

Assessing N emissions from organic residues spreading: Study of the on-field distribution patterns influence and of emission models sensitivity

A. Gueudet, Marilys Pradel, J.C. Roux, A. Lellahi, P. Vissac

▶ To cite this version:

A. Gueudet, Marilys Pradel, J.C. Roux, A. Lellahi, P. Vissac. Assessing N emissions from organic residues spreading: Study of the on-field distribution patterns influence and of emission models sensitivity. 15th International conference Ramiran, Jun 2013, Versailles, France. 4 p. hal-00947801

HAL Id: hal-00947801

https://hal.science/hal-00947801

Submitted on 17 Feb 2014

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

Assessing N emissions from organic residues spreading: study of the onfield distribution patterns influence and of emission models sensitivity

Alice Gueudet¹, Marilys Pradel^{2*}, Jean-Christophe Roux², Afsaneh Lellahi³, Philippe Vissac¹

- (1) ACTA, 149 rue de Bercy, Paris, 75595, FR
- (2) Irstea, UR TSCF, Domaine des Palaquins, Montoldre, 03150, FR
- (3) ARVALIS-Institut du végétal, Service Agronomie Economie Environnement, Boigneville, 91720, FR

*Corresponding author: marilys.pradel@irstea.fr

Abstract

Recycling organic residues by direct spreading not only brings ecosystemic services but also induces emissions on field. Estimating N-N₂O, N-NH₃ and N-NO₃ emissions on field according to the technological performances of spreading machines considering different cultural systems, weather and climate conditions, various distributions on the field and different emission models was studied.

Organic residue distributions in the field were considered as Gaussian distributions using mean, bias and standard deviation. The study showed that $N-N_2O$ emissions estimated with the STICS model are rarely sensible to distribution and mostly depend on weather and climate conditions. Concerning the other types of emissions obtained with the STICS model and all the results from the IPCC method, the bias of the distribution seems to be the parameter which influences most the results. Ammonia emissions, in particular, show variability according to the scenario and the distribution pattern for all the emission models used.

Introduction

A study conducted in 2012 between Irstea, ACTA and Arvalis-Institut du vegetal aims to analyse the sensitivity of nitrogenous emissions resulting from organic residues spreading according to different parameters: (i) the distribution of the product on the field, (ii) the type of spread products: sewage sludge, manure and slurries, (iii) the weather and climate conditions and associated cropping systems, (iv) the models used to estimate nitrogenous emissions.

Spreading scenarios were built considering different cropping system in specific weather and climate conditions and different spread products. N-N₂O, N-NH₃ and N-NO₃ emissions were calculated for each scenario with 4 different models regarding the possible variability of product distribution at the on-field scale. The main parameters used in this study are summed up in Table 1.

Table 1. Parameters used to assess the on-field emission variability

Parameters		Number	Comments			
Average/optimal	Bias	11	0.5, 0.6, 0.7, 0.8, 0.9, 1, 1.1, 1.2, 1.3, 1.4 and 1.5			
distribution	Standard deviation	5	10, 15, 20, 25, 30			
Organic residues	Sludge	4	Dry, composted, limed, liquid Cattle and poultry manure (with 2 different agronomic			
	Effluents	7				
			characteristics of each), pig slurry and cattle slurry (with 2 different			
			agronomic characteristics)			
Weather and	Sludge	9	Pays de Caux (Cx), Kerlavic (Kl1, Kl2, Kl3), Mons (Ms1, Ms2 fo			
climate	Effluents	8	sludge, Ms 1 for effluents), Montoldre (Mt1, Mt2), Poitou			
conditions			Charentes (Pt)			
Emission models	$N-N_2O$	2	STICS, IPCC			
	$N-NH_3$	3	STICS, IPCC, EMEP			
	$N-NO_3$	2	STICS, EMEP			

Material and Methods

The method used is based on those proposed by Pradel et al, 2012 [1]. On one side, nitrogenous emissions are calculated for different application rate classes (from 0 to 200% of the desired application rate) with the STICS (Simulateur multidisciplinaire pour les cultures standard) model

[2], IPCC and EMEP factors for 4 types of sludges and 4 types of effluent characterised by different agronomic composition. On the other side, distributions in the field were considered following a Gaussian distribution characterized by mean and standard deviation: a bias of the distribution is built based on the mean to represent the difference between the average and the desired application rate calculated in order to fulfil crop requirements. Standard deviation represents the distribution variability of the application rate classes encountered on the field (in kg per ha). Each obtained emission through simulation for each application rate classes and each organic residue is multiplied by the % of surface of each application rate classes statistically generated to obtain to total emissions at the field scale. Example of results is given in Table 3. The difference between emissions obtained for each distribution is compared to the optimum distribution represented by a bias equal to 1 and a standard deviation equal to 10 (in bold in Table 2). The sum up of these differences for all the application rate leads to a benefit or a loss compared to the reference.

Table 2. % of surface receiving each application rate according to different bias and standard deviation (results shows for 3 of the 11 bias: 0.5, 1 and 1.5 and 3 of the 5 standard deviation: 10, 20, 30). The optimum distribution is given for bias 1 and standard deviation 10 (in bold).

Bias	Standard	Application rate classes (in %)										
devia	deviation	0	20	40	60	80	100	120	140	160	180	200
0.5	10	0.01	2.41	48.3	46.67	2.61	0	0	0	0	0	0
0.5	20	2.42	13.65	34.58	34.02	13.17	2.06	0.1	0	0	0	0
0.5	30	9.69	15.36	25.48	24.29	16.12	6.68	1.86	0.47	0.05	0	0
••••												
1	10	0	0	0	0.14	15.92	68.23	15.56	0.15	0	0	0
1	20	0	0	0.58	5.93	24.28	38.68	24.3	5.69	0.53	0.01	0
1	30	0.17	0.94	3.68	10.7	20.98	26.36	21.24	11.15	3.79	0.85	0.14
1.5	10	0	0	0	0	0	0.01	2.15	47.4	48.08	2.35	0.01
1.5	20	0	0	0	0	0.2	2.34	13.04	33.55	35.06	13.72	2.09
1.5	30	0	0	0.01	0.31	1.81	6.66	16.4	24.93	24.83	16.34	8.71

Table 3. Example of N-NO₃ emissions results simulated with STICS model in Pays de Caux system for a bias equal to 1 and a standard deviation equal to 10 for different sludge (in kg N-NO₃)

	composted sludge	dry sludge	limed sludge	liquid sludge
0%	0	0	0	0
20%	0	0	0	0
40%	0	0	0	0
60%	0.0084	0.0098	0.0098	0.0098
80%	0.9552	1.2736	1.1144	1.1144
100%	4.0938	5.4584	4.7761	5.4584
120%	0.9336	1.2448	1.2448	1.2448
140%	0.009	0.012	0.012	0.012
160%	0	0	0	0
180%	0	0	0	0
200%	0	0	0	0
sum	6.00	8.00	7.16	7.84

Results

Nitrogenous emission variability according to bias and standard deviation

Between 3960 and 9240 data were generated. Variability analysis were done regarding the bias with the standard deviation equal to 10 and the standard deviation with bias equal to 1. These variability analyses were studied according to the different weather and climate conditions and the different organic residues [3]. Three different observations can be done:

- 1. If the sum of the nitrogenous emissions (in grey in Table 3) is constant and equal to zero, no variation was observed because no emission was generated by the organic residue spreading whatever the bias and the standard deviation used. This was observed using STICS model for N-N₂O emissions in Kerlavic system for all the effluents, N-N₂O emissions in all the systems except for Kerlavic and Poitou Charentes for all the sludges and N-NH₃ emissions for Kerlavic system for all the sludges.
- 2. If the sum of the nitrogenous emissions (in grey in Table 3) is constant and different from zero, no variation was observed whatever the bias and the standard deviation used because the emissions are the same whatever the application rate classes. This was observed using STICS model for N-N₂O emissions in Kerlavic system for all the sludge, N-NO₃ emissions for 1 to 5 system regarding the sludge, N-N₂O emissions in 3 systems for all the effluents and N-NO₃ emissions for Kerlavic and Mons systems for poultry manure.
- 3. If the sum of the nitrogenous emissions (in grey in Table 3) is variable regarding the bias and the standard deviation, we observe that for all the bias and standard deviation for all the systems and the organic residues using IPCC and EMEP factors, emissions are proportional to the nitrogen content of the organic residues. In those cases, the variability is not significant. Using STICS model, N-N₂O emission variability for sludges is significant for one system for a bias at least inferior 0.8 and superior to 1.2. N-NH₃ emission variability is significant for numerous biases for limed and liquid sludges and for composted sludge in 3 systems and dry sludge in 5 systems and for all tested effluents. N-NO₃ emission variability is only significant for the bias for sludge spreading. No significant variability was observed for N-NO₃ emissions whatever the bias and the standard deviation studied.

Comparison of nitrogenous emission variability according to the model used

The following tables represent the minus, maximum, average and standard deviation for each studied nitrogenous emission regarding the organic residues and the emission model used.

When the "distribution" factor is studied, sum of emissions are more sensitive to bias than standard deviation whatever the studied emission model. The variability of the sum of emissions is significant in numerous spreading scenarios for sludge and effluents for N-NH₃ and in some cases for N-NO₃. STICS N-NO₂ emissions are barely sensitive to this factor. When the "organic residues" factor is studied, a classification is possible for N-NH₃ and N-NO₃ emissions in case of sludge spreading scenarios. This is less clear for effluents, no classification can be established.

Table 4. Minus, maximum, mean and standard deviation (SD) of emission sums, benefits and loss compared to the reference for N-N₂O regarding products and emission models

Emission type Organic product Models		N-N ₂ O (UN/ha)					
		Slud	lges	Effluents			
		STICS	IPCC	STICS	IPCC		
	Minus	0,00	0,13	0,00	0,27		
Cum of amissions	Maximum	37,51	3,17	55,48	5,63		
Sum of emissions	Mean	3,33	1,70	23,70	2,56		
	SD	9,17	0,77	19,32	1,12		
D (". () 11	Minus	-3,46	-0,39	-1,26	-1,26		
Benefits (+) and losses	Maximum	3,35	0,41	0,78	1,26		
(-) compared to the	Mean	0,01	0,00	0,00	0,00		
optimal distribution	SD	0,49	0,17	0,10	0,55		

When the "weather and climate condition" factor is studied, no classification can be done for N-NO₂ ad N-NH₃ emissions but a classification is possible for N-NO₃ emissions estimated with STICS model. We observe that cropping system is less sensitive than the weather and climate condition in the generation of nitrogenous emissions. When the emission model factor is studied, orders of magnitude are different between STICS and IPCC for N-N₂O emissions where STICS emissions are greater than

IPCC ones. IPCC and EMEP models used for N-NH₃ emissions lead to greater values than STICS model ones. These differences are linked with the type of model (national inventory for IPCC and EMEP and cropping system for STICS).

Table 5. Minus, maximum, mean and standard deviation (SD) of emission sums, benefits and loss compared to the reference for N-NH₃ regarding products and emission models

			0.1				
Emission ty	N-NH ₃ (UN/ha)						
Organic product			Sludges			Effluents	
Models		STICS	IPCC	EMEP	STICS	IPCC	EMEP
	Minus	0,00	2,64	0,80	0,00	3,85	3,85
С С : :	Maximum	15,56	31,57	10,52	98,63	62,41	53,55
Sum of emissions	Mean	1,92	15,99	4,65	10,82	29,84	23,89
	SD	2,80	6,47	1,82	18,26	12,55	11,21
Benefits (+) and	Minus	-5,24	-5,51	-2,51	-32,98	-17,73	-19,18
losses (-) compared	Maximum	5,19	5,72	2,61	33,24	17,71	19,17
to the optimal	Mean	-0,01	0,01	0,00	0,00	0,01	0,01
distribution	SD	1,04	2,32	0,95	6,39	7,56	5,20

Table 6. Minus, maximum, mean and standard deviation (SD) of emission sums, benefits and loss compared to the reference for N-NO₃ regarding products and emission models

Emission	N-NO ₃ (UN/ha)					
Organic product Models		Slud	ges	Effluents		
		STICS	IPCC	STICS	IPCC	
	Minus	0,02	4,14	1,36	3,85	
Sum of emissions	Maximum	91,00	70,01	82,92	100,93	
Sum of emissions	Mean	27,17	38,34	18,97	43,89	
	SD	32,28	16,97	25,56	21,26	
Danafita (1) and leases	Minus	-13,48	-8,26	-1,13	-26,59	
Benefits (+) and losses	Maximum	15,07	8,57	0,59	26,56	
(-) compared to the	Mean	0,02	0,01	0,00	0,02	
optimal distribution	SD	1,80	3,49	0,11	11,49	

Conclusions

This sensitivity analysis on the sum of emission was a qualitative one. It would be interesting to analyse the robustness of what we observed using some statistical tools. Only theoretical distribution was used in this study but realistic distribution in the field is obviously different and cannot be therefore characterized by a bias and a standard deviation. A second study, based on this observation, was conducted in 2012 [4].

References

- [1] Pradel et al, 2012. Valorisation of Organic Wastes Through Agricultural Fertilization: Coupling Models to Assess the Effects of Spreader Performances on Nitrogenous Emissions and Related Environmental Impacts, Waste and Biomass Valorisation, 22 pages, DOI 10.1007/s12649-012-9162-2.
- [2] Brisson, N., Mary, B., 1998. STICS: a generic model for the simulation of crops and their water and nitrogen balances. I. Theory and parameterization applied to wheat and corn. Agronomie 18(5–6), 311-346.
- [3] Gueudet A., 2012. Etude de sensibilité relative aux émissions azotées générées par l'apport d'engrais organiques au champ selon différents facteurs : types de répartition du produit sur la parcelle, sites, couples machines produits et modèles d'émissions choisis. Rapport Action CASDAR « ACV et Fertilisation ». Décembre 2012. 41 pages
- [4] Gueudet A., 2012. Evaluation de la variabilité de la répartition et du dosage au cours de chantiers d'épandage en fonction de facteurs influant sur ces derniers Rapport Action CASDAR « ACV et Fertilisation ». Décembre 2012. 49 pages