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▶ To cite this version:

Vincent E. J. Jassey, Geneviève Chiapusio, Daniel Gilbert, Marie-Laure Toussaint, Philippe Binet. Phenoloxidase and peroxidase activities in Sphagnum-dominated peatland in a warming climate. Soil Biology and Biochemistry, 2012, 46, pp.49-52. hal-00682547

HAL Id: hal-00682547 https://hal.science/hal-00682547

Submitted on 26 Mar 2012

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18 Abstract

19 Peatlands still suffer from the scarcity of available data about the characterization and the response to climate forcing of the main oxidative enzymes that occur over the seasons. In the 20 present study, phenoloxidase and peroxidase activities were examined in Sphagnum lawns 21 along a narrow fen-bog gradient under experimental elevated temperatures. We showed that 22 peroxidase activities from Sphagnum mosses were 1000-fold higher than those of 23 phenoloxidases irrespective of seasons and sampling areas. Peroxidase activities increased (+ 24 30%) with the rise of air temperatures (an average of 1°C), while warming did not alter 25 phenoloxidase activities. These results suggest that the monitoring of peroxidase activities in 26 peatlands may represent a suitable and forward indicator of the impact of climate warming on 27 carbon cycle in peatlands. 28

Keywords: 2,7-diaminofluorene; peroxidases; phenoloxidases; climate warming; peatland;
open top chambers

Extracellular phenoloxidase and peroxidase measurements in ecosystems provide essential 32 information on the stability of the carbon cycle (Sinsabaugh, 2010; Theuerl et al., 2010). By 33 contributing to the oxidation and transformation of both complex and simple phenolic 34 35 compounds, these enzymes induce partial or complete degradation of such recalcitrant compounds, and finally act on carbon cycling (Baldrian, 2006; Sinsabaugh, 2010; Theuerl et 36 al., 2010). Considering the ongoing global warming, these enzymes gain scientific concern in 37 38 terrestrial carbon reservoirs, such as peatlands (Fenner et al., 2005; Laiho, 2006; Jassey et al., 2011b). The accumulation of carbon in peat soils is thought to partly result from a suppression 39 of the normal pathways of enzymatic decomposition in which oxidative enzymes, such as 40 phenoloxidases, play a key role (Freeman et al., 2001, 2004). 41

Although phenoloxidases (PO) involved in the degradation of polyphenols are divided 42 into PO O_2 (e.g. laccases, tyrosinases) and PO H_2O_2 (e.g. lignin and manganese peroxidases) 43 dependent (Criquet et al., 2000a; Duran & Esposito, 2000; Sinsabaugh et al., 2003, Alarcón-44 Gutiérrez et al., 2008; Sinsabaugh, 2010), only phenoloxidase O₂ dependent have been 45 predominantly investigated to date in peatlands. In forest litters, different spatiotemporal 46 variations of PO O₂ and H₂O₂ dependent were recorded, emphasizing that all of these 47 extracellular enzymes do not respond similarly to environmental changes (Criquet et al., 48 2000a; Duran and Esposito, 2000; Alarcón-Gutiérrez et al., 2008; Kaiser et al., 2010). Thus, 49 peatlands suffer from the scarcity of available data about the characterization and the response 50 to warming of these oxidative enzymes that occur over the seasons and ecological settings. 51

The purpose of the present paper was to determine the impact of an experimental climate warming on phenoloxidase activities O_2 and H_2O_2 dependent in peatlands over two seasons along a transitional fen-bog gradient. Because our assays did not discriminate individual enzymes, the generic terms phenoloxidase and peroxidase were chosen to describe the activity of enzymes that use O_2 and H_2O_2 as an acceptor, respectively.

During field campaigns of summer and autumn 2010, peroxidases and phenoloxidases 57 were investigated within a larger mire complex in fen and bog areas situated in the Jura 58 Mountains (France, 46°49'35''N, 6°10'20''E). Sampling areas were situated along a 59 transitional gradient between a poor fen and a raised bog with vegetation composition 60 dominated by Sphagnum fallax (Jassey et al., 2011b). Samples of S. fallax were collected and 61 cut into two levels: 0-3 cm (living segments = Top) and 3-10 cm (early declining segments = 62 Bottom) from the capitulum. In fen and bog areas, 6 Sphagnum plots were selected in 63 representative surfaces including 3 replicates as ambient treatment and 3 replicates as 64 warming treatment. The beginning of the warming treatment was on April 2008. Increasing of 65 air temperature was passively achieved in warming plots using open-top chambers (hereafter 66 referred as OTC) over the vegetation (Jassey et al., 2011b). Air temperatures (10 cm above 67 Sphagnum surface) were monitored continuously in each plot. 68

Because soil organic matter could affect enzyme activities, a specific method of 69 extraction was used (Criquet et al., 1999). 3 g FW of S. fallax in 50 mL 0.1 M CaCl₂ with 70 0.05% Tween 80 and 20 g PVPP were shaken for 1h. After centrifugation, the supernatant 71 was filtrated (0.2 µm) and concentrated in cellulose dialysis tubing (10 kDa molecular mass 72 cut-off) covered with polyethylene glycol. Then, concentrated extracts were resuspended in 73 phosphate buffer (pH 5.6) until 1/10 of the initial volume. Enzyme activities were measured 74 by spectrophotometry using a 96 wells microtiter plate. For phenoloxidase quantification, 75 each replicates wells contained 150 µL of enzyme-extract with, either 100 µL of L-DOPA (10 76 mM), or 2 μ L of 2,7-diaminofluorene (DAF; 0.68 mM; $\epsilon^{M} = 10$ 228 M⁻¹.cm⁻¹), or 2 μ L of 77 syringaldazine (5 mM; ϵ^{M} = 65 000 M⁻¹.cm⁻¹) or 5 µL of ABTS (0.1 mM; ϵ^{M} = 36 000 M⁻¹.cm⁻¹ 78 ¹) in assay wells, and monitored at 460, 600, 525 and 420 nm, respectively (Criquet et al., 79 2000a; Jassey et al., 2011b). Peroxidase activities were measured using 2 µL DAF (0.68 mM) 80 with 10 μ L of H₂O₂ (0.3%), and manganese peroxidases (Mn-peroxidases) with 12 μ L of 81 MnSO₄ (0.1 mM) (Criquet et al., 2001). Their oxidation rate was monitored at 600 nm. 82

Peroxidase activity was subtracted to the Mn-peroxidase assay to obtain the Mn-peroxidase activity. We also quantified fungal lignin-peroxidases using 12 μ L veratrylic alcohol (0.4 mM; $\epsilon^{M} = 9300 \text{ M}^{-1} \text{ cm}^{-1}$) with 10 μ L of H₂O₂ (0.3%) and monitored at 310 nm in quartz cuvettes (Tien and Kirk, 1984). Enzymatic activities were expressed as one μ mol of substrate oxidized per minute per gram of dry mass (U.g⁻¹ DM).

Differences of phenoloxidase and peroxidase pools between *Sphagnum* segments, sampling areas, seasons and climate treatments were analysed using repeated measures ANOVA with time as a within subject repeated factor (time = 2: Summer and Autumn) and warming treatment, area or sampling depth as between-subject factors. Interactions between sampling area, seasons and treatments were also considered. The assumptions of parametric tests were visualized and tested. An identical procedure was used to detect differences of air and soil temperatures between ambient and warming plots.

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96 The highest values of phenoloxidase activities were recorded with DAF. 97 Syringaldazine and ABTS substrates induced a formation of precipitates after 1h of enzyme 98 kinetic. Indeed, quinones produced from syringaldazine or ABTS were not soluble over time 99 in aqueous medium (Floch et al., 2007). Phenoloxidase activities quantified with DAF 100 showed significant differences along *Sphagnum* segments and between fen and bog ecological 101 areas, whereas phenoloxidases quantified with L-DOPA did not change (Fig. 1). Therefore, 102 the use of DAF is ideal to quantify phenoloxidases in *Sphagnum* peatlands.

We showed that peroxidases constituted the main oxidative system in *Sphagnum*peatlands, with values of peroxidase activities 1000-fold higher than those of phenoloxidases (Figs 1 and 2). Although peroxidase activities 120-fold higher than phenoloxidase activities have been recorded in surface forest litter (Alarcón-Gutiérrez et al., 2009), peroxidases 1000fold greater than phenoloxidases is unique, to our knowledge (Sinsabaugh, 2010). Moreover,

such ratio strongly suggested that peroxidases had a plant origin for several reasons. First, 108 Sphagnum lawns largely dominate fen and bog areas. Second, DAF-H₂O₂ is known to be the 109 most sensitive substrate for the detection of plant-peroxidases (Criquet et al., 2000b, 2001). 110 Third, no fungal lignin-peroxidase activity was highlighted ($< 10^{-7} \text{ U.g}^{-1} \text{ DM}$), corroborating 111 the Sphagnum origin of our high peroxidase activities. Furthermore, among the different 112 fungus taxa identified in Sphagnum litter, few (only 24%) are identified as phenoloxidase 113 producers (Thormann et al., 2001). Abiotic conditions (e.g. acidic conditions and 114 waterlogging) of peatlands are largely known to limit fungal oxidation activity (Williams et 115 al., 2000; Toberman et al., 2008, 2010). 116

117 Significant decreased of peroxidases were recorded both along *Sphagnum* segments 118 and over the seasons (ANOVA, P < 0.01; Fig. 2). Such temporal variations suggest a positive 119 relationship between peroxidase activities and polyphenol content. Jassey et al. (2011a, b) 120 actually demonstrated that phenolic release from *Sphagnum* mosses changed over seasons, 121 and decreased along *Sphagnum* shoots.

Warming by OTCs significantly increased the daily average air temperature (ANOVA, 122 P < 0.001) in both sampling areas (an average increase of 1 °C). The increase of air 123 temperatures also induce higher evapotranspiration in temperate zones, which result in lower 124 Sphagnum moisture content during summertime (Jassey et al., 2011b). Despite this rise of air 125 temperatures, phenoloxidase activities were not significantly influenced. Previous studies, 126 which attempted to evaluate warming effect on phenoloxidases in peatlands, found equivocal 127 results and concluded that the interactive effects of moisture and pH largely inhibited their 128 oxidation activity (Toberman et al., 2010; Jassey et al., 2011b). On the contrary, the rise of air 129 130 temperatures led to a significant increase of peroxidase activities in the fen area (+ 30%), especially in living top segments (ANOVA, P = 0.017; Fig. 2). The response of peroxidase 131 activities to climate warming between the fen and bog areas also showed that oxidative pools 132 change in different directions in response to climate warming, as already showed with 133

phenolics (Jassey et al., 2011b). Although temperature was identified as an enhancer of
peroxidase activities, it still remains difficult to predict the effect of global warming on soil
organic matter sequestration in peatlands because of multiple functions of peroxidases, both
in mineralization and humification pathways (Sinsabaugh, 2010).

To conclude, our results point out that (i) the DAF is a relevant oxidative substrate to quantify both phenoloxidase and peroxidase activities in enzymatic extract from *Sphagnum* lawns, (ii) *Sphagnum*-peroxidase activities constituted the main oxidative system in *Sphagnum*-peatlands and (iii) the monitoring of plant-peroxidases represents a suitable and forward indicator of changes in carbon cycle in peatlands under a climate warming.

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144 Acknowledgements

This research is a contribution of the ANR PEATWARM project (Effect of moderate warming on the functioning of *Sphagnum* peatlands and their function as carbon sink).
PEATWARM is supported by the French National Agency for Research under the Vulnerability: Environment—Climate Program (ANR-07-VUL-010). Further funding to VEJ Jassey by the Franche-Comté Region is kindly acknowledged.

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225 Figures:

Figure 1: Fungal phenoloxidase activities (mean \pm S.E.; n = 3) characterized by DAF or L-DOPA substrates along the fen-bog gradient of the Forbonnet peatland in summer and autumn autumn different *Sphagnum* segments. *Asterisks* indicate significant differences of phenoloxidase activities (ANOVA; *P* < 0.05) between *Sphagnum* segments. *Letters* indicate significant differences of phenoloxidase activities (ANOVA; *P* < 0.05) between the fen and the bog area. Top = 0-3 cm; Bottom = 3-10 cm from capitulum.

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Figure 2: Activity of peroxidases produced by *Sphagnum* (mean \pm S.E.; n = 3) along the fenbog gradient of the Forbonnet peatland in summer and autumn 2010 in different *Sphagnum* segments. *Asterisks* indicate significant differences of peroxidase activities (ANOVA; P <0.05) between *Sphagnum* segments. *Letters* indicate significant differences of peroxidase activities (ANOVA; P < 0.05) between seasons (summer/autumn). Triangles (Δ) indicate significant differences between ambient and warming plots. Perox = peroxidases; Mn-perox = manganese peroxidases. Top = 0-3 cm; Bottom = 3-10 cm from capitulum.

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