



HAL
open science

Biochemical, Transcriptional and Translational Evidences of the Phenol-meta-Degradation Pathway by the Hyperthermophilic *Sulfolobus solfataricus* 98/2

Alexia Comte, Pierre Christen, Sylvain Davidson, Matthieu Pophillat, Jean Lorquin, Richard Auria, Gwenola Simon, Laurie Casalot

► **To cite this version:**

Alexia Comte, Pierre Christen, Sylvain Davidson, Matthieu Pophillat, Jean Lorquin, et al.. Biochemical, Transcriptional and Translational Evidences of the Phenol-meta-Degradation Pathway by the Hyperthermophilic *Sulfolobus solfataricus* 98/2. PLoS ONE, 2013, 8 (12), pp.e82397. 10.1371/journal.pone.0082397 . hal-00999455

HAL Id: hal-00999455

<https://hal.science/hal-00999455>

Submitted on 3 Jun 2014

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

Biochemical, Transcriptional and Translational Evidences of the Phenol-*meta*-Degradation Pathway by the Hyperthermophilic *Sulfolobus solfataricus* 98/2

Alexia Comte^{1,2,9}, Pierre Christen^{1,2,*9}, Sylvain Davidson^{1,2}, Matthieu Pophillat³, Jean Lorquin^{1,2}, Richard Auria^{1,2}, Gwenola Simon^{1,2}, Laurence Casalot^{1,2}

1 Aix Marseille Université, CNRS/INSU, IRD, Mediterranean Institute of Oceanography (MIO), UM 110, Marseille, France, **2** Université du Sud Toulon-Var, CNRS/INSU, IRD, Mediterranean Institute of Oceanography (MIO), UM 110, La Garde, France, **3** Institut Paoli-Calmettes, Aix Marseille Université, CNRS, UMR7258, Centre de Recherche en Cancérologie de Marseille (CRCM), Marseille, France

Abstract

Phenol is a widespread pollutant and a model molecule to study the biodegradation of monoaromatic compounds. After a first oxidation step leading to catechol in mesophilic and thermophilic microorganisms, two main routes have been identified depending on the cleavage of the aromatic ring: *ortho* involving a catechol 1,2 dioxygenase (C12D) and *meta* involving a catechol 2,3 dioxygenase (C23D). Our work aimed at elucidating the phenol-degradation pathway in the hyperthermophilic archaea *Sulfolobus solfataricus* 98/2. For this purpose, the strain was cultivated in a fermentor under different substrate and oxygenation conditions. Indeed, reducing dissolved-oxygen concentration allowed slowing down phenol catabolism (specific growth and phenol-consumption rates dropped 55% and 39%, respectively) and thus, evidencing intermediate accumulations in the broth. HPLC/Diode Array Detector and LC-MS analyses on culture samples at low dissolved-oxygen concentration (DOC = 0.06 mg.L⁻¹) suggested, apart for catechol, the presence of 2-hydroxymuconic acid, 4-oxalocrotonate and 4-hydroxy-2-oxovalerate, three intermediates of the *meta* route. RT-PCR analysis on oxygenase-coding genes of *S. solfataricus* 98/2 showed that the gene coding for the C23D was expressed only on phenol. In 2D-DIGE/MALDI-TOF analysis, the C23D was found and identified only on phenol. This set of results allowed us concluding that *S. solfataricus* 98/2 degrade phenol through the *meta* route.

Citation: Comte A, Christen P, Davidson S, Pophillat M, Lorquin J, et al. (2013) Biochemical, Transcriptional and Translational Evidences of the Phenol-*meta*-Degradation Pathway by the Hyperthermophilic *Sulfolobus solfataricus* 98/2. PLoS ONE 8(12): e82397. doi:10.1371/journal.pone.0082397

Editor: Arnold Driessen, University of Groningen, The Netherlands

Received: June 19, 2013; **Accepted:** October 23, 2013; **Published:** December 11, 2013

Copyright: © 2013 Comte et al. This is an open-access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

Funding: Alexia Comte was the recipient of a CIFRE fellowship (Kemesys/ANRT). The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Competing Interests: The authors have declared that no competing interests exist.

* E-mail: pierre.christen@univ-amu.fr

⁹ These authors contributed equally to this work.

Introduction

Phenol is an aromatic-ring-containing compound largely used in organic chemical industry mostly for the production of bisphenol A, or as intermediate in resins, fibers, paints or pharmaceutical syntheses [1]. It also has applications in the perfumery, molecular-biology or medicine fields. Besides, it is considered as relatively dangerous for health, classified in the European Union as a mutagenic agent, with a threshold-value exposure for man set at 5 ppm in France. It is all the more harmful because of its high volatility and/or diffusivity providing a rapid propagation in the environment. Some authors have reviewed technologies used for phenol removal from wastewaters and gaseous streams [1]. A great number of microorganisms are able to degrade phenol and use it as sole energy or carbon source [2,3]. The aerobic biodegradation has received an increasing interest in the last few decades. Phenol has been chosen as a model molecule to study the aromatic-ring fission [4]. For monoaromatic compounds, such as phenol, benzoate, salicylate, benzene, etc., the first key degradation step involves its oxidation to catechol by a monooxygenase [5,6]. Then, catechol is degraded via two

alternative pathways, depending on the microorganism [4]. In the *ortho* route, aromatic ring is cleaved between the hydroxyl groups by the catechol 1,2 dioxygenase (C12D), leading to the *cis,cis* muconic acid. In the *meta* route, the ring cleavage occurs next to the two hydroxyl groups (Figure S1). It is catalyzed by the catechol 2,3 dioxygenase (C23D), and leads to the 2-hydroxymuconic semialdehyde (2-HMS) [7,8]. Then, the 2-HMS can be degraded either through the hydrolytic route or the 4-oxalocrotonate (4-OC) route [9,10,11]. Both *ortho* and *meta* routes can be active for the same microorganism depending on the substrate. For example, only the *ortho* route was involved on salicylate while both routes were activated on benzoate in *Pseudomonas cepacia* [12]. It can also depend on the aromatic concentration as demonstrated, in *Pseudomonas putida* grown on benzoate. At low concentrations (< 200–300 mg.L⁻¹), only the *ortho* pathway is involved, while both degradation routes are activated at higher concentrations [10,13]. In the latter case, a proteomic study showed the simultaneous presence of 8 catabolic enzymes: 3 corresponding to the *ortho*-cleavage route and 5 involved in the *meta*-cleavage one [10].

If most phenol biodegradation studies concerned mesophilic microorganisms, some involved thermophilic bacteria from the *Bacillus* genus [14,15] or hyperthermophilic archaea [16–18]. In the past two decades, some reviews have dealt with the promising future of thermophile and hyperthermophile enzymes for industrial applications: thermo-stable amylases, xylanases, proteases or DNA polymerases for potential use in food, chemical or pharmaceutical industries [19–21]. Besides their stability at high temperature, these enzymes are also known to withstand denaturant or acidic/alkaline conditions. Moreover, they are highly specific, robust and can be produced through either fermentation by the thermophilic microorganism or by cloning in fast-growing mesophiles by DNA recombinant technology [22]. Most of the thermophilic microorganisms belong to the Archaea group, grow at low pH and usually live in extreme environments such as solfataric fields or submarine hydrothermal areas [23]. Their physiological characteristics and the general features of their genome sequences have been reviewed elsewhere [24].

Oxygen availability is an important parameter in phenol aerobic biodegradation [25–27] especially at high temperature (80°C), for which its solubility in water is only 3 mg.L⁻¹. Viggor et al. [6] showed that oxygen was a co-substrate of the monooxygenase, responsible for the oxidation of phenol to catechol. In thermophilic *Bacilli*, it has been shown that maximal degradation was reached at an O₂ delivery of 1 vvm, while inactivation of the C23D was observed at high O₂ levels [28]. This limiting effect at low O₂ levels was also observed with *Pseudomonas* CF600 together with the accumulation in the medium of 2-HMS, an intermediate of the *meta* pathway [27].

Amongst high-temperature tolerant *Archaea*, our interest has focused on *Sulfolobus solfataricus* 98/2, which genome has been sequenced [29]. Its ability to grow on phenol was recently discussed [17,30], and kinetic parameters of phenol biodegradation were established [18]. Moreover, a C23D gene was identified in the 98/2 strain [31].

In this work, experiments were designed to study phenol-degradation route in *S. solfataricus* 98/2 cultivated in a fed-batch bioreactor. Growth, substrate and oxygen consumptions as well as product accumulated in the broth and CO₂ production are monitored. Transcriptomic and proteomic studies are also carried out in order to determine the metabolic pathway used for phenol degradation.

Materials and Methods

Strain and medium

S. solfataricus 98/2 was used in this study [32]. Cells were maintained at -80°C and reactivated on the mineral medium reported elsewhere [17]. The strain was previously adapted to phenol by repeated batches at concentrations up to 400 mg.L⁻¹, shown to be well tolerated by the strain [17,18].

Experimental set up

S. solfataricus was batch-cultivated in a 2.7 L reactor described in a previous paper [17]. Working volume was 1.8 L. The fermentor was equipped with pH, redox, dissolved-oxygen and temperature probes connected to an automat (Wago, France). The automat was connected to a computer for process monitoring and data capture (BatchPro Software, Decobecq Automatismes, France).

Experimental conditions

S. solfataricus was cultivated at 80°C by repeated additions of phenol at concentrations up to 400 mg.L⁻¹, this level being shown to be under the inhibition threshold [18]. Flask cultures (500 mL)

of phenol-adapted *S. solfataricus* 98/2 were used to inoculate the fermentor filled with standard mineral medium with phenol (initial optical density (OD) in a range of 0.15–0.20). Oxygen was first fed to maintain a dissolved-O₂ concentration (DOC) of 1.5 mg.L⁻¹. When biomass reached about 0.35 g.L⁻¹, the oxygen set up was decreased to a DOC of 0.06 mg.L⁻¹. In all experiments, stirring was adjusted to 300 rpm and total aeration-flow rate to 100 mL.min⁻¹. Under these conditions, the K_La value for oxygen was 82.8 h⁻¹ [17]. DOC was regulated through the O₂/N₂ ratio in the inlet gas. The pH was maintained at 3.2 with NaOH 0.5 mM. The exit gas was efficiently cooled to avoid phenol loss by evaporation as described elsewhere [17]. For proteomic and transcriptomic studies, experiments were also carried out on glucose (1.8 g.L⁻¹).

Analytical methods

Cell density was determined by OD measurement at 600 nm. Cell dry weight was calculated from OD data by using the relation of 1 OD unit = 320 mg.L⁻¹ [17].

Phenol consumption and intermediate-metabolite production were followed, after centrifugation of the sample (5 min, 14000 rpm), by HPLC equipped with a Diode Array Detector (DAD), as previously described [17]. Intermediary-metabolite concentrations are expressed as mg_{eq.phenol}.L⁻¹. Chemical structures of intermediates were explored by Liquid Chromatography Mass Spectrometry (LC-MS) using a Hitachi Elite LaChrom L-2130 liquid chromatograph coupled to a Bruker Esquire 6000 MS. The MS detector is equipped with an electro spray ionization in positive and negative mode and a quadrupole analyzer. Separation was achieved with a Varian Polaris C18 column eluted by a gradient of acetonitrile in water containing acetic acid (0.1% v/v), from 0 to 40% acetonitrile during 20 min, at a flow rate of 200 µL.min⁻¹. The cone and capillary voltages were maintained at 30 and 3500 V, respectively. To determine those chemical structures, 10 mL of culture were centrifuged; the supernatant extracted with ethyl acetate and analyzed by GC-MS as already described [33].

Both biomass and phenolic-compound analyses were made by triplicate and the average value was reported. Specific growth rate (µ, h⁻¹) and specific degradation rate (q_p, mg.g⁻¹.h⁻¹) were calculated from biomass and phenol concentration data.

Oxygen consumption was calculated through the O₂ mass flow needed to maintain the DOC set point in the broth. Carbon dioxide production was measured online in the exit gas by infrared analyzer. The oxygen yield factor (Y_{X/O₂}, g.g⁻¹), the respiratory quotient (Q_{resp}, mol.mol⁻¹), the biomass yield on phenol (Y_{X/P}, g.g⁻¹) and the carbon balance were calculated as previously described [17].

RNA extraction

Cells of *S. solfataricus* (50 mL), grown in different conditions, were harvested by centrifugation 10 min at 7400 rpm at 4°C. Pellets were snap-frozen in liquid N₂. RNA was extracted from the cell pellets using the High Pure RNA Isolation Kit (Roche Applied Science, USA). The purified nucleic acids were then treated with Turbo DNase (Ambion, USA) and RNA were purified again with the same kit. The quantity and quality of the obtained RNA were evaluated spectrophotometrically on a BioSpec-nano spectrophotometer (Shimadzu, Kyoto, Japan) and samples were diluted to 20 ng. µL⁻¹ before conservation at -80°C.

Semi-quantitative reverse-transcription PCR (RT-PCR) analyses

For RT-PCR analysis, total RNA (0.2 µg) was used to synthesize cDNA by triplicate in 20-µL reactions using the SuperScript® III Reverse Transcriptase and random primers (Invitrogen, USA). PCR was performed for 15, 20, 25 or 30 cycles using a step-cycle program of 98°C for 30 s, 50°C for 30 s, and 72°C for 20 s. Primers were designed using the Primer3 software. Primer pairs are listed Table 1. The amplification products were separated on a 1% agarose gel by electrophoresis, and the gel images were acquired using a Gel Doc Imager (Biorad, USA). DNA controls were carried out to exclude any DNA contamination.

2D-DIGE

The cells were harvested by centrifugation at 7000 g at 4°C for 15 minutes and washed with ice-cold phosphate buffered saline. The cells were broken by sonication in urea lysis buffer (8 M urea, 2 M thiourea, 4% (w/v) 3-[[3-Cholamidopropyl] dimethyl ammonio]-1-propane sulfonate (CHAPS), pH 8.5). The samples were then clarified by ultracentrifugation at 35000 g at 4°C for 1 h. Soluble proteins were purified and concentrated by precipitation with 4 volumes of ice-cold acetone, and solubilized for 1 h in 100 µL urea lysis buffer. The protein concentration was estimated using Bradford assay (Biorad, Hercules, CA, USA) according to the manufacturer's instructions. All samples were then washed with the 2D-Clean-up kit (GE Healthcare, USA) and solubilized in urea lysis buffer to a final concentration of 2.5 µg. µL⁻¹. The soluble *S. solfataricus*-protein fractions were labeled with cyanine dyes: Cy3, Cy5, Cy2 (CyDyes, GE Healthcare, USA). 25 µg of each protein extract was labeled separately at 4°C in the dark for 30 min with 200 pmol. µL⁻¹ of the N-hydroxysuccinimide esters of cyanine dyes (Cy3 or Cy5). The internal standard, corresponding to a pool of the samples (12.5 µg of each individual extract), was prepared in parallel and labeled with Cy2. Total protein labeled with Cy2, Cy3 and Cy5 were combined and mixed with an equal volume of 2x urea lysis buffer containing 1% carrier ampholytes pH 4–7 according to the manufacturer's instructions. For the first dimension (IsoElectric Focusing, IEF), precasted IPG (Immobilized pH Gradient) strips were used (pH 4–7, non linear (NL), 11 cm length; Immobiline DryStrips, GE Healthcare, USA). Typically, 75 µg of protein (25 µg for each dye) was loaded onto each IPG strip and the IEF was carried out (IPGPhor III, GE Healthcare, USA). The IEF protocol was as follows: 0–300 V gradient for 1h; 300–1000V gradient for 1.5 h; 1000–6000 V

gradient for 2 h; 6000 V for 2 h. Temperature was set up at 20°C. Prior to SDS PAGE, IPG strips were equilibrated during 20 min in an equilibration buffer (6 M urea, 50 mM Tris pH 8.8, 2% SDS, 38.5% glycerol) added with 65 mM DTT for the first 10 min and 2% iodoacetamide for the further 10 min. The second dimension was performed using a Criterion Dodeca Cell separation unit (Biorad, Hercules, CA, USA) and precast 10% SDS-PAGE gels (Biorad, Hercules, CA, USA). IPG strips were placed on the top of the precast gels, overlaid with 0.5% agarose in 2x running buffer containing bromophenol blue. Gels were run at 20°C using the following XT-MES running buffer (Biorad, Hercules, CA, USA): 1X for the cathode and 2X running buffer for the anode part. Electrophoresis was conducted overnight at 15 V and stopped when the bromophenol-blue-dye front has reached the bottom of the gel. After SDS PAGE, cyanine-dye-labeled-protein gels were scanned directly using the Typhoon FLA9000 scanner (GE Healthcare, USA). All gels were scanned with a resolution of 50 µm. Determination of protein abundance and statistics based on 2D DIGE were carried out with the Decyder Software (version 6.5, GE Healthcare, USA). First step for the spot detection is the creation of crop images of the region of interest. Cropped images were imported onto Decyder. Spot detection was set as 10000 with a filter volume set at 30000. Spot selection was performed for a ratio up to 2 and a t-test p-value < 1%.

Gel digestion and MALDI-TOF MS

MALDI-TOF MS is based on the Decyder analysis. Spots of interest were excised using Shimadzu Biotech Xcise System (Champs sur Marne, France). The proteins were subjected to in-gel digestion with trypsin, (Sequencing-grade modified porcine trypsin; Promega, Madison, WI, USA). Tryptic peptides were then extracted from the gel by successive treatment with 5% formic acid and 60% acetonitrile/5% formic acid. Each treatment is followed by sonication (5 min). Extracts were pooled and dried in a Speedvac evaporator. Peptides, resuspended in an α-cyano-4-hydroxycinnamic-acid-matrix solution (prepared by diluting 6 times a saturated solution in 50% acetonitrile/0.3% trifluoroacetic acid), were then spotted on the metal target. Mass analyses were performed on a MALDI-TOF Bruker Ultraflex III spectrometer (Bruker Daltonics, Wissembourg, France) controlled by the Flexcontrol 2.0 package (Build 51). This instrument was used at a maximum accelerating potential of 25 kV and was operated in reflector mode and the m/z ranges from 600 to 3500. Six external standards (Peptide Calibration Standard II, Bruker Daltonics, Wissembourg, France) were used to calibrate each spectrum to a

Table 1. Primers used in gene-expression analyses in *S. solfataricus* 98/2.

Genes	Proteins	Primers
<i>ssol_0230</i>	4-hydroxyphenyl pyruvate dioxygenase	5'-CACTGTGGCCAAGTTTCTGA-3' 5'-CCATAACGCTTTTGGGATGT-3'
<i>ssol_1707</i>	Gentisate 1,2 dioxygenase	5'-AGGGGACTAACGCCTACGAT-3' 5'-AAACATCACCTGCCTCCAC-3'
<i>ssol_2369</i>	Homogentisate 1,2 dioxygenase	5'-TACGCATCCTTTTGACGTG-3' 5'-GGAACAGACTGGGGTGATA-3'
<i>ssol_2712</i>	Extradiol cleavage dioxygenase	5'-CGGTTTCCCAGAAGAGACCT-3' 5'-GGGCTATTGTCGGTAATGGA-3'
<i>ssol_2912</i>	Catechol 2,3 dioxygenase	5'-TGCGCTAATTTCTGTCTGA-3' 5'-ATTGGGAGCCAATAGTGTGG-3'

doi:10.1371/journal.pone.0082397.t001

mass accuracy within 10 ppm with a minimal resolution of 10000 the angiotensin II peak (Monoisotopic mass = 1046.542 Da). Peak picking was performed with Flexanalysis 2.0 software (Bruker Daltonics, Wissembourg, France) with an adapted-analysis method. Parameters used were as follows: SNAP peak detection algorithm, S/N threshold fixed to 6 and a quality factor threshold of 30. A list of contaminations was constituted from a blank sample (blank piece of gel treated and analyzed exactly as a true sample). The trypsin peaks of this blank sample were excluded of the database search.

Protein identification

Database searches for MS spectra were conducted using MASCOT software 2.2 (Matrix Science), available online, against the *S. solfataricus* P2-complete-proteome NCBI database for known proteins. Peptide-mass tolerance was set to 100 ppm. Criteria used for protein identification are given by Mascot as a Probability Based Mowse Score. Ion score is $-10 \cdot \log(P)$, where P is the probability that the observed match is a random event. Protein scores greater than X (X is a number between 55 and 74, from one search to the other) are significant ($p < 0.05$). The matching of theoretical pI and PM of the identified protein with those observed in the 2D-gel experiments was also used as another criterion of confidence.

Results

Growth kinetics and yields, degradation products and carbon balance

As suggested previously, modifying the operating conditions for culture, such as DOC, can be a way to identify the phenol degradation pathway through the apparition of intermediary metabolites [27]. When *S. solfataricus* was grown with a DOC of 1.5 mg.L^{-1} after two sequential phenol batches (phenol initial concentration $< 400 \text{ mg.L}^{-1}$), biomass reached a concentration of 0.35 g.L^{-1} within 4 days (data not shown). At the third phenol addition (350 mg.L^{-1}), the DOC was drastically decreased to 0.06 mg.L^{-1} . Kinetic and respirometric yields and rates are presented in Table 2. Although cells were repeatedly grown on phenol (no lag phase was observed at 1.5 mg.L^{-1}), at the lower O_2

level, a short adaptation period was needed for growth and phenol consumption (Figure 1A). Then, biological activity started again and was characterized by μ of 0.0106 h^{-1} and a q_p of $29.0 \text{ mg.g}^{-1}.\text{h}^{-1}$. These parameters were lower than those observed at 1.5 mg.L^{-1} ($\mu = 0.0235 \text{ h}^{-1}$ and $q_p = 47.5 \text{ mg.g}^{-1}.\text{h}^{-1}$, respectively (Table 2)). In the same way, when the O_2 concentration decreased, so did the $Y_{X/P}$ (from 0.539 to 0.426 g.g^{-1}).

At DOC of 0.06 mg.L^{-1} , carbon dioxide production and redox potential profiles were similar to those observed at 1.5 mg.L^{-1} , i.e. O_2 consumption and CO_2 production increased and redox potential decreased with growth (data not shown). This has already been described [17]. However, the yield coefficient for oxygen (Y_{X/O_2}) was 0.415 g.g^{-1} for a DOC of 1.5 mg.L^{-1} and decreased significantly to 0.265 g.g^{-1} for a DOC of 0.06 mg.L^{-1} . The respiratory quotient (Q_{resp}) was closed to the theoretical value of $0.87 \text{ mol.mol}^{-1}$ and was not affected by the oxygen level (Table 2).

At 0.06 mg.L^{-1} , during phenol consumption, catechol – identified with a standard solution by its absorption spectrum ($\lambda_{\text{max}} = 275 \text{ nm}$) and its retention time (7.20 min) in HPLC/DAD analysis – appeared in the broth after 20 h (Figure 1). Catechol reached a maximum concentration of $90 \text{ mg}_{\text{eq.phenol}}.\text{L}^{-1}$ at 32 h and then decreased (Figure 1A). Its dissimilation is correlated with the appearance of various, but minor, compounds, also evidenced by HPLC/DAD. Only three of the peaks corresponded to compounds with a defined λ_{max} . They are characterized by retention times of 2.55, 3.2 and 3.6 min, respectively, with the analysis conditions given in the Material and Methods section. They displayed a λ_{max} of 290, 254 and 265 nm, respectively (Figure 1B). Trace amounts of them appeared after 28 h (Figure 1A) and their concentration increased strongly when catechol concentration decreased (36 h). At the end of the run (40 h), they reached a maximum concentration of 6.4, 47.4 and $12.3 \text{ mg}_{\text{eq.phenol}}.\text{L}^{-1}$, respectively.

LC/MS analysis formally confirmed the presence of catechol in all the samples withdrawn after 20 h of run, but not the *cis,cis* muconic acid ($\lambda_{\text{max}} = 260$). In the same way, 4-OC, belonging to the *meta* pathway, was also identified.

Moreover, at the same time that the intermediate metabolites are detected (28 h), a yellow color appeared instantaneously in the samples in contact with air. Its intensity grew with time. In contrast, none of the samples at 1.5 mg.L^{-1} displayed this phenomenon.

At both DOC, carbon balances are closed to 100% (Table 2). However, at 1.5 mg.L^{-1} , carbon from phenol is exclusively directed toward CO_2 (63.1%) and biomass (34.3%), while at 0.06 mg.L^{-1} , carbon distribution is different. In this case, the CO_2 proportion is maintained relatively constant (60.2%) while the biomass part decreased (27.2%) and a significant intermediary metabolite amount was observed (15.7%).

Genomic analyses

Few studies concerned the enzymes involved in phenol degradation through the *meta* or *ortho* pathways in Sulfolobales. Chae et al. [31] have clearly identified the presence of a C23D-coding gene in *S. solfataricus* 98/2 through PCR amplification, sequencing and heterologous production in *E. coli* (Accession number EF494887). This enzyme showed the highest activity against catechol and 4-chlorocatechol. The corresponding *orf* was not identified in the genome annotation (Accession number NC_017274.1). Using a primer pair designed on the *S. solfataricus* P2 C23D genomic sequence (ss01223), we were able to amplify a fragment which sequence perfectly matched the one amplified by

Table 2. Kinetic and respirometric rates and yields of phenol degradation by phenol-grown *S. solfataricus* 98/2 cells at two dissolved-oxygen concentrations (DOC).

Parameters	DOC	
	1.5 mg.L^{-1}	0.06 mg.L^{-1}
μ (h^{-1})	0.0235 ($R^2 = 0.996$)	0.0106 ($R^2 = 0.995$)
q_p ($\text{mg.g}^{-1}.\text{h}^{-1}$)	47.5 (± 1.4)	29.0 (± 1.1)
$Y_{X/P}$ (g.g^{-1})	0.539 (± 0.023)	0.426 (± 0.057)
Q_{resp} (mol.mol^{-1})	0.828 (± 0.008)	0.805 (± 0.005)
Y_{X/O_2} (g.g^{-1})	0.415 (± 0.005)	0.265 (± 0.035)
Carbon balance (%)	97.4 (± 5.5)	103.1 (± 0.7)
- biomass* (%)	34.3 (± 1.5)	27.2 (± 3.7)
- carbon dioxide (%)	63.1 (± 4.0)	60.2 (± 2.9)
- metabolites (%)	0	15.7 (± 0.2)

*on the basis of a phenol-grown biomass empirical formulae of $\text{CH}_{1.8}\text{O}_{0.5}\text{N}_{0.2}$ and a molecular weight of 24.6 g.mol^{-1} [15].

doi:10.1371/journal.pone.0082397.t002

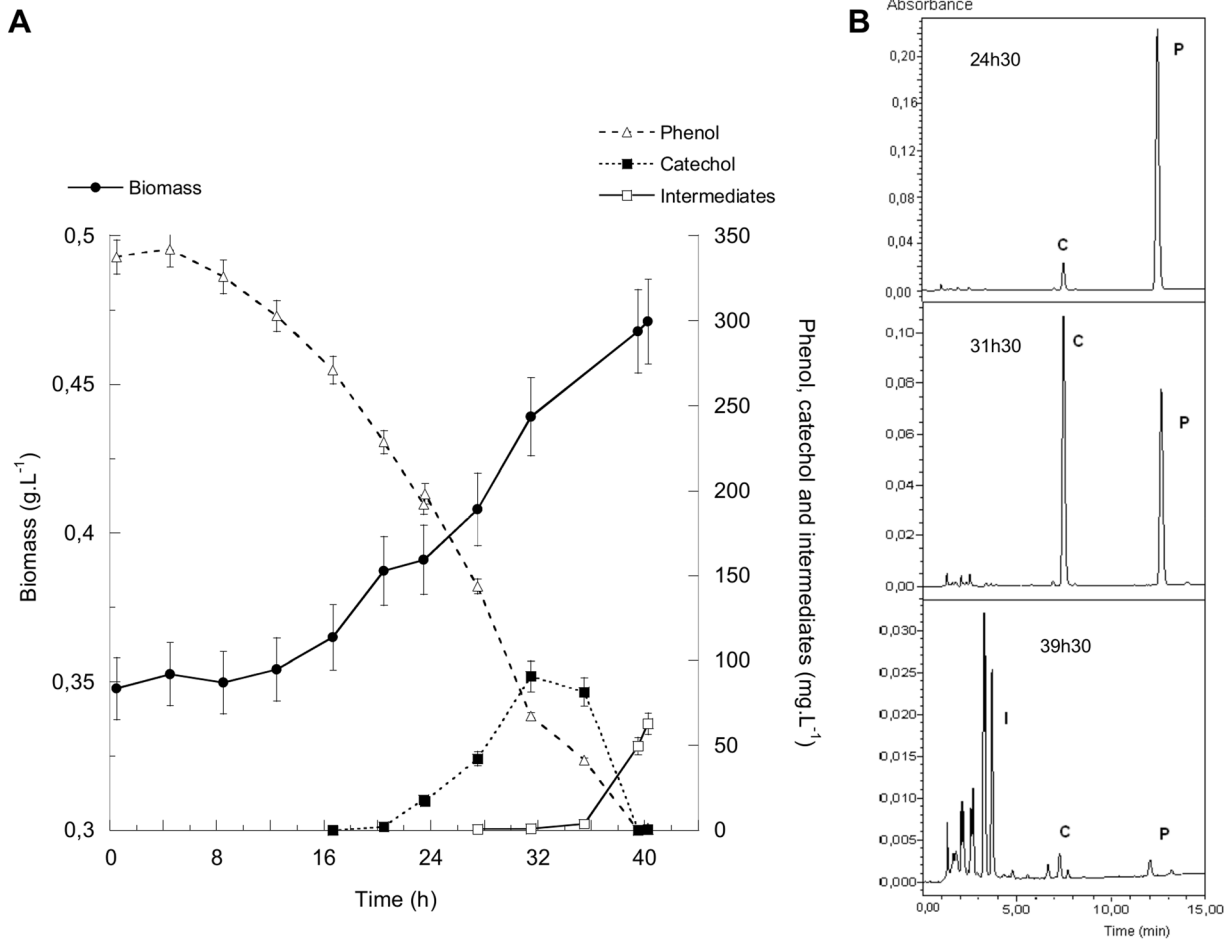


Figure 1. Analysis of *S. solfataricus* 98/2 cultures grown on phenol at a DOC of 0.06 mg.L⁻¹. (A) biomass, catechol and other intermediate productions and phenol consumption. The intermediate curve corresponds to the sum of all the detected compounds. (B) HPLC chromatograms ($\lambda = 270$ nm) obtained at three different time course of the run (P: phenol, C: catechol, I: intermediates). doi:10.1371/journal.pone.0082397.g001

Chae et al. [31]. We decided to name this *orf*, according to the *S. solfataricus* 98/2-genome annotation, *ssol_2912* (KF701465). No other gene coding for protein potentially involved in the *meta* pathway were identified in the genome of *S. solfataricus* 98/2. However, 4 genes coding for putative C12D, involved in the *ortho* degradation pathway, were found in the genome: *ssol_0230* (4-hydroxyphenyl pyruvate dioxygenase), *ssol_1707* (gentisate 1,2 dioxygenase), *ssol_2369* (homogentisate 1,2 dioxygenase) and *ssol_2712* (extra diol ring clivage dioxygenase).

Transcriptional analysis

To investigate which nutrient conditions induce the transcription of these different oxygenases in *S. solfataricus*, the transcription levels of these genes were examined by semi-quantitative RT-PCR analysis. For this analysis, mRNA were isolated from samples of *S. solfataricus* grown in liquid media with glucose or phenol as the sole carbon source and different oxygen concentrations. RT-PCR controls, using primer pairs designed to amplify 16S rRNA gene, showed the same expression level in each culture condition (Figure 2A). Using the primer pair designed to amplify *ssol_0230*, *ssol_1707*, *ssol_2369*, *ssol_2712* and *ssol_2912* genes, a unique band of the expected size was obtained (Figure 2B-F). The mRNA transcription levels of *ssol_0230* were equivalent when the strain was harvested from the medium with glucose or phenol and with a

DOC of 1.5 mg.L⁻¹ (Figure 2B, lanes 2-3), but decreased when the strain was cultivated with phenol at a DOC of 0.06 mg.L⁻¹ (Figure 2B, lane 1). The mRNA transcription levels of *ssol_1707* were equivalent with low and high oxygen concentrations in presence of phenol (Figure 2C, lanes 1-2) and slightly lower when the strain was harvested from the medium with glucose as substrate (Figure 2C, lane 3). The mRNA transcription levels of *ssol_2369* were equivalent in the strain cultivated with glucose and phenol (Figure 2D, lanes 2-3). However, it increased when the strain was harvested from the medium with phenol at a DOC of 0.06 mg.L⁻¹ (Figure 2D, lane 1). The mRNA transcription level of *ssol_2712* was slightly lower when the strain was harvested from the medium with glucose compared to phenol (Figure 2E, lanes 2-3). Yet, the transcriptional level of the extra-diol-ring-clivage-dioxygenase gene was higher in the media with high oxygen concentration (Figure 2E, lanes 1-2). The mRNA transcription level of *ssol_2912* was undetectable when the strain was harvested from the medium with glucose as the sole carbon source (Figure 2F, lane 3). Nevertheless, it greatly increased when the strain was harvested from the media with phenol (Figure 2F, lanes 1-2). This differential expression of the C23D gene was detected in the medium whatever the DOC.

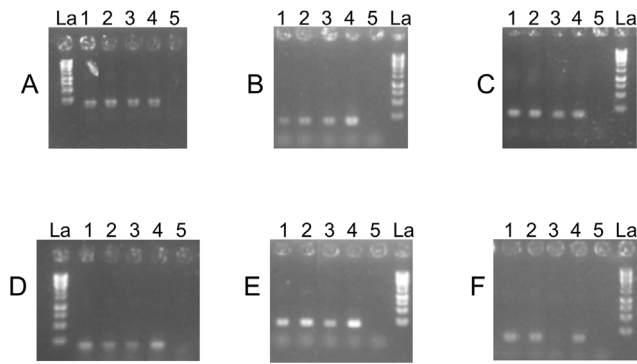


Figure 2. RT-PCR analyses. (A) 16S rRNA, (B) *sso_0230* (4-hydroxyphenyl pyruvate dioxygenase), (C) *sso_1707* (gentisate 1,2 dioxygenase), (D) *sso_2369* (homogentisate 1,2 dioxygenase), (E) *sso_2712* (extra diol ring cleavage dioxygenase), and (F) *sso_2912* (C23D) genes from *S. solfataricus*. Amplified fragments are 200-bp long. Lane 1: cells of *S. solfataricus* cultivated with phenol and a DOC of 0.06 mg.L^{-1} . Lane 2: cells of *S. solfataricus* cultivated with phenol and a DOC of 1.5 mg.L^{-1} . Lane 3: cells of *S. solfataricus* cultivated in glucose and a DOC of 1.5 mg.L^{-1} . Lane 4: positive control, PCR amplification on genomic DNA as template. Lane 5: negative control experiment performed omitting reverse transcriptase during RT-PCR reaction and showing complete absence of DNA in the RNA samples. La: 1 kb Ladder (Fermentas, France).

doi:10.1371/journal.pone.0082397.g002

Proteomic analysis

To investigate the influence of phenol on protein production in *S. solfataricus* 98/2, a 2D-DIGE experiment was performed. We compared the whole protein content of the cells grown on phenol or glucose. The representative 2D-DIGE gel is shown in Figure 3A. Interestingly, among the proteins more abundant on phenol only, spot 181 (Figure 3A) was identified as the C23D by MALDI-TOF MS. This enzyme is only detected on phenol (Figure 3B). Besides, no other dioxygenase could be identified as produced more abundantly on phenol. In particular, no differential

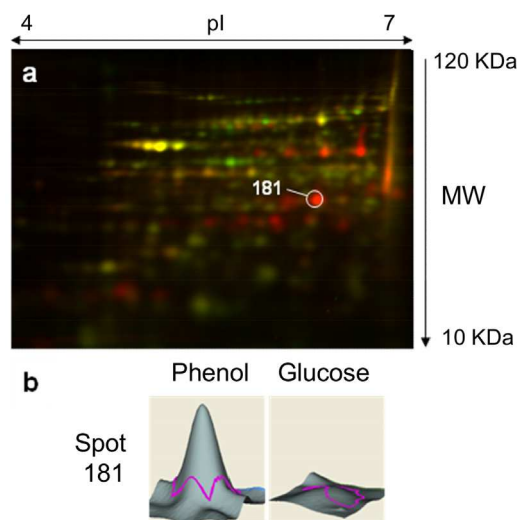


Figure 3. Differential proteome of *S. solfataricus* 98/2 grown on phenol or glucose. (A) 2D-DIGE gel. Cy2 (yellow): glucose + phenol conditions, Cy3 (green): glucose condition, and Cy5 (red): phenol condition. (B) Relative intensity of spot 181.

doi:10.1371/journal.pone.0082397.g003

production was observed for the putative C12D (*Sso_0230*, *Sso_1707*, *Sso_2369* and *Sso_2712*).

Discussion

Phenol and its derivatives are widely distributed environmental pollutants that are responsible for many unhealthy effects on humans. Degradation of such compounds has become of increasing interest for many years. As pointed out by Cao et al. [10], the most effective and economical way to remove them from the environment is the microbiological way. The catabolism of phenol involves the action of a monooxygenase responsible for the oxygenation of phenol to catechol. The latter is then subjected to either a *meta* or an *ortho* cleavage of the aromatic ring yielding 2-HMS or *cis-cis* muconic acid, respectively [2,4,8].

In this paper, we designed experiments to understand which pathway is involved in phenol degradation in *S. solfataricus* 98/2. One of the easiest ways to define the degradation pathway is to identify the reaction products. However, in *S. solfataricus* 98/2 cultivated in standard conditions, phenol is completely metabolized into CO_2 , H_2O and biomass [17]. In their paper, Kapley et al. [27] observed that 2-HMS accumulated at low DOC (2 mg.L^{-1}), probably because of a slower metabolism, which, in turn, allowed observing transitory accumulation of intermediary compounds. In our first experiment, performed at 1.5 mg.L^{-1} , the growth parameters were determined. Nevertheless, in this condition and in contrary to what was expected, we were unable to detect any intermediate for the phenol degradation. Our experimental set up enables to work at a regulated and constant DOC. The problem of reducing the DOC to a very low level is that, in this case, the oxygen might become the limiting factor for the growth. Previous experiments showed that, with our experimental set up, for DOC as low as 0.06 mg.L^{-1} , the carbon source still remains the limiting factor (data not shown). With such operating conditions, the behavior of the strain was notably modified. At this level, in comparison to the parameters measured at 1.5 mg.L^{-1} , growth (μ) and phenol-biodegradation rates (q_p) were slowed down (55% and 39%, respectively). These results can be compared to those reported by Ali et al. [28] who demonstrated that phenol specific degradation decreased with the oxygen concentration. Biomass yields on phenol ($Y_{X/P}$) and on oxygen (Y_{X/O_2}) at 1.5 mg.L^{-1} are in the range of those reported in the literature (Table 2). For example, for two different *P. putida* strains cultivated on phenol in a continuous fed stirred-tank reactor, Seker et al. [34] and Nikakhtari and Hill [25] reported values of $Y_{X/P}$ of 0.521 g.g^{-1} and 0.73 g.g^{-1} , respectively and of Y_{X/O_2} of 0.338 g.g^{-1} and 0.360 g.g^{-1} , respectively. Feitkenhauer et al. [35] reported, with a *Bacillus thermooleovorans* strain, an Y_{X/O_2} of 0.48 g.g^{-1} . At 0.06 mg.L^{-1} , we found that both parameters dropped 21% and 36%, respectively, indicating a less effective use of phenol or oxygen for biomass build up. Moreover, these coefficients ($Y_{X/S}$ and Y_{X/O_2}) have already been shown to be sensitive to O_2 levels for *S. solfataricus* as demonstrated by Simon et al. [36].

The slower growth of the strain (characterized by the 55% reduction of the growth rate), at a DOC of 0.06 mg.L^{-1} , gave suitable conditions for the accumulation of some of the intermediates (15.7%), as confirmed by the carbon balance (Table 2). After 20 h of culture, catechol, resulting from the oxidation of the aromatic ring, appeared in the culture. It accumulated up to 90 mg.L^{-1} and decreased (Figure 1). Its consumption is correlated to the appearance of, at least, three other compounds. The λ_{max} of two of these compounds are characteristic of two intermediates of the *meta*-degradation pathway: the 2-hydroxymuconic acid (2-

HMA, $\lambda_{\max} = 290$ nm) from the 4-OC route and the 4-hydroxy-2-oxovalerate (HOV, $\lambda_{\max} = 265$ nm) (Figure S1). Concomitantly, a yellow coloration is detected in the sample after contact with air. As soon as the early seventies, Buswell and Twoney [37] reported the presence of this color during the degradation of phenol and cresol by a *Bacillus stearotherophilus* strain. This color has been reported as typical of the *meta* pathway biodegradation of monoaromatic compounds and associated with the presence of 2-HMS in thermophilic *Bacilli* strains [14,38]. This has also been extensively reported in mesophilic strains (*Alcaligenes*, *Ralstonia*, *Pseudomonas*) [27,39,40]. The analysis of the last sample, using a more sensitive MS detector, strongly suggested the presence of 2-HMS. Moreover, the *cis,cis* muconic acid, characteristic of the *ortho* route, was not detected neither by UV_{260nm} nor by MS.

Dioxygenases are responsible for the opening of the ring. C12D open the ring between the C1 and C2 carbons of the catechol (*ortho* cleavage), while C23D open between C2 and C3 (*meta* cleavage). In the genome of *S. solfataricus* 98/2, four open reading frames putatively coding for 1,2 dioxygenases have been identified: *ssol_0230* (4-hydroxyphenyl pyruvate dioxygenase gene), *ssol_1707* (gentisate 1,2 dioxygenase gene), *ssol_2369* (homogentisate 1,2 dioxygenase gene), and *ssol_2712* (extra diol ring cleavage dioxygenase gene). The expression of these genes has been semi-quantitatively followed depending on the growth conditions. Even if the expression of some of these genes seems to be oxygen dependant (*ssol_0230* and *ssol_2369*), it is worth noting that all of the genes were expressed whether the carbon source was glucose or phenol. One open reading frame putatively coding for a C23D was identified in *S. solfataricus* 98/2 genome: *ssol_2912* (C23D gene). The semi-quantitative analysis of its expression showed that, in presence of glucose, the transcript was not detectable indicating a regulation of the expression depending on the carbon source. No difference observed in the expression of the gene coding for 16S rRNA validated the approach.

To confirm the presence or the absence of the protein depending on the growth condition, a comparative analysis of the proteome between cells harvested on glucose or on phenol was performed. The only detectable difference among the dioxy-

genases is on the production of the C23D. The four C12D were not identified in this experiment. However, their theoretical pI being within the tested range (*Ssol_1707*: 6.11, *Ssol_2369*: 6.24, *Ssol_0230*: 6.37 and *Ssol_2712*: 5.06, respectively), their absence means that the four proteins are equally produced on glucose and phenol.

In conclusion, our set of results (transcriptomic and proteomic) seems to indicate that both degradation pathways are functional in presence of phenol. However, the activation of the C23D, only when phenol is present, and the accumulation of only intermediary compounds related to this pathway lead us to the conclusion that the aromatic ring is preferentially opened through the *meta* pathway.

Supporting Information

Figure S1 Phenol degradative pathways. Dot arrow, *ortho* pathway. Dash arrow, *meta* pathway. MO, monooxygenase; C12D, catechol 1,2 dioxygenase; C23D, catechol 2,3 dioxygenase; 2-HMS H, 2-HMS hydrolase, 2-HMS DH, 2-HMS dehydrogenase; 4OT, 4-OC tautomerase; 4OD, 4-OC decarboxylase; OE H, OE hydratase; 2-HMS, 2-hydroxy muconic semialdehyde; 2-HMA, 2-hydroxy muconic acid; 4-OC, 4-oxalocrotonate; OE, 2-oxopent-4-dienoate; HOV, 4-hydroxy-2-oxovalerate; TCA: Tricarboxylic acid (adapted from Omokoko et al. [11]). (TIF)

Acknowledgments

The authors thank Corinne Valette for technical help. We also would like to thank Dr B. Omokoko and Dr. W. Hartmeier for allowing the use of their figure to inspire Figure S1.

Author Contributions

Conceived and designed the experiments: AC LC PC RA. Performed the experiments: AC GS LC MP PC SD. Analyzed the data: AC GS LC MP PC RA SD. Contributed reagents/materials/analysis tools: JL MP SD. Wrote the paper: AC GS JL LC MP PC RA SD.

References

- Busca G, Berardinelli S, Resini C, Arrighi L (2008) Technologies for the removal of phenol from fluid streams: a short review of recent developments. *J Hazard Mat* 160: 265–288.
- Agarry SE, Durojaiye AO, Solomon BO (2008) Microbial degradation of phenols: a review. *Int J Environ Poll* 32: 12–28.
- Sridevi V, Chandana Lakshmi MVV, Manasa M, Sravani M (2012) Metabolic pathways for the biodegradation of phenol. *Int J Eng Sci Adv Technol* 2: 695–705.
- van Schie PM, Young LY (2000) Biodegradation of phenol: Mechanisms and applications. *Bioremediation J* 4: 1–18.
- Tao Y, Fishman A, Bentley WE, Wood TK (2004) Oxidation of benzene to phenol, catechol, and 1,2,3-trihydroxybenzene by toluene 4-monooxygenase of *Pseudomonas mendocina* KR1 and toluene 3-monooxygenase of *Ralstonia pickettii* PKO1. *Appl Environ Microbiol* 70: 3814–3820.
- Viggor S, Heinaru E, Kümnapas A, Heinaru A (2008) Evaluation of different phenol hydroxylase-possessing phenol-degrading pseudomonads by kinetic parameters. *Biodegradation* 19: 759–769.
- Tsirogiani I, Aivaliotis M, Karas M, Tsiotis G (2004) Mass spectrometric mapping of the enzymes involved in the phenol degradation of an indigenous soil pseudomonad. *Biochim Biophys Acta* 1700: 117–123.
- Banerjee A, Ghoshal AK (2010) Phenol degradation by *Bacillus cereus*: Pathway and kinetic modelling. *Biores Technol* 101: 5501–5507.
- Powlowski J, Shingler V (1994) Genetic and biochemistry of phenol degradation by *Pseudomonas* sp. CF600. *Biodegradation* 5: 219–236.
- Cao B, Geng A, Loh KC (2008) Induction of *ortho*- and *meta*-cleavage pathways in biodegradation of high benzoate concentration: MS identification of catabolic enzymes. *Appl Microbiol Biotechnol* 81: 99–107.
- Omokoko B, Jäntges UK, Zimmermann M, Reiss M, Hartmeier W (2008) Isolation of the *phe*-operon from *G. stearotherophilus* comprising the phenol degradative *meta*-pathway genes and a novel transcriptional regulator. *BMC Microbiol* 8: 197.
- Hamzah RY, Al-Baharna BS (1994) Catechol ring-cleavage in *Pseudomonas cepacia*: the simultaneous induction of *ortho* and *meta* pathways. *Appl Microbiol Biotechnol* 41: 250–256.
- Loh KC, Chua SS (2002) *Ortho* pathway of benzoate degradation in *Pseudomonas putida*: induction of *meta* pathway at high substrate concentrations. *Enz Microbiol Technol* 30: 620–626.
- Mutzel A, Reinscheid UM, Antranikian G, Müller R (1996) Isolation and characterization of a thermophilic bacillus strain, that degrades phenol and cresols as sole carbon source at 70°C. *Appl Microbiol Biotechnol* 46: 593–596.
- Feitkenhauer H, Schnicke S, Müller R, Märkl H (2001) Determination of the kinetic parameters of the phenol-degrading thermophile *Bacillus thermoleovorans* sp. A2. *Appl Microbiol Biotechnol* 57: 744–750.
- Izzo V, Notomista E, Picardi A, Pennacchio F, Di Donato A (2005) The thermophilic archaeon *Sulfolobus solfataricus* is able to grow on phenol. *Res Microbiol* 156: 677–689.
- Christen P, Davidson S, Combet-Blanc Y, Auria R (2011) Phenol biodegradation by the thermoacidophilic archaeon *Sulfolobus solfataricus* 98/2 in a fed batch bioreactor. *Biodegradation* 22: 475–484.
- Christen P, Vega A, Casalat L, Simon G, Auria R (2012) Kinetics of aerobic phenol biodegradation by the acidophilic and hyperthermophilic archaeon *Sulfolobus solfataricus* 98/2. *Biochem Eng J* 62: 56–61.
- Niehaus F, Bertoldo C, Kähler M, Antranikian G (1999) Extremophiles as a source of novel enzymes for industrial application. *Appl Microbiol Biotechnol* 51: 711–719.
- Bouzas TD, Barros-Velazquez J, Villa TG (2006) Industrial applications of hyperthermophilic enzymes: A review. *Protein and Peptide Lett* 13: 645–651.
- Turner P, Mamo G, Karlsson N (2007) Potential and utilization of thermophiles and thermostable enzymes in biorefining. *Microbial Cell Factories* 6: 9.
- Haki GD, Rakshit SK (2003) Developments in industrially important thermostable enzymes: a review. *Biores Technol* 89: 17–34.

23. Stetter KO (2013) A brief history of the discovery of hyperthermophilic life. *Biochem Soc Trans* 41:416–420.
24. Bertoldo C, Dock C, Antranikian G (2004) Thermophilic microorganisms and their novel biocatalysts. *Eng Life Sci* 4: 521–532.
25. Nikakhtari H, Hill GA (2005) Modelling oxygen transfer and aerobic growth in shake flasks and well-mixed bioreactors. *Can J Chem Eng* 83: 493–499.
26. Contreras EM, Albertario ME, Bertola NC, Zaritzky NE (2008) Modelling phenol biodegradation by activated sludges evaluated through respirometric techniques. *J Hazard Mat* 158: 366–374.
27. Kapley A, Tolmare A, Purohit HJ (2001) Role of oxygen in the utilization of phenol by *Pseudomonas* CF600 in continuous culture. *World J Microbiol Biotechnol* 17: 801–804.
28. Ali S, Fernandez-Lafuente R, Cowan DA (1998) Meta-pathway degradation of phenolics by thermophilic *Bacilli*. *Enz Microbiol Technol* 23: 462–468.
29. Lucas S, Copeland A, Lapidus A, Glavina del Rio T, Tice H et al. (2009) *Sulfolobus solfataricus* 98/2, complete genome. Available: <http://www.ncbi.nlm.nih.gov/nuccore/261600703>.
30. Ulas T, Riemer SA, Zaparty M, Siebers B, Schomburg D (2012) Genome-scale reconstruction and analysis of the metabolic network in the hyperthermophilic archaeon *Sulfolobus solfataricus*. *PLoSone* 7:e43401.
31. Chae JC, Kim E, Bini E, Zylstra GJ (2007) Comparative analysis of the catechol 2,3-dioxygenase gene locus in thermoacidophilic archaeon *Sulfolobus solfataricus* strain 98/2. *Biochem Biophys Res Com* 357: 815–819.
32. Rolfmeier M, Blum P (1995) Purification and characterization of a maltase from the extremely thermophilic crenarchaeote *Sulfolobus solfataricus*. *J. Bacteriol* 177: 482–485.
33. Liebgott PP, Labat M, Casalat L, Amouric A, Lorquin J (2007) Bioconversion of tyrosol into hydroxytyrosol and 3,4-dihydroxyphenylacetic acid under hypersaline conditions by a new *Halomonas* sp. strain HTB24. *FEMS Microbiol Lett* 276: 26–33.
34. Seker S, Beyenal H, Salih B, Tanyolac A (1997) Multi-substrate growth kinetics of *Pseudomonas putida* for phenol removal. *Appl Microbiol Biotechnol* 47: 610–614.
35. Feitkenhauer H, Schnicke S, Müller R, Märkl H (2003) Kinetic parameters of continuous cultures of *Bacillus thermoleovorans* sp. A2 degrading phenol at 65°C. *J Biotechnol* 103: 129–135.
36. Simon G, Walther J, Zabeti N, Combet-Blanc Y, Auria R, et al. (2009) Effect of O₂ concentrations on *Sulfolobus solfataricus* P2. *FEMS Microbiol Lett* 299: 255–260.
37. Buswell JA, Twoney DG (1975) Utilization of phenol and cresols by *Bacillus stearotheophilus*, strain PH24. *J Gen Microbiol* 87: 377–379.
38. Adams D, Ribbons DW (1988) The metabolism of aromatic ring fission products by *Bacillus stearotheophilus* strain IC3. *J Gen Microbiol* 134: 3179–3185.
39. Leonard D, Lindley ND (1998) Carbon and energy flux constraints in continuous cultures of *Alcaligenes eutrophus* grown on phenol. *Microbiol* 144: 241–248.
40. Leonard D, Ben Youssef C, Destruhaut C, Lindley ND, Queinnec I (1999) Phenol degradation by *Ralstonia eutropha*: colorimetric determination of 2-hydroxymuconate semialdehyde accumulation to control feed strategy in fed-batch fermentations. *Biotechnol Bioeng* 65: 407–415.