of OH formation in the reactions of
558	HO_2 with $CH_3C(O)O_2$ and other substituted peroxy radicals. Atmospheric
559	Chemistry and Physics 8, 4877-4889.
560	Dillon T. J., Horowitz A., Hölscher D., Crowley J. N., Vereecken L., Peeters J. (2006).
561	Reaction of HO with hydroxyacetone (HOCH ₂ C(O)CH ₃): rate coefficients
562	(233–363 K) and mechanism. Physical Chemistry Chemical Physics 8, 236–
563	246.
564	El Dib G., Sleiman C., Canosa A., Travers D., Courbe J., Sawaya T., Mokbel I.,
565	Chakir A. (2013). First experimental determination of the absolute gas-phase
566	rate coefficient for the reaction of OH with 4-hydroxy-2-butanone (4H2B) at
567	294 K by vapor pressure measurements of 4H2B. Journal of Physical
568	Chemistry A 117, 117-125.
569	Goldstein A. H., Galbally I. E. (2007). Known and unexplored organic constituents in
570	the Earth's atmosphere. Environmental Science and Technology March 1, 1515-1521.
571 572	Gross C. B. M., Dillon T. J., Schuster G., Lelieveld J., Crowley J. N. (2014). Direct
572	kinetic study of OH and O_3 formation in the reaction of CH ₃ C(O)O ₂ with HO ₂ .
575 574	Journal of Physical Chemistry A 118, $974-985$.
575	Henry K. M., Donahue N. M. (2012). Photochemical aging of a-pinene secondary
576	organic aerosol: Effects of OH radical sources and photolysis. Journal of
577	Physical Chemistry A 116, 5932-5940.
578	IUPAC. (2013). "IUPAC Task Group on Atmospheric Chemical Kinetic Data
579	Evaluation, (http://iupac.pole-ether.fr)."
580	Madronich S. (2006). Chemical evolution of gaseous air pollutants down-wind of
581	tropical megacities: Mexico City case study. Atmospheric Environment 40,
582	6012-6018.
583	Magneron I., Bossoutrot V., Mellouki A., Laverdet G., Le Bras G. (2003). The OH-
584	initiated oxidation of hexylene glycol and diacetone alcohol. Environmental
585	Science and Technology 37, 4170-4181.

586	Matsunaga S. N., Kawamura K. (2000). Determination of a- and b-hydroxycarbonyls
587	and dicarbonyls in snow and rain samples by GC-FID and GC-MS employing
588	benzyl hydroxyl oxime derivatization. Analytical Chemistry 72, 4742-4746.
589	Mellouki A., Le Bras G., Sidebottom H. (2003). Kinetics and Mechanisms of the
590	Oxidation of Oxygenated Organic Compounds in the Gas Phase. Chemical
591	Reviews 103, 5077-5096.
592	Messaadia L., El Dib G., Ferhati A., Roth E., Chakir A. (2012). Gas phase UV
593	absorption cross-sections for a series of hydroxycarbonyls. Chemical Physics
594	Letters 529, 16-22.
595	Messaadia L., El Dib G., Lendar M., Cazaunau M., Roth E., Ferhati A., Mellouki A.,
596	Chakir A. (2013). Gas-phase rate coefficients for the reaction of 3-hydroxy-2-
597	butanone and 4-hydroxy-2-butanone with OH and CI. Atmospheric
598	Environment 77, 951-958.
599	Moortgat G. K. (2001). Important photochemical processes in the atmosphere. Pure
600	Applied Chemistry 73, 487-490.
601	Orlando J. J., Tyndall G. S. (2012). Laboratory studies of organic peroxy radical
602	chemistry: an overview with emphasis on recent issues of atmospheric
603	significance. Chemical Society Reviews 41, 6294-6317.
604	Orlando J. J., Tyndall G. S., Fracheboud J. M., Estupinan E. G., Haberkorn S.,
605	Zimmer A. (1999). The rate and mechanism of the gas -phase oxidation of
606	hydroxyacetone. Atmospheric Environment 33, 1621-1629.
607	Reisen F., Aschmann S. M., Atkinson R., Arey J. (2003). Hydroxyaldehyde products
608	from hydroxyl radical reactions of z-3-hexen-1-ol and 2-methyl-3-buten-2-ol
609	quantified by SPME and API-MS. Environmental Science and Technology 37,
610	4664-4671.
611	Sander S. P., Abbatt J. P. D., Barker J. R., Burkholder J. B., Friedl R. R., Golden D.
612	M., Huie R. E., Kolb C. E., Kurylo M. J., Moortgat G. K., Orkin V. L., Wine P. H.
613	(2011). Chemical kinetics and photochemical data for use in atmospheric
614	studies, Evaluation No. 17. Pasadena, Jet Propulsion Laboratory,
615	http://jpldataeval.jpl.nasa.gov.
616	Singh H., Chen Y., Staudt A., Jacob D., Blake D., Heikes B., Snow J. (2001).
617	Evidence from the Pacific troposphere for large global sources of oxygenated
618	organic compounds. Nature 410, 1078-1081.
619	Sleiman C., El Dib G., Tabet A., Canosa A. (2013). Atmospheric degradation of 4-
620	hydroxy-4-methyl-2-pentanone with OH in the gas phase at 297 K. Energy
621	Procedia 36, 502-510.
622	Spaulding R., Charles M. (2002). Comparison of methods for extraction, storage, and
623	silylation of pentafluorobenzyl derivatives of carbonyl compounds and multi-
624	functional carbonyl compounds. Analytical and Bioanalytical Chemistry 372,
625	808-816.
626	Szabo E., Djehiche M., Riva M., Fittschen C., Coddeville P., Sarzynski D., Tomas A.,
627	Dobé S. (2011). Atmospheric chemistry of 2,3-pentanedione: Photolysis and
628	reaction with OH radicals. Journal of Physical Chemistry A 115, 9160–9168.
629	Szabo E., Tarmoul J., Tomas A., Fittschen C., Dobe S., Coddeville P. (2009).
630	Kinetics of the OH-radical initiated reactions of acetic acid and its deuterated
631	isotopes. Reaction Kinetics and Catalysis Letters 96, 299-309.
632	Téton S., Mellouki A., Le Bras G., Sidebottom H. (1996). Rate constants for reactions
633	of OH radicals with a series of asymmetrical ethers and tert-butyl alcohol.
634	International Journal of Chemical Kinetics 28, 291-297.

Tuazon E. C., Aschmann S. M., Arey J., Atkinson R. (1998). Products of the gas phase reactions of a series of methyl-substituted ethenes with the OH radical. Environmental Science and Technology 32, 2106-2112.

- Turpin E., Tomas A., Fittschen C., Devolder P., Galloo J.-C. (2006). Acetone-h₆ or -d₆
 + OH reaction products: Evidence for heterogeneous formation of acetic acid
 in a simulation chamber. Environmental Science and Technology 40, 59565961.
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- 646
- 647

Tables and Figures

Table 1: Experimental conditions used for the reaction of 3H3M2B with OH

radicals in the Pyrex simulation chamber coupled to a FTIR spectrometer.

	3H3M2B + OH \rightarrow products
Femperature (K)	298-356
Pressure (Torr)	500-600
Reference compound	Tert-butyl alcohol
Optical path (m)	24-56
[3H3M2B] ₀ (ppm)	10-50
[Reference]₀ (ppm)	10-50
Spectral range (cm ⁻¹): 3H3M2B	990-950 and 3530-3480
Spectral range (cm ⁻¹): reference	1060-1000 and 3665-3620

Table 2: Photolysis frequencies, global quantum yields and product yields (corrected for wall loss and photolysis) of the photolysis of 3H3M2B in the absence and presence of OH-scavenger. Errors represent the global uncertainty (2 σ).

	Experiments	
	Without OH-scavenger	With OH-scavenger
[3M3H2B] ₀ (ppm)	18 - 60	9 - 52
J _{3H3M2B} (h ⁻¹)	0.33 ± 0.02	0.24 ± 0.02
Ф _{3Н3M2B} (270-390 nm)		0.79 ± 0.30
Molar yields of primary products (%)		
Acetone	86 ± 13	98 ± 14
Formaldehyde	16 ± 4	23 ± 4
Methanol	6 ± 1 -	11 ± 3 ^a 6 ± 1 ^b
2,3-butanedione	1.7 ± 0.5	0.4 ± 0.14
Methylglyoxal ^c	0.11 ± 0.05	0
Hydroxyacetone	NQ ^d	NQ ^d
: OH seavenger: m xylene		

^a: OH-scavenger: m-xylene ^b: OH-scavenger: carbon monoxide

^c: yield without photolysis correction

d: NQ: not quantified

Temperature (K)	k _{3H3M2B} /k _{ref}	k _{знзм2в} (/ 10 ⁻¹³ сm ³ molecule ⁻¹ s ⁻¹)
298	0.67 ± 0.01	7.3 ± 1.0
298	0.69 ± 0.01	7.4 ± 1.0
298	0.75 ± 0.02	8.1 ± 1.0
		7.6 ± 0.5*
314	0.81 ± 0.01	9.2 ± 2.0
315	0.78 ± 0.01	8.9 ± 2.0
		9.0 ± 0.3*
334	0.81 ± 0.01	9.6 ± 1.5
334	0.75 ± 0.02	8.9 ± 1.5
335	0.76 ± 0.02	9.0 ± 1.5
		9.2 ± 0.4*
353	0.97 ± 0.01	12.0 ± 2.0
356	0.73 ± 0.02	9.2 ± 2.0
		10.6 ± 2.0*

Table 3: Average rate constants for the reactions of 3H3M2B with OH 669

670

* uncertainties correspond to the standard deviation of the rate constant

671

Figure 1: Kinetic plots of 3H3M2B photolysis. Empty symbols are for
experiments without OH-scavenger (3 experiments), while full symbols are for
experiments with excess m-xylene as OH-scavenger (2 experiments).

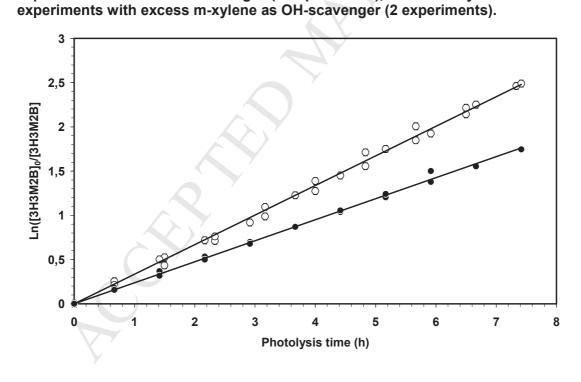
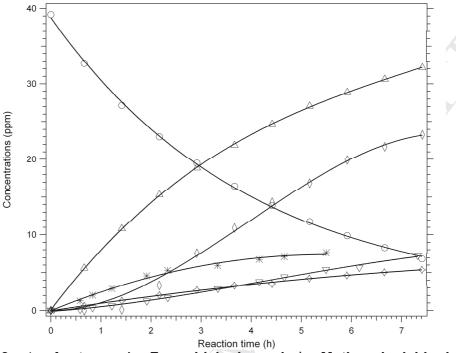


Figure 2: Time profiles of 3H3M2B and reaction products (experiment with mxylene as OH-scavenger). Symbols: O: 3H3M2B; \triangle : Acetone; \diamond : Acetic acid; *: Formaldehyde; ∇ : Carbon monoxide; \diamond : Methanol. Lines correspond to polynomial fits (except an exponential fit for 3H3M2B).



682

Figure 3: \triangle : Acetone, *****: Formaldehyde and \diamond : Methanol yield plots for experiments performed in the presence of m-xylene as OH-scavenger. Lines correspond to linear fits for acetone and formaldehyde and to a polynomial fit (second order) for methanol.

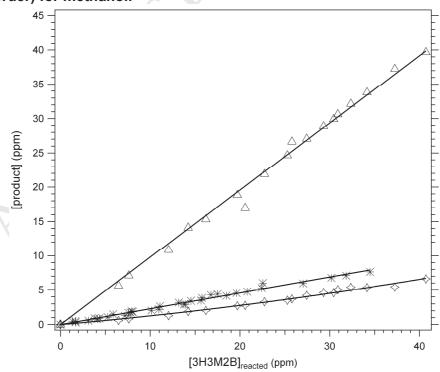
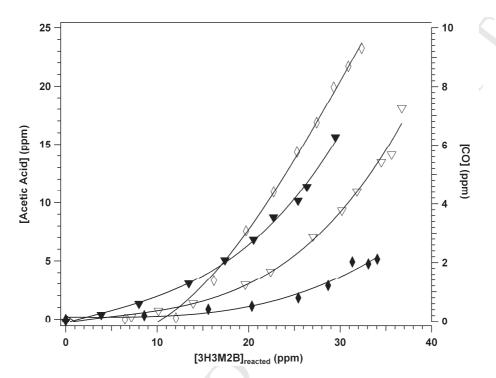
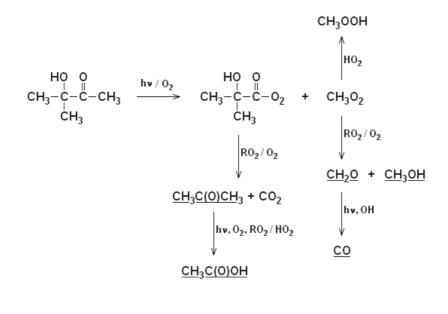
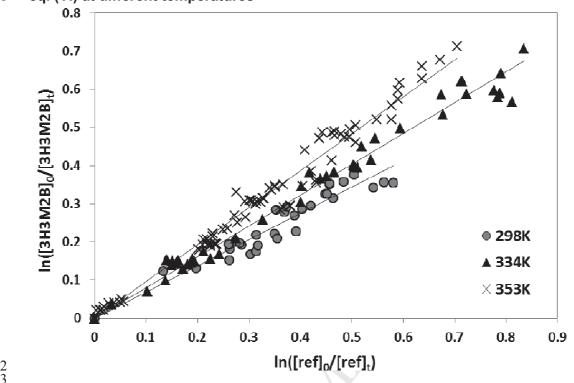


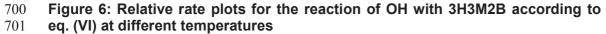
Figure 4: Plots of acetic acid (diamonds - left scale) and carbon monoxide
 (triangles - right scale) formation against 3H3M2B reacted. Full symbols are for
 experiments without OH-scavenger, while empty symbols are for experiments
 with OH-scavenger. Full lines correspond to polynomial fits of the data to help
 visualizing the data points.







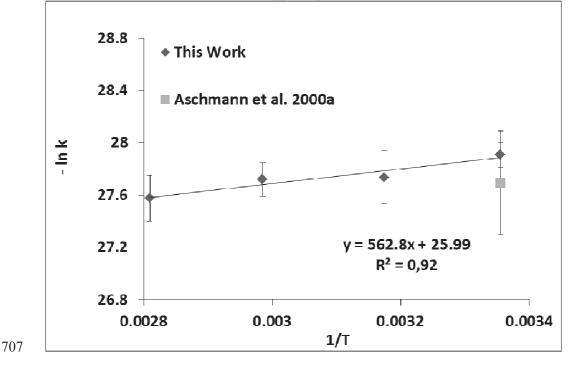












Reactivity of 3-hydroxy-3-methyl-2-butanone: Photolysis and OH reaction kinetics

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Highlights

- 3H3M2B photolysis is the major oxidation channel in the atmosphere
- Slight positive temperature dependence in the OH + 3H3M2B reaction kinetics
- Need for a better understanding of the atmospheric chemistry of oxygenated VOC