



HAL
open science

Home and community composting for on-site treatment of urban organic waste: perspective for Europe and Canada

Benu Adhikari, A. Trémier, José Martinez, S. Barrington

► **To cite this version:**

Benu Adhikari, A. Trémier, José Martinez, S. Barrington. Home and community composting for on-site treatment of urban organic waste: perspective for Europe and Canada. *Waste Management and Research*, 2010, 28 (11), p. 1039 - p. 1053. 10.1177/0734242X10373801 . hal-00677364

HAL Id: hal-00677364

<https://hal.science/hal-00677364v1>

Submitted on 8 Mar 2012

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

25 **Abstract**

26
27 Management of municipal solid waste (MSW) is a challenge faced by urban centres worldwide,
28 including the European Union (EU) and Canada, as a result of urbanization and economic
29 prosperity accelerating the generation of municipal solid wastes (MSW) along with its organic
30 fraction. Within a concept of waste recovery, source separation and on-site treatment of urban
31 organic waste (UOW) can resolve major economic issues and faced by urban centres along with
32 environmental and social issues associated with landfilling. In this context and as compared to
33 the traditional landfilling practice (Base Sce), this paper examines on-site UOW composting
34 strategies using a combination of centralized composting facilities (CCF), community
35 composting centres (CCC) and home composting (HC) (Sce 1, 2 and 3). This study consists of a
36 feasibility and economic study based on available data and waste management costs. This study
37 indicates that on-site treatment of UOW using practices such as home and community
38 composting can lower management costs by 50, 37 and 34 % for the rich European countries
39 (annual GDP over \$25000 US), the poorer European countries (annual GDP under \$25000 US),
40 and Canada, respectively. Furthermore, on-site composting can reduce greenhouse gas (GHG)
41 emissions by 40 % for Europe and Canada, despite gas capture practices on landfill sites.
42 However, the performance of home composters and the quality of the compost products are
43 issues to be further addressed for the successful implementation of UOW on-site composting.

44
45 **Keywords:** municipal solid waste, urban organic waste, greenhouse gas, landfill, composting

46 **Introduction**

47 Depending on the country's economic activity, organics represent 20 to 80 % of the municipal
48 solid waste (MSW) main stream and therefore constitute one of its major fractions (Adhikari et
49 al. 2009; Adhikari et al. 2006; EEA & ETC-WMF 2002; Papadopoulos et al. 2009). The
50 improper disposal of urban organic wastes (UOW), composed mostly of food and green wastes,
51 results in well-known health and environmental issues: attraction of insects and rodents;
52 development sites for parasites, pathogens and viruses; contamination of drainage waters, and;
53 emissions of unpleasant odours and greenhouse gases (Kumar et al. 2009; Moghadam et al.
54 2009; Rasapoor et al. 2009; Turan et al. 2009).

55 All countries worldwide can benefit from reducing the generation of MSW through
56 recycling and reusing. In Asian countries, the expansion of urban centres and their economic
57 growth have exponentially increased the production of MSW along with the mass of UOW
58 (Adhikari et al. 2009). Since several major cities in Asia can only afford to collect 30 % of their
59 MSW, their growth has further stressed issues associated with collection and disposal
60 (Guermoud et al. 2008; Harjula et al. 2001; Parrot et al. 2009; Troschinetz & Michelcic 2009;
61 Vehlow et al. 2007). In industrialized countries of Europe and North America, landfilling is still
62 the most common practice for the disposal of MSW along with UOW although social pressures
63 are making it harder to find proper sites (De Baere 2000; Environment Canada 2009; Messineo
64 & Panno 2008). Furthermore, landfilled organic wastes (OW) emit greenhouse gases which can
65 be recovered at a cost, generate leachate which requires treatment and remove land which
66 otherwise could be used for agriculture (Machado et al. 2009; Official Journal 2000; Wagner &
67 Arnold 2008). In an attempt to reduce the number of landfills along with their social and
68 environmental impacts, European and North American countries have adopted policies aimed at
69 reducing the generation of wastes through for example recycling and reuse (Landfill Directive
70 1999; CCEM 1989).

71 To be diverted from landfills, OW must be stabilized with the objective among others of
72 producing a soil amendment. In North Canada, composting to produce a soil amendment is likely
73 the most popular treatment, while in Europe, countries such as Germany and the Netherlands
74 have encouraged source separation for composting and biogas production through anaerobic
75 digestion (Table 1). Other European countries such as Spain and France, allow the stabilization
76 of the OW fraction by composting the entire MSW mainstream before landfilling; these
77 countries also mechanically and source separate the OW to produce soil amendments (Kelleher
78 2007). Although the source separated OW produces a compost of higher quality and value, this
79 management option through centralized composting facilities increases the collection and
80 transportation costs, besides that of the disposal method. Source separation of UOW for its
81 composting or anaerobic digestion requires a separate collection increasing transportation cost.
82 Cities have generally coped with this issue by collecting UOW weekly and the rest of MSW
83 twice monthly. In terms of treating source separated UOW and as compared to landfilling costs
84 of \$30 to 50 US ton⁻¹ of MSW including greenhouse gas capture, composting and anaerobic
85 digestion cost from \$50 to 400 US ton⁻¹, with an end product not even meeting the recycling cost
86 at \$5 to 10 US ton⁻¹ wholesale.

87 **Table 1**

88 Therefore, the recycling of OW into a high quality soil amendment requires additional
89 investments as compared to the traditional method of landfilling even where greenhouse gases
90 are captured and treated (Burnley 2007; Dunne et al. 2008; Hazra & Goel 2009). This additional
91 investment is an economic burden for developed as well as developing countries. As an
92 alternative method of reducing if not eliminating collection and transport costs to compensate for
93 the treatment cost, community composting centres (CCC) and home composting systems (HC)
94 are proposed. Nevertheless, such systems have not yet been demonstrated as sanitary, and
95 economically and environmentally advantageous.

96 Within the recycling legislative framework of Europe and North America, the aim of this
97 study was to investigate the economical and environmental advantages of the on-site treatment of
98 UOW by community or individual household composters. The present feasibility and economic
99 study is based on available waste management costs and environmental knowledge. In this study,
100 food and garden wastes generated from households, institutions and businesses make up the
101 UOW fraction of MSW.

102

103 **The European and Canadian UOW generation and management**

104 In this study, the 27 member states of the European Union (EU) will be split into groups 1 and 2
105 (EUG1 and EUG2) consisting of countries with a gross domestic product (GDP) in excess of and
106 under \$25000 US capita⁻¹ year⁻¹ (Table 2), respectively. Canada and the EU are similar in
107 economy but differ in their landmass and population density. The EUG1 and EUG2 have
108 population densities of 136 and 92 persons km⁻² whereas Canada has a population density of
109 only 4 persons km⁻² (World population prospects, 2007). Nevertheless, the Canadian population
110 is mainly concentrated along its southern border, for a more representative density of 20 person
111 km⁻². To consider the different contexts of economy, urbanization and population density, the
112 following sections separately discuss the MSW generation and management for the EU, the
113 EUG1, the EUG2 and Canada.

114

115 ***MSW and UOW generation***

116 The MSW and UOW generated by the EUG1, the EUG2 and Canada are presented in Table 2.
117 For Canada, the quantities of MSW and UOW correspond only to household waste whereas for
118 the EU, the quantities correspond to household, institutions and commercial wastes (OECD
119 2006-2007). In 2005, the EUG1 generated 207 million tonnes of MSW or 2.0 kg person⁻¹ day⁻¹,
120 representing 82 % of total MSW generated by the EU. The OW fraction represented 32 % of the

121 total MSW main stream and amounted to 0.63 kg person⁻¹ day⁻¹. Also in 2005, the EUG2
122 generated 44 million tonnes of MSW or 1.56 kg person⁻¹ day⁻¹ representing 18 % of the total
123 MSW generated by the EU. The generated OW fraction amounted to 25 % of the MSW
124 production for the EUG2. With an average gross domestic product (GDP) of \$35000 US capita⁻¹,
125 the EUG1 produced 28 % more MSW per person than the EUG2 with half the GDP of \$17000
126 US capita⁻¹.

127 In North America and for 2005, the Canadian urban population of 25.8 million produced
128 13.4 million tonnes of MSW or 1.42 kg person⁻¹ day⁻¹ for a GDP of \$33400 US capita⁻¹. In
129 comparison, the US and Mexican urban populations of 242.3 and 79.6 million, respectively,
130 produced 180.1 and 27.5 million tonnes of MSW or 2.5 and 1.2 kg person⁻¹ day⁻¹, for a GDP of
131 \$41400 and \$10700 US capita⁻¹. Accordingly, the calculated UOW generation for Canada, the
132 US and Mexico was 0.43, 0.63 and 0.63 kg person⁻¹ day⁻¹, respectively.

133 For the EU and Canada, UOW is one of the major fractions of the MSW main stream
134 (Table 3) representing in 2005, 30 and 25 % of the MSW main stream, respectively. Other
135 reported components of the MSW main stream were paper and paperboard, glass, metal, plastics
136 and textile.

137 **Insert Table 2 & Table 3**

138

139 *European and Canadian UOW management practices*

140 Landfilling is still the most common MSW disposal method in both the European Union and
141 Canada. In 2005, the EUG1 and EUG2 landfilled 36 and 82 % of the total MSW main stream,
142 whereas Canada landfilled 60 % (Table 2). Diversion of MSW from landfilling practices vary
143 widely among the EUG1 countries with Germany and the Netherlands reaching over 98% as
144 compared to the UK still at 22 % (Table 1). In Malta, Greece and the Czech Republic, more than
145 80 % of all MSW was still being landfilled in 2005 (EEA 2009).

146 The low landfill diversion rate for most European Countries and Canada is far from
147 meeting environmental policy expectations. In 1999 and for the biodegradable fraction of MSW,
148 the EU Landfill Directive 99 (Article 5) set diversion objectives of 25 % by 2006, 50 % by 2009,
149 and 65 % by 2016, based on 1995 levels. Similarly in Canada, the Canadian Council of Ministers
150 of the Environment (CCME) proposed a national diversion goal of 50 % of all MSW by 2000
151 based on that disposed in 1989, without specifically targeting the organic fraction (Wagner &
152 Arnold 2008).

153 Individual country policies have influenced the level of diversion and the technology
154 preferred for this diversion. Incineration is not widely used because of issues of atmospheric
155 emissions and the fact that the high moisture content of UOW reduces the caloric value of the
156 process (Zsigraiová et al. 2005; El Asri & Baxter 2004; Marton & Alwast 2002; Environment
157 Canada, 2009). While in the EUG1 countries, 21 % of all MSW was incinerated in 2005, less
158 than 6 % was treated by this process in the EUG2 and Canada (EEA, 2009). The high EUG1
159 incineration level results in part from countries such as France, Germany and The Netherlands
160 using this technology to divert over 30 % of their MSW while generating energy. In 2007 and
161 within the EUG1, Germany and The Netherlands were diverting 98 to 99 % of their MSW (Table
162 1), respectively, because of strict and costly landfilling and incineration regulations encouraging
163 composting, and a subsidy on biogas production from the anaerobic digestion of organic wastes.
164 In Canada since 2000, only the province of Nova Scotia succeeded in composting 60 % of its
165 OW (Wagner & Arnold 2008), through strict regulations prohibiting among others, the disposal
166 of OW through landfills. However, recycling dropped to 41 % in 2006 because of a higher level
167 of waste generation as a result of economic growth and changes in consumer habits. British
168 Columbia is second in Canada with 32 % diversion from landfill through recycling mainly
169 because of a bylaw requiring manufacturers to recover packaging (Table 1). With voluntary
170 recycling policies for the rest of the Canadian provinces, less than 20 % of all MSW is diverted

171 from landfills excess for Quebec thanks to its public organization Recy-Quebec managing the
172 recycling of metals, paper and glass (Environment Canada 2009).

173 Composting and anaerobic digestion are commonly used to treat and recycle the organic
174 fraction of MSW. In 2006 and for the EU, 124 central composting facilities were treating 4
175 million tonnes of MSW annually (Kelleher 2007). In 2005, France, Spain and The Netherlands
176 were composting 14, 33 and 24 % of all MSW, respectively (Table 4), while in Canada, 12 %
177 was composted. In the past, centralized mixed MSW composting facilities were built and
178 operated, while at present, the composting of source separated OW facilities are preferred to
179 assure the quality and value of the finished product. In Germany, source separation of the
180 organic fraction of MSW is mandatory in many municipalities and at present, 700 to 900
181 composting facilities are in operation (Table 5). In France, amongst 119 composting facilities
182 treating MSW, 54 use source separated UOW. In Canada, 54 facilities compost source separated
183 food waste generated from residences, industries, businesses and institutions (Table 5). The
184 construction of a centralized composting facility can easily cost \$7 million US, for a UOW
185 processing cost of at least \$140 US ton⁻¹.

186 Anaerobic digestion is another technology used to divert UOW from landfill sites and
187 produce energy, but its application is generally accompanied by an incentive to generate energy
188 (Table 1). The highest mass of organic waste diverted using this technology is found in Germany
189 and the United Kingdom with 2000 kilo tonne of oil equivalent (ktoe) yr⁻¹ followed by Italy and
190 Spain with 300 ktoe yr⁻¹ and the France, Austria and The Netherlands with 100 ktoe yr⁻¹
191 (European Biomass Industry Association 2006). To encourage such source of green energy, EU
192 countries must generally offer a subsidy equivalent to the cost of electricity produced by
193 conventional technologies. At a crude oil price of \$100 US barrel⁻¹, such a green energy policy
194 costs the EU some \$4 billion US year⁻¹.

195 **Insert Table 4 & Table 5**

196 In order to reduce the cost of treating and recycling UOW, a more interesting diversion
197 method is needed. Within this objective, the on-site treatment of source separated UOW was
198 proposed to produces a dry, stabilized and volume reduced soil amendment (Adhikari et al.
199 2009) readily available for urban gardens. Nevertheless, the real economical impacts are issues
200 to be addressed before recommending the wide use of home and community composting centres.
201 Within this context, the following sections formulate and compare strategies for on-site
202 composting of UOW in the EUG1, EUG2, and Canada.

203

204 **Comparison of on-site UOW composting scenarios diverting MSW from landfills**

205 The following sections examine the growth of MSW and UOW over the upcoming 15 years and
206 then, predict the economic and environmental advantages associated with on-site composting to
207 divert the UOW fraction from the MSW mainstream and landfilling operations.

208

209 ***Estimated growth in MSW and UOW production***

210 The Adhikari et al. (2006) study is updated study by including rather than estimating the 2005
211 urban population (UP) and gross domestic product (GDP) data according to UNPD (2007) and
212 UNSD (2008). The growth of MSW and UOW was estimated by correlating the 2005 urban
213 population of each country (UNPD 2007) with its 2005 GDP (UNSD 2008) (Figure 1a & b).
214 Similarly, MSW and UOW generation rates were correlated with GDP (Figure 1c) according to
215 Adhikari et al. (2006). Estimating urban population expansion based on future economic
216 improvement and population growth, MSW and UOW production were computed from the
217 following equations for 2009, 2016, 2020 and 2025:

$$218 \quad (MSW)_{CRY} = 3.65 \times 10^{-9} \times (UP)_{CRY} \times (MSWR)_{CRY} \times (TP)_{CRY} \quad (1)$$

$$219 \quad (UOW)_{CRY} = 3.65 \times 10^{-9} \times (UP)_{CRY} \times (UOWPR)_{CRY} \times (TP)_{CRY} \quad (2)$$

220 where, $(MSW)_{CRY}$ is the MSW production (million tonne yr^{-1}); $(UP)_{CRY}$ is the urban population
221 (%); $(TP)_{CRY}$ is the total population; $(MSWR)_{CRY}$ is the MSW production rate ($\text{kg capita}^{-1}\text{day}^{-1}$);
222 $(UOW)_{CRY}$ is the UOW production (million tonne yr^{-1}); $(UOWPR)_{CRY}$ is the UOW production rate
223 ($\text{kg capita}^{-1}\text{day}^{-1}$), and; in the subscripts CRY , C refers to the country, R to the continents of
224 Europe and North America and Y to the year.

225

226 **Insert Figure 1**

227

228 The short term global economic recession was assumed to have a negligible impact on
229 MSW generation and management systems. From Equations (1) and (2) and for 2009 to 2025,
230 the estimated growth in MSW and UOW in the EU, the EUG1, the EUG2 and Canada are
231 presented in Figure 2. With the largest population, the EUG1 is expected to increase its MSW
232 production from 202 to 263 million tonnes yr^{-1} , over the next 15 years (2009 to 2025), resulting
233 in an UOW production increasing from 75 to 92 million tonnes yr^{-1} . Over the same period, the
234 EUG2 with the second largest population is expected to increase its UOW production from 11 to
235 13 million tonnes yr^{-1} . The UOW production is expected to increase by 23 % and 18 % in the
236 EUG1 and the EUG2 respectively. With 8 million tonnes yr^{-1} of UOW in 2009, Canada's
237 production is expected to reach 10 million tonnes yr^{-1} in 2025 representing an increase of 25 %.

238

239 **Insert Figure 2**

240

241 *Scenarios for UOW treatment strategies*

242 The economic and environmental impacts of the proposed scenarios are evaluated in this section
243 for the upcoming decades. The Base Scenario (Base Sce) assumes that all UOW will continue to
244 be landfilled, but that 80% of their biogas generation will be captured; Scenario one (Sce 1)

245 considers the practices of diverting 57, 16, and 44 % of all UOW from landfilling for the EUG1,
246 the EUG2 and Canada respectively, and treating the diverted waste through a centralized
247 composting facility (CCF) except for 1 % which would be treated through home composter
248 (HC); Scenario two (Sce 2) assumes that UOW diversion from landfill increases from 25 % in
249 2006 to 65 % in 2016 and 80 % in 2025 and the diverted UOW is composted at centralized
250 composting facilities (CCF), and; scenario three (Sce 3) considers zero landfilling with 10, 60,
251 and 30 % of UOW treated using CCF, HC and community composting centres (CCC)
252 respectively by 2025 (Table 6). Community composting centres (CCC) are considered to be on-
253 site treatment systems because of their close proximity to producers who can drop off the waste
254 on their way to work or during other activities.

255

256 **Insert Table 6**

257

258 *Economical assumptions*

259 Table 7 compares the cost of various composting strategies to that of landfilling UOW. Disposal
260 of UOW through landfilling requires: land acquisition; capital, operating and closure costs, and;
261 collection and transportation of UOW to landfill sites generally located at some distance from the
262 city (Adhikari et al. 2009). Besides the collection and transportation costs estimated at \$115 US
263 tonne⁻¹, a landfill-dumping fee of \$50 US tonne⁻¹ is quite common in Europe and America, for a
264 total landfilling costs \$165 US tonne⁻¹ of MSW (Eunomia 2002). The cost of capturing landfill
265 biogases of \$1.50 US tonne⁻¹ was added to represent more current practices. Centralized
266 composting facilities (CCF) were assumed to cost 33 % more in collection and transportation as
267 compared to landfilling because of the double collection required, but at a less frequent interval
268 for MSW other than UOW, and the location of CCF within the urban perimeter. With a capacity
269 of 20×10^3 tonne yr⁻¹ and at 7 % interest rate, CCF can compost UOW at a cost of \$241 US
270 tonne⁻¹. Community composting centres (CCC) are estimated to treat UOW at a lower cost of

271 \$118 US tonne⁻¹ (Table 7) because of the time volunteered to operate the centre. Community
272 composters with a 11.5 m³ capacity were presumed to cost \$30000 US, if not automated and
273 built of polyethylene (Eco-quartier, 2009). As compared to CCC, HC can cost in the range of
274 \$31 US tonne⁻¹ of capacity and are expected to have a 10 yr life. For HC, no collection and
275 transportation costs are involved and the time required to manage the system is free. Purchasing
276 the composter, promoting the use of HC and training the community, are the only costs
277 amounting to \$42 US tonne⁻¹. The cost of bulking agent is considered negligible when food
278 waste is composted along with yard trimmings.

279

280 **Insert Table 7**

281

282 *Environmental assumptions*

283 For all management methods, UOW generated CH₄. Landfilled UOW contribute to global
284 warming by generating potentially 204 kg CH₄, 500 kg CO₂ and 0.13 kg N₂O tonne⁻¹ wet OW
285 (Barton & Atwater 2002; Pettus 2009; SITA Australia Pty Ltd 2008; US CESLG 2008; Wang et
286 al. 1997), where modern technology can capture and oxidize up to 80 % of the CH₄ generated
287 (SITA Australia Pty Ltd 2008). In addition, the garbage trucks collecting and transporting the
288 MSW generate some 25 kg CO₂ tonne⁻¹ wet UOW (Clean Energy 2007), assuming that UOW
289 constitute 46 % of the volume handled.

290 According to IPCC (2006), CCF generate 4 kg CH₄ and 0.3 kg N₂O tonne⁻¹ wet organic
291 waste and the transportation contribution was presumed increased by 33% as compared to
292 landfilling. This compares favourably with Amlinger et al. (2008) reporting that HC generates
293 0.8 to 2.2 kg CH₄, 139 to 215 kg CO₂ and 0.076 to 0.186 kg N₂O tonne⁻¹ of OW. Although
294 composting is an aerobic process, some CH₄ is formed by anaerobic pockets within the mass,
295 when initiating the process because microbes can consume O₂ faster than it can be transferred. In

296 the present analysis, CH₄, CO₂ and N₂O emissions from CCC and HC were assumed to respect
297 the upper limits found by Amlinger et al. (2008), because of limited data on CCC emissions. The
298 GHG (CO₂, CH₄ and N₂O) emissions were expressed as units of CO₂ equivalent based on the
299 global warming potential (GWP) of CH₄ and N₂O valued at 21 and 310 times that of CO₂,
300 respectively (US EPA 2005).

301

302 **Comparison of the results for the different scenarios**

303 *Economic implications of various scenarios from 2009 to 2025*

304 Considering the Base Sce, the cost of UOW landfilling in the EUG1 will increase from \$12300
305 to \$15100 million US from 2009 to 2025, an increase of 23 % based on 2009 values (Figure 3a).
306 Compared to the base scenario and for 2025, Sce 1 & 2 will increase the cost of handling and
307 treating UOW by 25 & 37 %, respectively, whereas Sce 3 will lower the cost by 49%.
308 Accordingly from 2009 to 2025 and for Sce 1, 2 and 3, the cost of treating UOW will increase
309 from \$15400 to \$18900 million US, \$15100 to \$20700 million US and \$6300 to \$7800 million
310 US, respectively.

311 In the EUG2, the landfilling cost for UOW will increase by 14 % from \$1800 to \$2100
312 million US. Over the same period, Sce 1, 2 and 3 could increase the cost of handling and treating
313 UOW from \$1950 to \$2225 million US, \$2250 to \$2850 million US and \$935 to \$1070 million
314 US, respectively (Figure 3b). In comparison to landfilling, adopting CCF (Sce 2) will increase
315 the cost by 57 % in 2025 while adopting HC (Sce 3) will drop the cost by 41 %.

316 In the EU27, the cost of UOW landfilling is expected to increase from \$14100 to \$17200
317 million US, if the Base Sce is maintained, an increase of 22 % (Figure 3c), whereas Sce 1 & 2
318 will increase this cost by 50 and 67 %, respectively and Sce 3 can drop this cost by 37 %.

319 For Canada, maintaining the Base Sce will increase UOW handling and treatment costs
320 by 29 %, in 2025 as compared to 2009. If Sce-3 is adopted, the cost will increase from \$660 to

321 \$850 million US over the same period. By adopting Sce 3, the cost can be lowered by 34% as
322 compared to the Base Sce (Figure 3d). Accordingly, Sce 3 appears to be the most feasible option
323 from an economic point of view for the on-site treatment of UOW over the next 15 years.

324

325 **Insert Figure 3**

326

327 ***GHG (CO₂, N₂O and CH₄) emissions under the various scenarios for 2025***

328 From 2009 to 2025, UOW management can increase GHG emission if the Base Sce, mainly
329 landfilling, is maintained along with its fossil fuels consumption for waste collecting and
330 transportation. Considering Base Sce for the EUG1, GHG emission from landfilled UOW will
331 grow from 37 to 45 million tonne CO₂-eq (Figure 4a), whereas Sce 1 and 3 could emit 34 and 27
332 million tonne CO₂-eq, respectively by 2025, representing a drop of 25 and 40 %. The Sce 2 is
333 will increases emissions by 7 % in 2025, because non CH₄ capture is assumed from 2016
334 onwards with the assumption that less UOW landfilling makes CH₄ capture uneconomical.
335 Similarly for the EUG2 by 2025, the GHG emissions for the Base Sce could increase from 5 to 6
336 million tonne CO₂-eq whereas Sce 1, 2 and 3 could limit GHG emissions to 6, 7 and 3 million
337 tonne CO₂-eq, respectively (Figure 4b). Again with the Base Sce, GHG emissions by the EU27
338 could increase from 42 to 51 million tonne CO₂-eq (Figure 4c) whereas Sce 1, 2 and 3 could
339 limit GHG emissions to 40, 55 and 31 million tonne CO₂-eq. Compared with 2005 emissions
340 (Table 4), the Base Sce, and Sce 1, 2 and 3 will contribute 1.4, 1.1, 1.5 and 0.8 % of the EUG1
341 anthropogenic GHG emissions respectively in 2025.

342 Similarly in Canada, GHG emission from landfilled UOW or the Base Sce is expected to
343 increase from 4 to 5 million tonne CO₂-eq from 2009 to 2025 (Figure 4d), but to drop to 4 and 3
344 million tonne CO₂-eq, if Sce-1 and 3 are adopted and increases to 5 % if Sce 2 is implemented.

345 Therefore, in the EU and Canada, the implementation of Sce-3 will reduce GHG emissions
346 remarkably in upcoming years.

347

348 **Insert Figure 4**

349 ***Land used for UOW landfilling from 2006 to 2025***

350 Landfilling requires 33 ha of tillable land per million tonne of UOW (Adhikari et al. 2009).
351 Accordingly and from 2009 to 2025, maintaining the Base Sce in the EUG1 will waste 32%
352 more land annually for landfilling (Figure 5a) which is equivalent to twice the Luxemburg
353 permanent crop area of 1780 ha (CIA 2009). Similarly, the EUG2 will require 21% more land
354 (Figure 5b) which is equivalent to twice the Malta permanent crop area of 990 ha (CIA 2009). In
355 Canada, maintaining the Base Sce will increase the annual land usage for landfilling by 38 %
356 (Figure 5d).

357

358 **Insert Figure 5**

359

360 By 2025 in the EUG1, implementing Sce 1, 2 and 3 will reduce the annual land
361 requirements for UOW landfilling by 56, 80 and 100 % as compared to the Base Sce, while for
362 the EUG2, land use will be reduced by 15, 80 and 100 %, respectively (Figure 5b). Similarly, in
363 Canada, Sce 1, 2 and 3 will reduce land use by 41, 80 and 100 %, respectively.

364

365 **Proposed CCC and HC in Paris, France and Toronto, Canada**

366 The cities of Paris in France, and Toronto in Canada, are large cities with a respective
367 population density of 3400 and 2500 km⁻² (Table 9) where the residential, commercial and
368 institutional generation of UOW amounts to 0.63 kg person⁻¹ day⁻¹ (OECD 2006/2007). In this
369 projection, it is proposed to use for Paris, 2 CCC km⁻² with 3 - 15 m³ in-vessel composters and

370 438 individual home 400 L composting bins km⁻², and for Toronto, 1 CCC km⁻² with 2 - 15 m³
371 in-vessel composters and 255 individual home 400 L compost bins km⁻² (Table 8).

372

373 **Insert Table 8**

374 Compared to landfilling, CCC and HC can save annual UOW treatment cost by \$25760
375 and \$28900 US km⁻² in Paris while Toronto can benefit from annual savings of \$8150 and
376 \$49570 US km⁻², respectively. Furthermore, GHG emissions will drop drastically and urban air
377 quality can benefit from less garbage collection and transportation (Adhikari et al. 2009).
378 Therefore, the on-site composting of source separated UOW can offer interesting environmental
379 and economic benefits for the years to come. However, the successful implementation of on-site
380 composting offers some challenges.

381 To recycle a high fraction of the UOW, the implementation challenge for CCC and HC
382 are numerous. The first prerequisite is the participation and involvement of waste producers
383 (urban residents), because most need as stimulus, tax incentives or legislative pressures, besides
384 education on the benefits of compost as soil amendment. Finding space for CCC in highly
385 populated cities such as Paris and Toronto is another challenge, although the Montreal City
386 Tournesol Centre owes its success to its location on the edge of the impressive Jeanne Mance
387 park. The compost produced from CCC and not used by the UOW producers will have to be
388 transported to city gardens.

389

390 **Conclusion and recommendations**

391 The production of municipal solid waste (MSW) and its urban organic waste (UOW) fraction is
392 expected to increase exponentially over the next 15 years, as a result of economic growth and
393 urban expansion. If landfilling is maintained as the main treatment option, such growth will
394 further add to already existing waste management issues and resource shortages. To divert the

395 organic fraction from landfills, the European Union (EU) and Canada have promulgated and
396 implemented waste management legislations with emphasis on reduction, reuse and recycling.

397 The economic and environmental impact of promoting community composting centres
398 (CCC) and home composters (HC) to recycle UOW was investigated in this project as an
399 alternative to landfilling. By 2025, such on-site practices could reduce costs and greenhouse gas
400 emissions by 34 to 50 and 40 %, respectively, as compared to maintaining landfilling practices.
401 Furthermore and annually, some 3440 and 330 ha of agricultural land could be saved for the
402 generation of food in the EU and Canada. By eliminating collection, transport and labour costs,
403 HC are an interesting solution to the recycling of UOW. However, the shift of MSW
404 management systems from landfill disposal to resource recovery requires technological input,
405 population participation, compost quality assurance and sufficient urban gardens to divert the
406 mass produced (Burnley 2007; Hargreaves et al. 2008). In this context, the performance of HC is
407 an issue to be addressed for its successful implementation as on-site treatment system.

408

409 **Acknowledgement**

410
411 The authors gratefully acknowledge the financial and all necessary logistics supported by the
412 Regional Council of Brittany, France, Cemagref Rennes, France, and the Natural Science and
413 Engineering Research Council of Canada.

414

415 **References**

- 416 Adhikari, B.K., Barrington, S.F. & Martinez, J. (2009) Urban food waste generation: challenges
417 and opportunities. *Int. J. Environment and Waste Management*, 3(1/2), 4-21.
- 418 Adhikari, B.K., Barington, S. & Martinez, J. (2006) Predicted growth of world urban food waste
419 and methane production. *Waste Management & Research*, 24, 421-433.
- 420 Alamgir, M. & Ahsan, A. (2007) Municipal solid waste and recovery potential: Bangladesh
421 perspective. *Iran J. Environ. Health. Sci. Eng*, 4 (2), 67-76.

422 Al-Yousfi, A.B. (2003) Environmentally sound technologies (EST) for designing and operating
423 solid wastes landfills, United Nations Environment Program (UNEP), Regional Office for
424 West Asia (ROWA), proceedings of international conference on wastes management and
425 pests control, Muscat Municipality-Oman.

426 Amlinger, F., Peyr, S. & Cuhls, C. (2008) Green house gas emissions from composting and
427 mechanical biological treatment. *Waste Management & Research*, 26, 47-60.

428 Antler, S. (2009) Compost grows stronger. The Compost Council of Canada:
429 www.compost.org/compostinggrowsstronger.html (3 June 2009).

430 Bank of Canada (1995-2010) Public information, Bank of Canada, Ottawa, Ontario, Canada:
431 www.bankofcanada.ca (24 January 2010).

432 Barton, P.K. & Atwater, J.W. (2002) Nitrous oxide emissions and the anthropogenic nitrogen in
433 wastewater and solid waste, *Journal of Environmental Engineering*, 128 (2), 137-150.

434 Berg, C.-G. (2005) Separate collection of organic waste-how does it work in Germany? BMU,
435 Bonn University, Germany.

436 Bhide, A.D. (1994) Methane emission from landfills. *Journal of Indian Association for*
437 *Environmental Management*, 21, 1-7.

438 Burnley, S.J. (2007) A review of municipal solid waste composition in the United Kingdom.
439 *Waste Management*, 27, 1274-1285.

440 CCC (2010) Composting regulations and guidelines across Canada, Compost Council of Canada
441 (CCC), Toronto, Ontario, Canada: www.compost.org/AboutCompostingindustry.html (2 Feb
442 2010).

443 CCEM (1989) National contamination site remediation program. Environment Canada, Ottawa,
444 Canada.

445 CIWMB (2008) Technologies and management options for reducing greenhouse gas emissions
446 from landfills. California Integrated Waste Management Board (CIWMB), publication #200-
447 08-001, California, USA.

448 CIA (2009) The world fact book, Central Intelligence Agency, USA:
449 www.cia.gov/library/publications/the-world-factbook/geos/us.html (4 February 2009).

450 Clean Energy (2007) Refuse truck powered by clean-burning natural gas fuel handles trash
451 collection duties at July 7 live earth New York concert event, USA:
452 www.cleanenergyfuels.com/07/07-6-07.html (13 January 2010).

453 De Baere, I. (2000) Anaerobic digestion of solid waste: state-of-the-art. *Water Science and*
454 *Technology*, 41, 283-290.

455 Demographia (2009) Demographia world urban areas & population projections, 5th
456 comprehensive edition, April 2009 revision: <http://demographia.com/db-worldua2015.pdf> (5
457 June 2009).

458 Dunne, L., Convery, F.J. & Gallagher, L. (2008) An investigation into waste charges in Ireland,
459 with emphasis on public acceptability. *Waste Management*, 28, 2826-2834.

460 ECN (2009) Country report, The European Compost Network, ECN/ORBIT e.V., Postbox 22 29,
461 D-99403 Weimar, Germany. www.compostnetwork.info/index.php?id=36 (3 June 2009).

462 Eco-quartier (2009) Personnel conversation with respect to the purchase of a new non-automated
463 composter built of polyethylene tubing, and site preparation and installation costs, Montréal,
464 Canada.

465 EEA (2009) Diverting waste from landfill, effectiveness of waste-management policies in the
466 European Union, European Environmental Agency, EEA Report NO. 7/2009, ISSN
467 17259177, Konges, Nytorv 6, 1050 Copenhagen K, Denmark.

468 EEA & ETC-WMF (2002) Biodegradable municipal waste management in The Europe. The
469 European Environmental Agency (EEA) and The European Topic Center on Waste and
470 Material Flows (ETC-WMF), Report Number 15/2001, Copenhagen, Denmark.

471 El Asri, R. & Baxter, D. (2004) Process control in municipal solid waste incinerators: survey and
472 assessment. *Waste Management & Research*, 22, 177-185.

473 Environment Canada (2009) National inventory report: Greenhouse gas sources and sinks in
474 Canada 1990 – 2007, National Inventory Report. Cat. No.: En81-4/2007E.

475 Eunomia (2002) Cost for the municipal waste management in the EU, final report to directorate
476 general environment, The European Commission. The Eunomia research & consulting on
477 behalf of ECOTEC research and consulting: [www.edis.sk/ekes/the Eucostwaste.pdf](http://www.edis.sk/ekes/the_Eucostwaste.pdf) (26
478 November 2008).

479 European Biomass Industry Association (2006) Anaerobic digestion.
480 <http://www.eubia.org/108.0.html> (22 February 2010).

481 Eurostat (2009) Structural indicators, environment, municipal waste generated and municipal
482 waste by type of treatment, European Commission, Eurostat:
483 http://epp.eurostat.ec.europa.eu/portal/page/portal/structural_indicators/environment (1Feb.
484 2010)

485 Eurostat (2008) New country profile interface, The Euro indicators: [http://epp.the Eurostat.ec.the
486 Europa.the EU](http://epp.the Eurostat.ec.the Europa.the EU) (31 October 2008).

487 Gervais, C. (2002) An overview of UK waste and resource management policies. Royal Society
488 for Nature Conservation, Forum for the future, London, England.

489 Government of India (2008) Ministry of non-conventional energy sources, energy recovery from
490 wastes, CGO complex, Lodi road, New Delhi – 110 003, India.

491 Guermoud, N., Ouadjnia, F., Abdelmalek, F., Taleb, F. & addou, A. (2008) Municipal solid
492 waste in Mostaganem city (Western Africa). *Waste Management*, 29(2), 896-902.

493 Harjula, H., De Tilly, S. & Stutz, J. (2001) Waste trends and outlook. In: Zacarias-Farah, A. &
494 Geyer-Allely, E. (2003) Household consumption patterns in OECD countries: trends and
495 figures. *Journal of Cleaner Production*, 11, 819-827.

496 Hazra, T. & Goel, S. (2009) Solid waste management in Kolkata, India: Practices and challenges.
497 *Waste Management*, 29(1), 470-478.

498 IPCC (2006) Biological treatment of solid waste, volume 5: waste. Intergovernmental Panel on
499 Climate Change Guidelines for National Greenhouse Gas Inventories.

500 Kelleher, M. (2007) Anaerobic digestion outlook for MSW streams. *BioCycle*, 48(8), 51.

501 Kumar, S., Bhattacharya, J.K., Vaidya, A.N., Chakrabarti, T., Devotta, S. & Akolkar, A.B.
502 (2009) Assessment of the status of municipal solid waste management in metro cities, state
503 cities, class I cities, and class II towns in India: An insight. *Waste Management*, 29(2), 883-
504 895.

505 Kanbour, F. (1997) General status on urban waste management in west Asia, paper delivered at
506 the United Nations Environment Programme (UNEP) regional workshop on urban waste
507 management in west Asia, Bahrain, 23 – 27 November.

508 Landfill Directive (1999) Council directive 1999/31/EC of 26 April 1999 on the landfill of waste.
509 Official Journal of the European Communities, I. 182/1, Office for official publications of the
510 European Communities, 2, rue Mercier, L – 2985 Luxembourg.

511 LAP (2003) The national waste management plan (LAP), Part 1 Policy framework:
512 www.vrom.nl/docs/internationaal/National%20Waste%20Management%20Plan.pdf (1
513 February 2010).

514 Machado, S.L., Carvalho, M.F., Gourc, J.P., Vilar, O.M. & do Nascimento, J.C.F. (2009)
515 Methane generation in tropical landfills: simplified methods and field results. *Waste*
516 *Management*, 29(1), 153-161.

517 Marton, C. & Alwast, H. (2002) Report: Operational experiences and legal aspects of co-
518 combustion in Germany. *Waste Management & Research*, 20, 476-483.

519 Messineo, A. & Panno, D. (2008) Municipal waste management in Sicily: Practices and
520 challenges. *Waste Management*, 28, 1201-1208.

521 MEEDDM (2009) La politique des dechets 2009-2012. Ministere de l'Ecologie, de l'Energie, du
522 Developpement durable et de la Mer, Paris, France.

523 Moghadam, M.R., Mokhtarani, N. & Mokhtarani, B. (2009) Municipal solid waste management
524 in Rasht City, Iran. *Waste Management*, 29(1), 485-489.

525 Nash, C. (1992) Backyard composting: the first step in organic waste management. *Resource*
526 *Recycling*, May 1992.

527 OECD (2006-2007) *OECD Environmental Data; Compendium 2006/2007, Waste.*
528 *Environmental Performance and Information Division OECD, Environmental Directorate.*
529 *Working Group on Environmental Information and outlooks, Paris, France.*

530 Official Journal (2000) Commission Decision 2000/738/EC of 17 November 2000 concerning a
531 questionnaire for member states reports on the implementation of Directive 1999/31/EC on
532 the landfill of waste, 25.11.2000, L298/24. Office for official publications of the European
533 Communities, 2, rue Mercier, L – 2985 Luxembourg.

534 Papadopoulos, A.E., Stylianou, M.A., Michalopoulos, C.P., Moustakas, K.G. & Hapeshis, K.M.
535 (2009) Performance of a new household composter during in-home testing. *Waste*
536 *Management*, 29(1), 204-213.

537 Pettus, A. (2009) Methane: Tapping the untapped potential, Clean Air Task Force, 18 Tremont
538 Street, Boston, USA.

539 Parrot, L., Sotamenou, J. & Dia, B.K. (2009) Municipal solid waste management in Africa:
540 strategies and livelihoods in Yaounde, Cameroon. *Waste Management*, 29(2), 986-995.

541 Peavy, H.S., Rowe, D.R. & Tchobanoglous, G. (1985) *Environmental Engineering*
542 McGraw-Hill, Inc. USA. pp 640-644.

543 Rasapoor, M., Nasrabadi, T., Kamali, M. & Hoveidi, H. (2009) The effects of aeration rate on
544 generated compost quality, using aerated static pile method. *Waste Management*, 29 (2), 570-
545 573.

546 Recycle Works Ltd (2010) *Compost bins and composting*, Unit 1, Bee Mill, Ribchester, UK;
547 www.recycleworks.co.uk (28 January 2010).

548 SITA Australia Pty Ltd. (2008) *Methane captured for greenhouse neutrality*, prepared by Hyder
549 consulting Pty Ltd., North Sydney, Australia, Report no. NS04054-R01-04.

550 Statistics Canada (2008) *Waste management industry survey: Business and government sectors*
551 *2006. Catalogue no. 16F0023X: www.statcan.gc.ca/pub/16f0023x/16f0023x2006001-eng.pdf*
552 (1 February 2010).

553 Schnurer, H.L. (2000) *Regulation on waste management- the situation in Germany. Federal*
554 *Ministry of the Environment, Nature Conservation and Nuclear Safety, Bonn, Germany.*

555 Sufian, M. A. & Bala, B.K. (2007) Modeling of urban solid waste management system: The case
556 of Dhaka city. *Waste Management*, 27, 858-868.

557 Turan, N.G., Coruh, S., Akdemir, A. & Ergun, O.N. (2009) Municipal solid waste management
558 strategies in Turkey. *Waste Management*, 29(1), 465-469.

559 Troschinetz, A.M. & Michelcic, J.R. (2009) Sustainable recycling of municipal solid waste in
560 developing countries. *Waste Management*, 29(2), 915-923

561 UN (2007) Urban population development and the environment, United Nations, Department of
562 Economic and Social Affairs, population Division: www.unpopulation.org (23 May 2008).

563 UNPD (2007) World urbanization prospects: The 2007 revision population database, department
564 of economics and social affairs, United Nations Population Division: www.unpopulation.org
565 (21 January 2009).

566 UNSD (2008) Department of economic and social affairs, United Nations Statics Division:
567 <http://unstats.un.org/unsd/snaama/SelectionCountry.asp> (21 January 2009).

568 US CESLG (2008) Landfill methane utilization, United States Clean Energy Strategies for Local
569 Government . USA.

570 US EPA (2005) Emission facts: metrics for expressing greenhouse gas emissions: carbon
571 equivalents and carbon dioxide equivalents, United States Environmental Protection Agency.
572 EPA 420-F-05-002. 1200 Pennsylvania Ave., NW (6406J), Washington, DC, USA:
573 www.epa.gov/oms/climate/420f05002.pdf (15 January 2010)

574 Van der Sloot, H.A. (1996) Present status of waste management in The Netherlands. *Waste
575 Management*, 16 (5/6), 375-383.

576 Vehlow, J., Bergfeldt, B., Visser, R. & Wilen, C. (2007) The European Union waste
577 management strategy and the importance of biogenic waste. *Journal of Material Cycles and
578 Waste Management*, 9, 130-139.

579 Wang, Y., Odle, W. S., Eleazer, W. E. & Barlaz (1997) Methane potential of food waste and
580 anaerobic toxicity of leachate produced during food waste decomposition. *Waste
581 Management & Research*, 15, 149- 167.

582 Wagner, T. & Arnold, P. (2008) A new model for solid waste management: an analysis of the
583 Nova Scotia MSW strategy. *Journal of cleaner Production*, 16, 410-421.

584 WHO (1995) Solid waste management in some countries – environmental dimensions of waste
585 disposal, World Health Organization regional center for environmental health activities
586 (CEHA), P.O. Box 926967, Amman 111 90, Jordan.

587 World Resources (1998 – 99) A Guide to the Global Urban Environment, city level, 1993. Data
588 table 9.3, pp. 278.

589 World Population prospects (2007) United Nations Population Division: <http://esa.un.org/unpp>
590 (6 July 2008).

- 591 Zhang, D., Keat, T.S. & Gersberg, R.M. (2009) A comparison of municipal solid waste
592 management in Berlin and Singapore. Waste Management, doi:
593 10.1016/j.wasman.2009.11.017.
- 594 Zsigraiová, Z., Tavares, G., Semião, V. & Carvalho, M.G. (2005) Municipal solid waste
595 incineration- contribution to sustainable development of energy and environment. Acta
596 Metallurgica Slovaca, 11(4), 450-459.

597

598 **List of symbols and acronyms**

599

600	Base Sce	–	base scenario
601	CCC	–	community composting centres
602	CCF	–	centralized composting facility
603	CH ₄	–	methane
604	CO ₂ -eq	–	carbon dioxide equivalent
605	CO ₂	–	carbon dioxide
606	EU	–	european union
607	EUG1	–	european union group one
608	EUG2	–	european union group two
609	GDP	–	gross domestic product
610	GHG	–	greenhouse gas
611	GMP	–	glass, metal and plastics
612	HC	–	home composter
613	IC & I	–	institutions, commercials and industries
614	LF CH ₄	–	landfill methane
615	LF	–	landfilling
616	MSW	–	municipal solid waste
617	NUP	–	national urban population
618	OW	–	organic waste
619	P&PB	–	paper & paperboard
620	Scce 1	–	scenario one
621	Scce 2	–	scenario two
622	Scce 3	–	scenario three
623	TA CH ₄	–	total anthropogenic methane
624	TA GHG	–	total anthropogenic greenhouse gas
625	TMSWG	–	total municipal solid waste generated
626	T&O	–	textile & others
627	TP	–	total population
628	UOW	–	urban organic waste
629	UOWPR	–	urban organic waste production rate
630	UP	–	urban population

Table 1: Legislation to divert MSW from landfill in some EU countries and provinces of Canada in 2006.

Country	Diversification	Legislation for diversion
Germany	99 %	<ul style="list-style-type: none"> - Waste Disposal Act (1972): closure of uncontrolled landfills and introduction of strict and costly regulations for new and larger landfills. - Waste Avoidance and Management Act (1986) and German Packaging Ordinance (1991): obligatory packaging recovery by producers and users. - Technical Guidelines for Hazardous Wastes (1991) and Solid Household Wastes and Wastes Similar to Household Wastes (1993): more stringent and costly regulations on landfills and incinerators; realisation that organic wastes create most leachate and gas issues. - Ordinance on Biowaste (1998): control for the sanitary treatment of organic wastes. - Renewable Energy Law (2000): subsidizing of biogas production from organic wastes and municipal support for home composting.
The Netherlands	98 %	<ul style="list-style-type: none"> - National Environmental Policy (1988): brought awareness for recycling with slow progress in terms of mass recycled. - Regulation Air Emissions on Incineration (1989): strict and costly regulations imposed on incinerators. - Regulation on the anaerobic digestion of UOW: subsidization of OW to produce electricity.
France	56 %	<ul style="list-style-type: none"> - Directive aimed at MSW (2005): has producers share the cost of waste management. - Action Plan for 2009-2012: aims at increasing recycling, by applying a higher tax on waste management and requiring manufacturers to help finance packaging management. - Incineration: air quality policies improved from 1995 to 2006; early softer policies allowing the incineration of 47% of all MSW to generally produce energy.
United Kingdom	22 %	<ul style="list-style-type: none"> - Hazardous Waste Directive (1991), Packaging Waste Directive (1994), White paper <i>Making Waste Work</i> (1995), Landfill Directive (1999): UK fails to draw 1999 waste management plan. - Waste Strategy for England and Wales (2000): making local authorities responsible for waste management has limited impact on landfilled MSW. - Renewable Obligation Order (2002): sets a 2010 target for the UK to generate 10 % of its electricity from renewable sources with power to be purchased by the Non-Fossil Purchasing Agency (NFPA).
Canada British Columbia	32%	<ul style="list-style-type: none"> - Emission Criteria for Municipal Solid Waste Incinerators (1991): imposed emission standards and gives all incinerators 5 years to comply. - Landfill Criteria for Municipal Solid Wastes (1993) and Guidelines for Environmental Monitoring (1996): controls and sets standards for the management of landfills; landfill operators have to show that they meet performance criteria. - Regional District Solid Waste Management plan Guidelines (1994): each district sets its plan and several have banned the disposal by landfill of recyclable, hazardous and organic materials. - Recycling Regulation - Product Stewardship (revised 2009): industry and the consumer are responsible for the disposal of packaging and containers.
Nova Scotia	41 %	<ul style="list-style-type: none"> - Nova Scotia Environment Act (1994): proposed to develop a Solid Waste Resource Development Strategy with the 55 municipalities, introduced strict regulations on new landfills and required existing landfills to conform by 2006. - Environment Goals and Sustainable Prosperity Act (1996): committed to recycling 50% of all municipal solid wastes; ban on disposal of all containers and packaging, refillable or recyclable beverage containers, and all organics which can be composted.
Ontario	19 %	<ul style="list-style-type: none"> - Waste Diversion Act (2002) giving Board the responsibility to develop and implement waste diversion programs: reviewed by public consultation to recommend path towards zero waste.
Quebec	27 %	<ul style="list-style-type: none"> - Bylaw on landfill tax (2006): a landfill tax of \$10 ton⁻¹. - Bylaw on the disposal by landfill and incineration of municipal solid wastes (2006): revision with strict environmental requirements whereas before, in many instances a landfill liner was not even required. - Law and several bylaws on recovery and recycling (1990): creation of Recy-Quebec to manage the recycling of glass, paper and metal.

Note: regulations on compost quality are excluded. References : Eurostat (2009); Statistics Canada (2008); CCC (2010); MEEDDM (2009); LAP (2003); Zhang et al. (2009); Schnurer (2000); Bergs (2005); van der Sloot (1996); Gervais (2002).

Table 2: Population and waste generation for countries categorized by GDP per capita (US\$) in 2005.

Category	Countries	GDP ^a per capita US\$	TP ^b (‘1000)	UP ^b (‘1000)	UP %	MSW ^c generated million tonne	UOW ^d generated million tonne	MSW ^c landfilled million tonne
EU Group 1 (EUG1) GDP> 25000	Luxembourg	64102	457	378	82.8	0.32	0.14	0.06
	Ireland	39040	4143	2507	60.5	3.07	0.77	1.84
	Denmark	34298	5417	4653	85.9	3.99	1.16	0.21
	Netherlands	34289	16328	13095	80.2	10.19	3.57	0.18
	Austria	33299	8292	5514	66.5	5.13	1.80	0.66
	UK	33125	60245	54040	89.7	35.18	13.13*	22.65
	Belgium	32955	10398	10117	97.3	4.95	1.93	0.38
	Sweden	32750	9038	7619	84.3	4.36	2.12*	0.21
	Finland	32467	5246	3274	62.4	2.49	1.13*	1.48
	France	31846	60991	46780	76.7	33.06	10.58	11.89
	Germany	29913	82652	60667	73.4	46.62	6.53	3.97
	Italy	29189	58646	39645	67.6	31.79	9.22	17.30
	Spain	26792	43397	33285	76.7	25.91	12.69	12.67
Total/Average		34928	365250	281574	77	207	65	74
EU Group 2 (EUG2) GDP <25000	Greece	23386	11100	6704	60.4	4.86	1.40*	4.30
	Slovenia	22294	1999	990	49.5	0.85	0.24*	0.66
	Portugal	21168	10528	6064	57.6	4.70	1.60	2.93
	Czech Republic	20931	10192	7491	73.5	2.95	0.90*	2.13
	Cyprus	19724	836	579	69.3	0.62	0.11*	0.55
	Malta	19239	403	377	93.6	0.25	0.05*	0.22
	Hungary	18257	10086	6687	66.3	4.64	1.35	3.85
	Slovakia	15991	5387	3027	56.2	1.56	0.44*	1.23
	Estonia	15990	1344	933	69.4	0.59	0.10*	0.37
	Lithuania	14538	3425	2281	66.6	1.29	0.21*	1.17
	Poland	14156	38196	23491	61.5	9.36	3.08*	8.63
	Latvia	13692	2302	1565	68	0.71	0.15*	0.56
	Bulgaria	9220	7745	5437	70.2	3.68	0.42*	3.14
Romania	9067	21628	11614	53.7	8.15	1.17*	6.40	
Total/Average		16975	125171	77241	65	44	11	36
North Americas	US	41410	299846	242276	80.8	222.90	55.73	121.03
	Canada	33400	32271	25849	80.1	13.40	3.22	8.05 ^e
	Mexico	10689	104266	79555	76.3	36.00	18.36	34.81 ^f

GDP - Gross domestic product; UP - Urban population; TP - Total population; MSW - Municipal solid waste; OW - Organic waste; ^aUN (2007); ^bWorld Population Prospects (2007); ^cEurostat (2008); ^dOECD (2006/2007); *Estimated from Adhikari et al. (2006) with 2005 UOW and GDP data; ^e2004; ^f2006; MSW generated and landfilled data for North America from OECD (2006/2007).

Table 3: Municipal solid waste (MSW) composition for the EU and North America in 2005.

Country	TMSWG million tonne	OW %	P& PB %	T&O %	GMP %	OW million tonne	P&PB million tonne	T&O million tonne	GMP million tonne
Luxembourg	0.32	45	22	16	16.8	0.14	0.07	0.05	0.05
Ireland	3.07	25	31	23	20	0.77	0.95	0.71	0.61
Denmark	3.99	29	27	32	11.8	1.16	1.08	1.28	0.47
Netherlands	10.19	35	26	12	27	3.57	2.65	1.22	2.75
Austria	5.13	35	22	19	24	1.80	1.13	0.98	1.23
Belgium	4.95	39	17	29	15	1.93	0.84	1.44	0.74
France	33.06	32	20	26	22	10.58	6.61	8.59	7.27
Germany	46.62	14	34	12	39	6.53	15.85	5.59	18.18
Italy	31.79	29	28	22	20	9.22	8.90	6.99	6.36
Spain	25.91	49	21	7	24	12.69	5.44	1.81	6.22
Hungary	4.64	29	15	35	21	1.35	0.70	1.62	0.97
Portugal	4.70	34	21	23	22	1.60	0.99	1.08	1.03
Slovakia	1.56	38	13	31	18	0.59	0.20	0.48	0.28
USA	222.90	25	34	16	25	55.73	75.79	35.66	55.73
Canada	13.40	24	47	8	22	3.22	6.30	1.07	2.95
Mexico	36.00	51	15	16	15	18.36	5.40	5.76	5.40

Source: OECD (2006/2007); Eurostat (2008)

TMSWG - Total municipal solid waste generated; OW - Organic waste;

P&PB - Paper and cardboard; T&O - Textile and others; GMP - Glass, metal and plastics.

Table 4: Composting of urban organic waste (UOW) in some of the EU countries and Canada.

Country	Composting	Composting	Composting	Composting
	million tonne	%	million tonne	%
	1995		2005	
Luxembourg	0.02	7.04	0.06 ^b	19.28 ^b
Ireland	-	-	0.03 ^a	1.25 ^a
Denmark	0.32	10.69	0.55 ^b	15.28 ^b
Netherlands	2.01	23.72	2.38 ^c	23.49 ^c
Austria	0.94	26.56	2.06 ^c	44.7 ^c
Belgium	0.27	5.95	1.05 ^b	22.76 ^b
France	2.53	9.14	4.87	14.33
Germany	5.59	10.96	8.30 ^c	17.14 ^c
Italy	0.12	0.45	10.55	33.30
Spain	2.09	10.38	7.43 ^c	32.69 ^c
Slovenia	0.00	0.00	0.01 ^a	1.15 ^a
Portugal	0.50	12.90	0.31	6.26
Hungary	-	-	0.05 ^b	1.07 ^b
Canada	-	-	1.66 ^c	12.50 ^c

Source: OECD (2006/2007); Eurostat (2008)

UOW - Urban organic waste

^a2002; ^b2003; ^c2004.

Table 5: Municipal Solid Waste (MSW) composting facilities in some EU countries and Canada.

Country	No of composting facilities	Description	References
France	119	65 mixed MSW and 54 source-separated urban organic waste (UOW) composting facilities.	ECN (2009)
Germany	700-900	Source separated UOW composting is the main focus in Germany for quality assurance.	ECN (2009)
Finland	20	In-vessel composting facilities with a capacity ranging from 5 to 35 million kg.	ECN (2009)
Canada	54	Composting of food waste from institutions, commercials & industries (IC&I) and residential sectors	Antler (2009)

Table 6: Various scenarios for Urban Organic Waste (UOW) treatment strategies.

Scenarios	LF %	CCF %	CCC %	HC %
Base Scenario (Base Sce)	100	0	0	0
Scenario one (Sce 1)				
EUG1	42	57	0	1
EUG2	84	15	0	1
Canada	56	43	0	1
Scenario two (Sce 2)				
Year				
2006	75	25	0	0
2009	50	50	0	0
2016	35	65	0	0
2020	25	75	0	0
2025	20	80	0	0
Scenario three (Sce 3)	0	10	30	60

LF - Landfilling; CCF - Centralized composting facility; CCC - Community composting Centre; HC - Home composters.

Table 7: Cost of landfilling and composting systems for urban organic waste (UOW).

Costs (\$ US tonne ⁻¹ wet UOW)	LF ^a	CCF ^b	CCC	HC
Collection	115	153	0	0
Capital costs				
Land acquisition	2	4	0	0
^h Civil works	19	17	20 ^c	0
Equipments	0	18	64 ^d	0
Site assessment/restoration	2	0	0	0
After care	7	0	0	0
Compost bin	0	0	0	31 ^f
Sub-total	30	39	84	31
Variable costs				
Annual maintenance	0	7	0	0
ⁱ Manpower/operation	18	23	28 ^e	0
Fuel and disposal of rejects	0	19	0	0
Training/Promotion	0	0	6	11 ^g
^j Gas capturing	2	0		
Sub-total	20	49	34	11
Total	165	241	118	42

LF - Landfilling; CCF - Centralized Composting Facility; CCC - Community Composting Center;

HC - Home Composter; UOW - Urban Organic Waste;

^{a&b} values adjusted using an annual inflation rate of 2.5% from 2002 to 2009 and an exchange of \$1.00 US = 0.72 € mid-month average for 2009 (Eunomia 2002; Bank of Canada 1995-2010); for EU, collection cost averaged from a range of \$50 to \$210 US tonne⁻¹; CCF increased by 33 % because of double collection at a lower frequency for MSW other than UOW; other costs for LF and CCF were based on UK and Italian values;

^c Site preparation and installation costs for CCC of \$40000 US (Eco- Quartier 2009);

^d CCC costing \$30000 US for 11.5 m³ total capacity (Eco-Quartier 2009) serving 360 persons generating 0.63 kg UOW person⁻¹day⁻¹ (OECD 2006/2007);

^e Annual management cost of \$5700 US;

^f Wooden 400L composting bin at \$200 US (Recycle works Ltd 2010) serving 4 persons per household generating 0.63 kg UOW person⁻¹day⁻¹ (OECD 2006/2007);

^g Promotion of bins and education of producer (Nash 1992);

^h Civil works for LF including capital investment and interest;

ⁱ Manpower/operation for labour, power and equipment maintenance;

^j Gas capturing cost consisting of vertical extraction wells at \$246 US m⁻¹ drilling and installation cost with a 25 m influence radius, and at \$525 US per wellhead assembly with piping and valves (CIWMB 2008), based on a landfill 15 m deep with an average waste density of 380 kg m⁻³ (Peavy et al. 1985; Bhide 1994).

Table 8: Proposed community and home composting in Paris, France and Toronto, Canada.

Description	Paris		Toronto	
	CCC	HC	CCC	HC
^a TP million(2005):	10.4		^a TP million(2006):	5.671
^a Total area (km ²):	3043		^a Total area (km ²):	2279
% NUP:	22		% NUP:	22
Population km ⁻² :	3400		Population km ⁻² :	2500
^b UOWperson ⁻¹ day ⁻¹ :	0.63		UOWperson ⁻¹ day ⁻¹ :	0.63
	CCC	HC	CCC	HC
^c composting centre km ⁻²	2	-	1	-
In-vessel composters centre ⁻¹	3	-	2	-
In-vessel composter size	6 m long x 1.8 m diameter	-	6 m long x 1.8 m diameter	-
^d Compost bins km ⁻²	-	255	-	438
Capacity of home composter bin ⁻¹	-	400L	-	400L
Composting costs	\$118 US tonne ⁻¹	\$42 US tonne ⁻¹	\$118 US tonne ⁻¹	\$42 US tonne ⁻¹
Total UOW composted yr ⁻¹ km ⁻²	548 tonne	235 tonne	173 tonne	403 tonne
Cost saving compared to landfilling	\$47 US tonne ⁻¹	\$123 US tonne ⁻¹	\$47 US tonne ⁻¹	\$123 US tonne ⁻¹
Cost saving compared to CCF	\$123 US tonne ⁻¹	\$199 US tonne ⁻¹	\$123 US tonne ⁻¹	\$199 US tonne ⁻¹

TP - Total population; NUP - National urban population; UOW - Urban Organic Waste;

CCC - Community composting centre; HC - Home composter;

The density and capacity of CCC and HC corresponds to the waste generated by the population.

^a Demographia (2009);

^b OECD (2006/2007) and assuming same for Toronto;

^c Assuming 70 % and 30 % population live in multi-storey buildings in Paris and Toronto, respectively; ^d Assuming 30 % and 70 % people live in single unit house in Paris and Toronto, respectively with 4 persons household⁻¹.

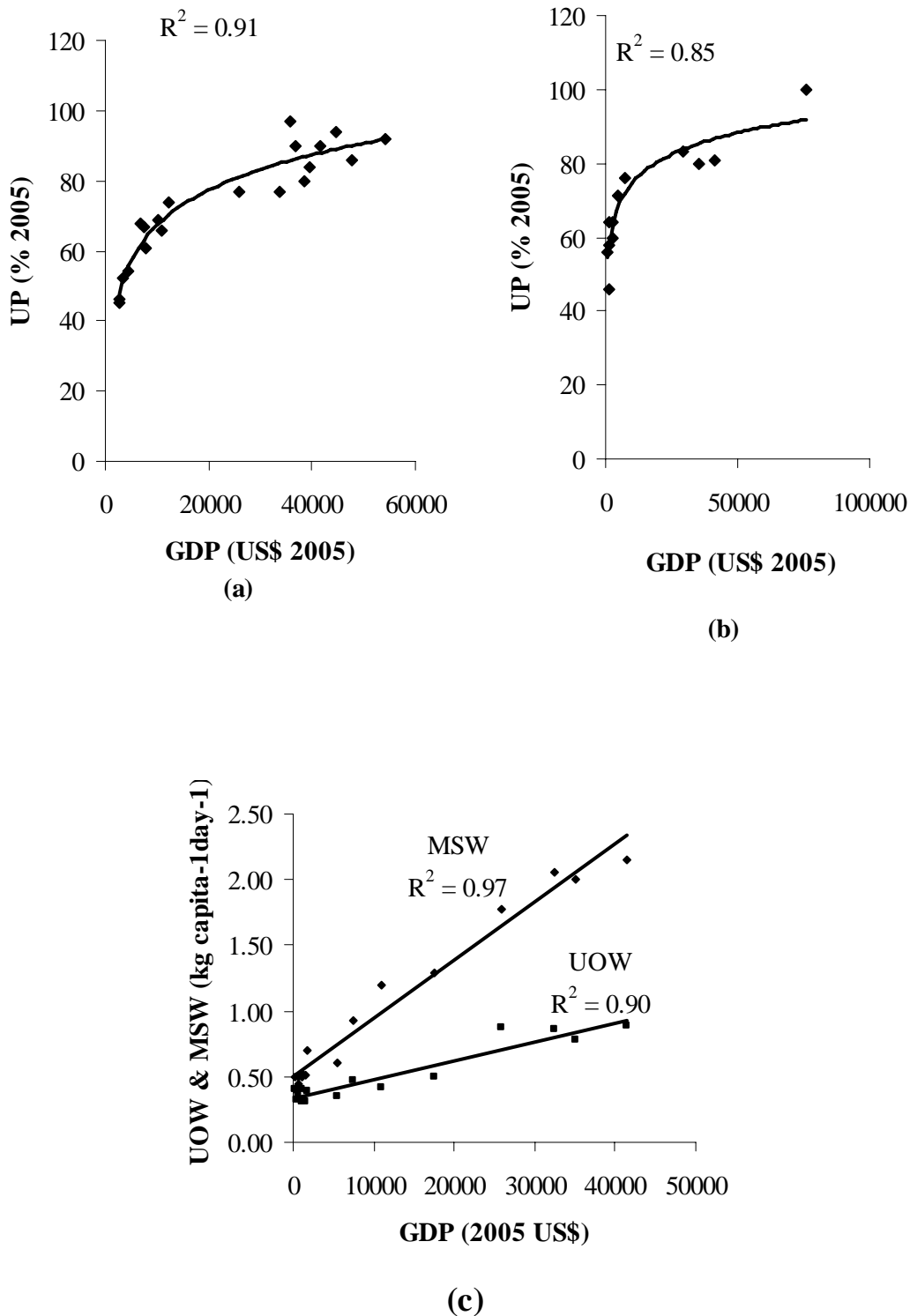
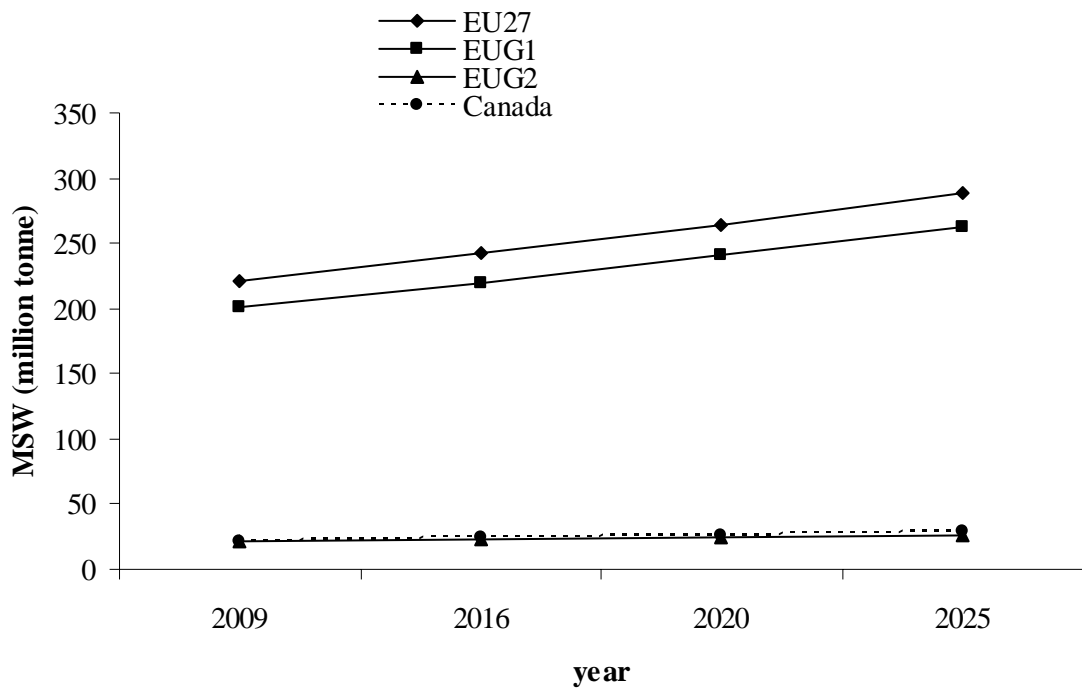
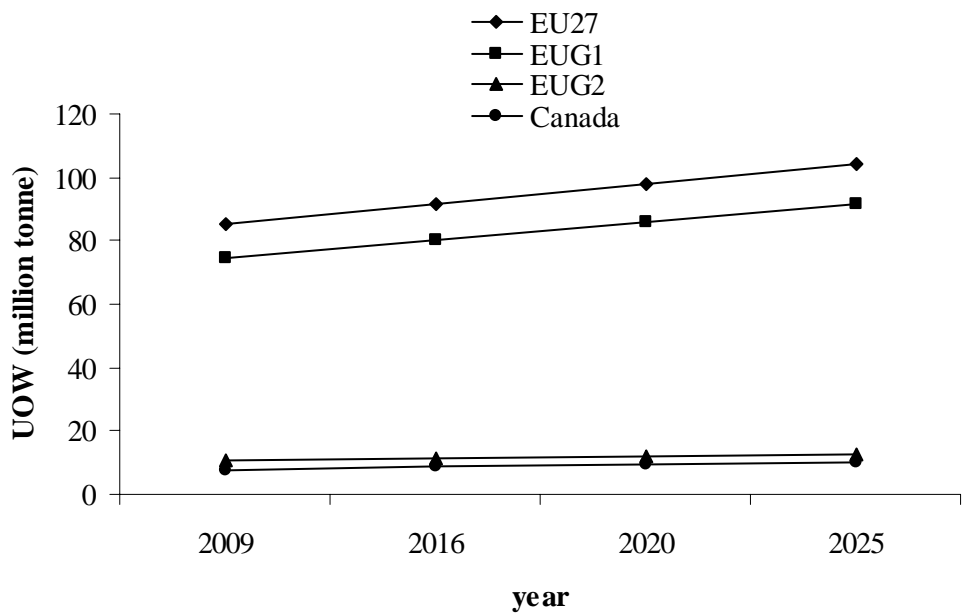


Fig. 1: Correlation of urban population (% UP) with per capita gross domestic product (GDP) according to UNPD (2007) and UNSD (2008) for (a) Europe; and (b) The Americas. (c) correlation of per capita urban organic waste (UOW) and municipal solid waste (MSW) production with per capita gross domestic production (GDP) according to OECD (2006-2007); Eurostat (2008); Sufian & Bala (2007); Government of India (2008); Kanbour (1997); WHO (1995); Al-Yousfi (2003); Alamgir & Ahsan (2007); and World Resources (1998-99).



(a)



(b)

Fig. 2: Urban solid waste generation in the European Union (EU) and Canada on an annual basis (a) Municipal Solid Waste (MSW); and (b) Urban Organic Waste (UOW).

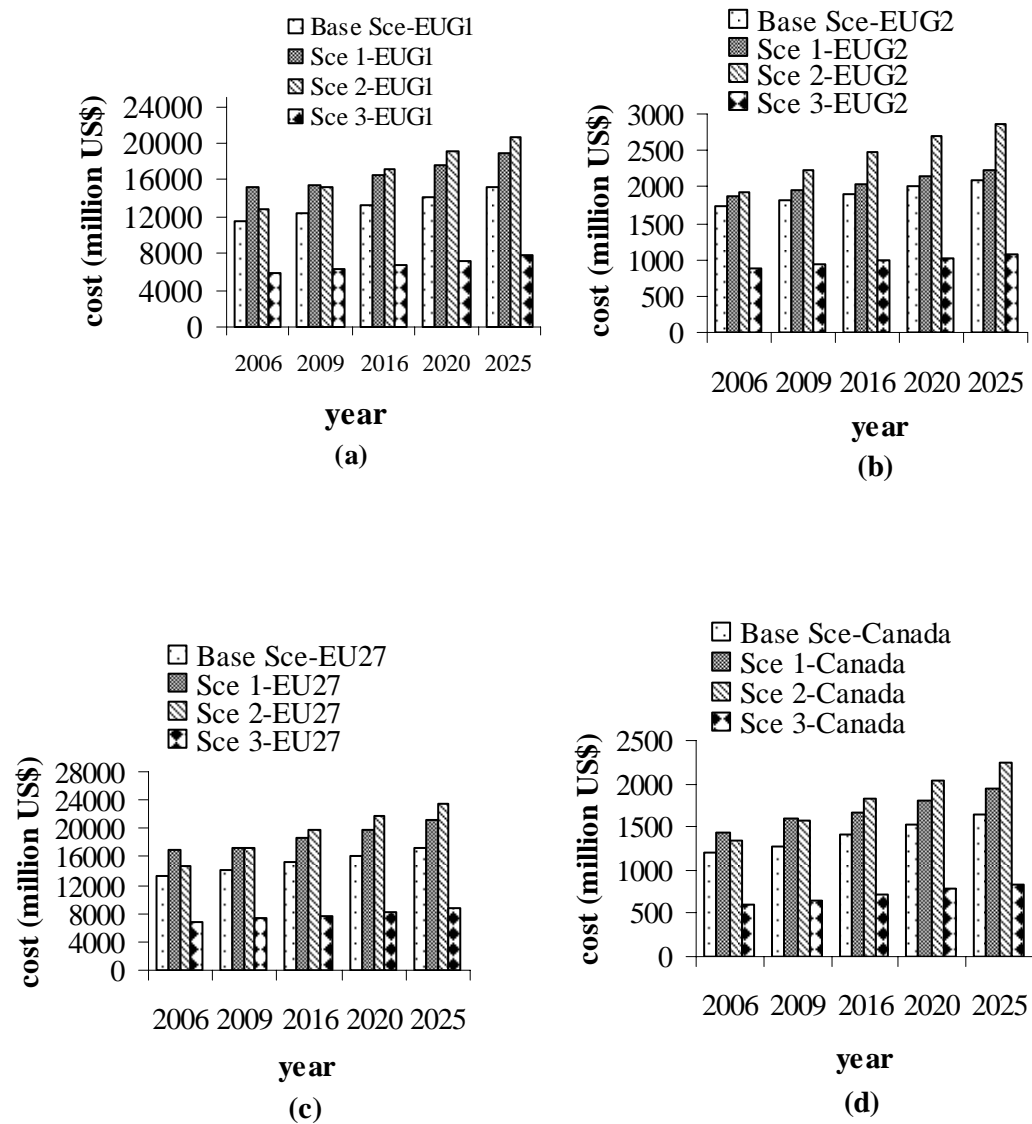


Fig. 3: Cost of urban organic waste (UOW) handling and treatment in consideration of various scenarios for (a) The European Union Group one (EUG1 with annual GDP over \$25000 US); (b) The European Union Group two (EUG2 with annual GDP under \$25000 US); (c) 27 countries of The European Union (EU27 or EUG1 + EUG2); and (d) Canada.

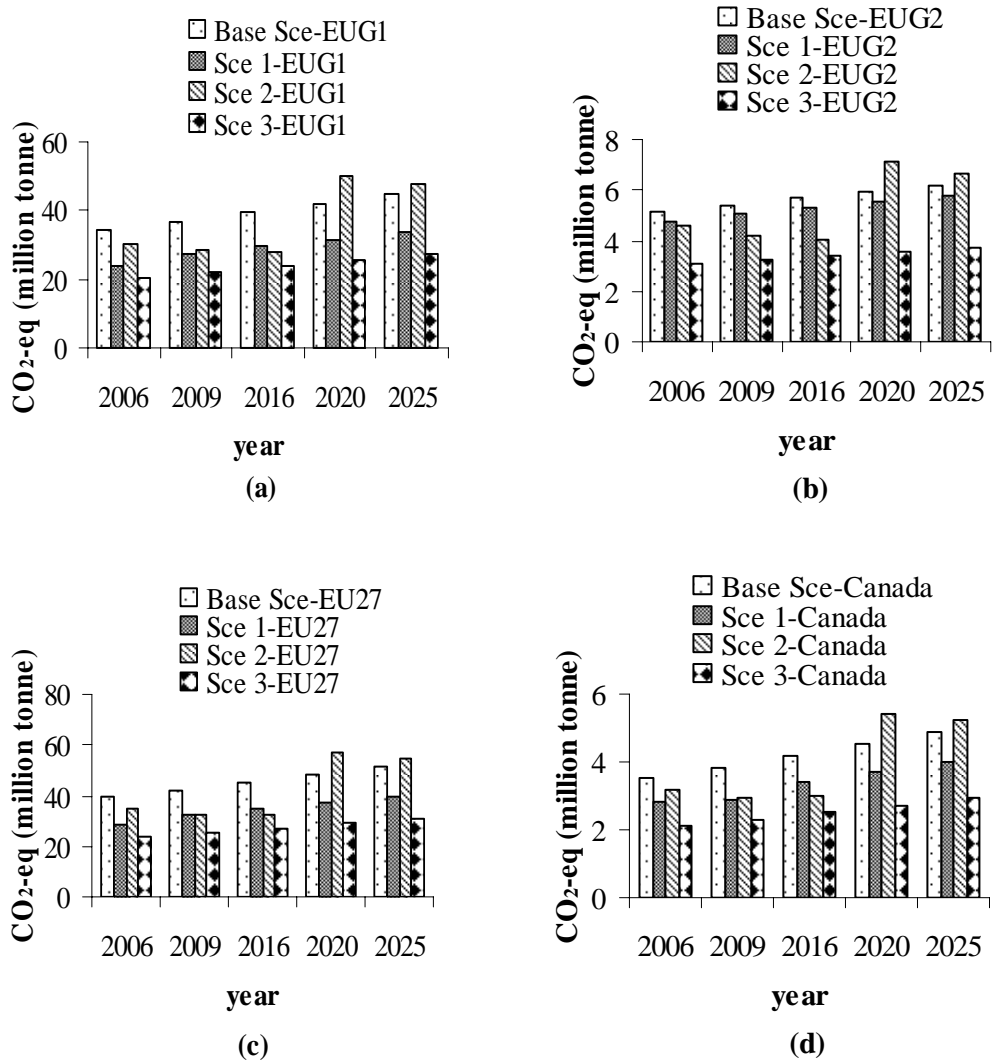
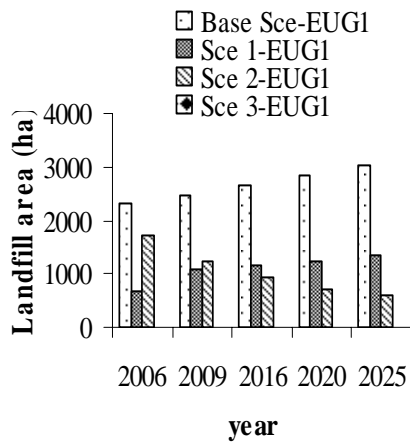
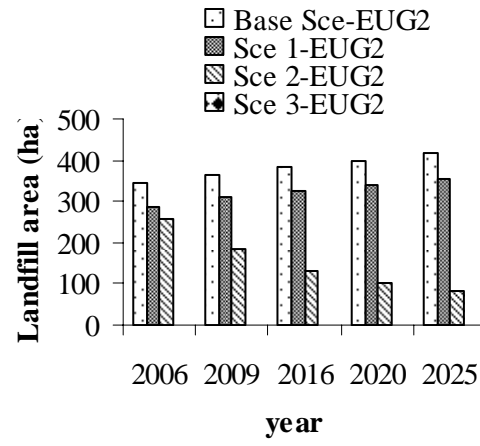


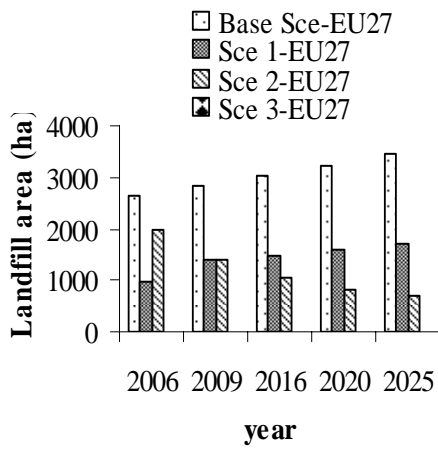
Fig. 4: Computed Greenhouse Gases (CH₄, CO₂ & N₂O) emissions from UOW treatment in consideration of various scenarios for (a) The European Union Group one (EUG1 with annual GDP over \$25000 US); (b) The European Union Group two (EUG2 with annual GDP under \$25000 US); (c) 27 countries of The European Union (EU27 or EUG1 + EUG2); and (d) Canada.



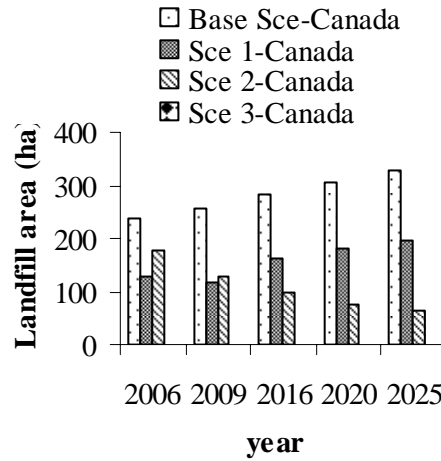
(a)



(b)



(c)



(d)

Fig. 5: Computed landfill area required for urban organic waste (UOW) landfilling in consideration of various scenarios for (a) The European Union Group one (EUG1 with annual GDP over \$25000 US); (b) The European Union Group two (EUG2 with annual GDP under \$25000 US); (c) 27 countries of The European Union (EU27 or EUG1 + EUG2); and (d) Canada.

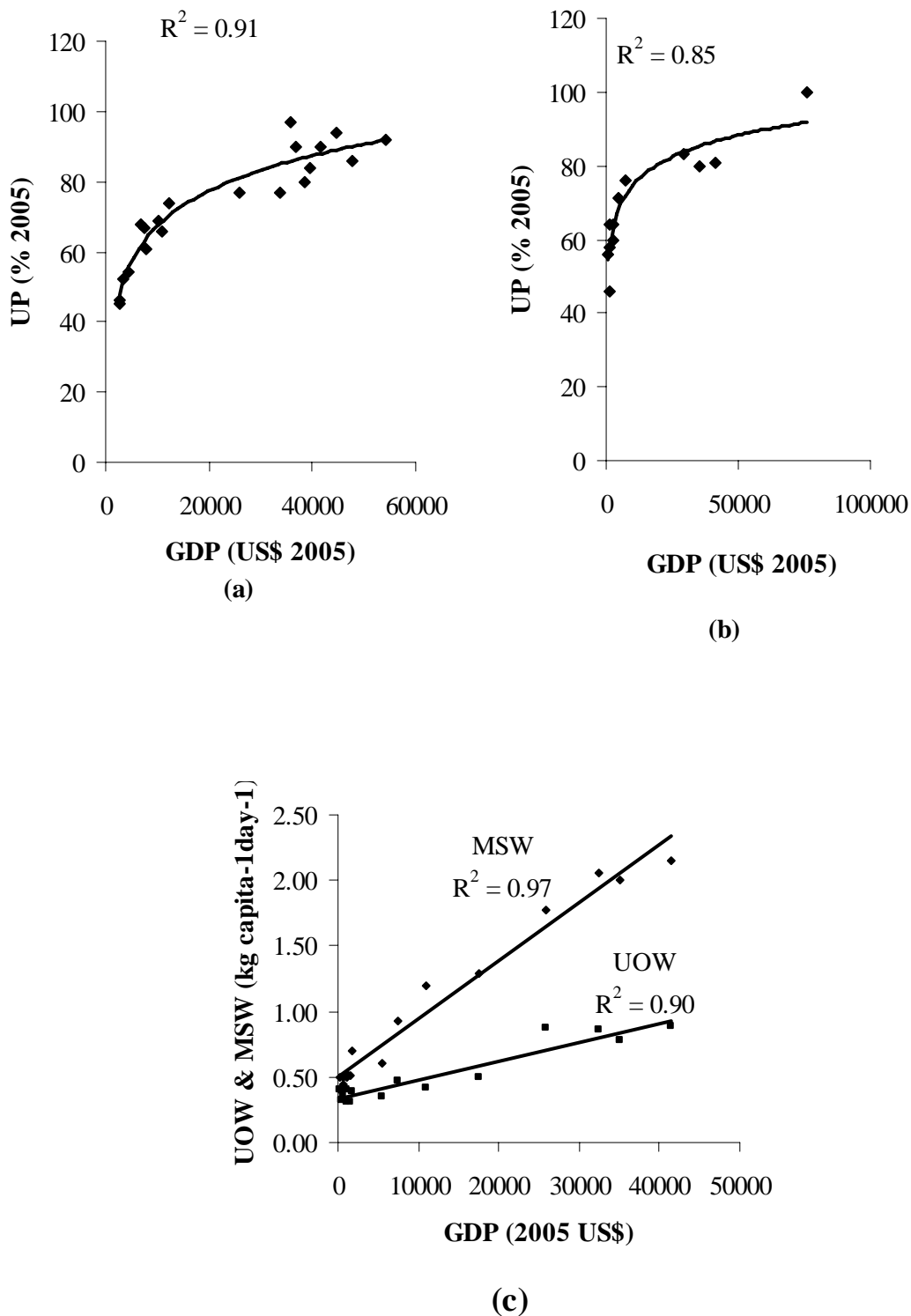
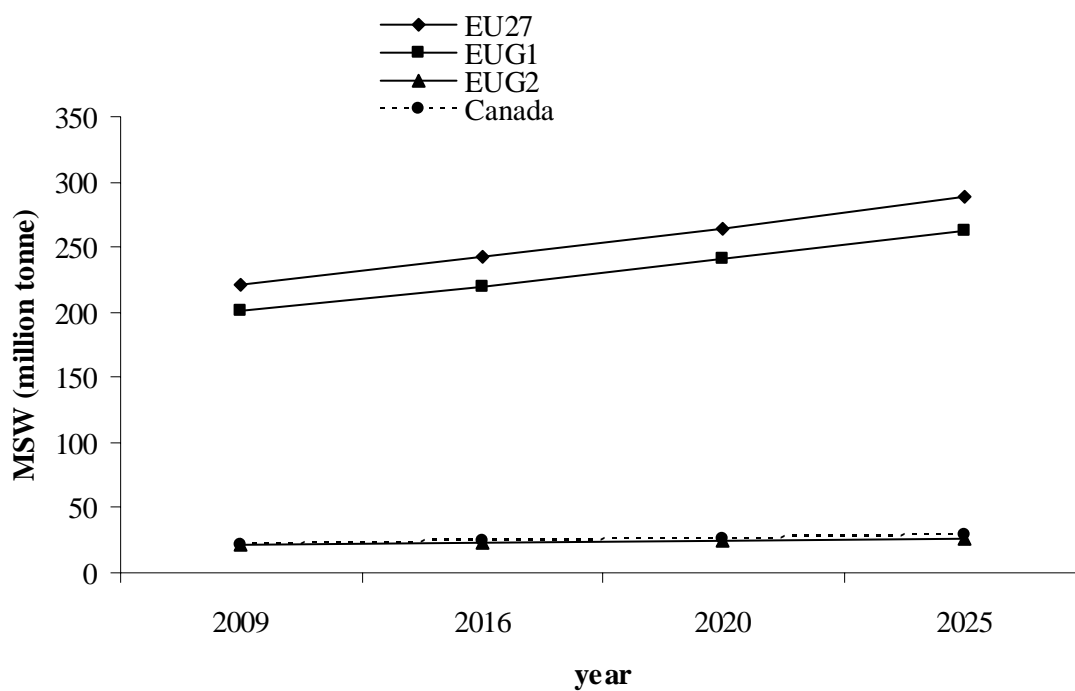
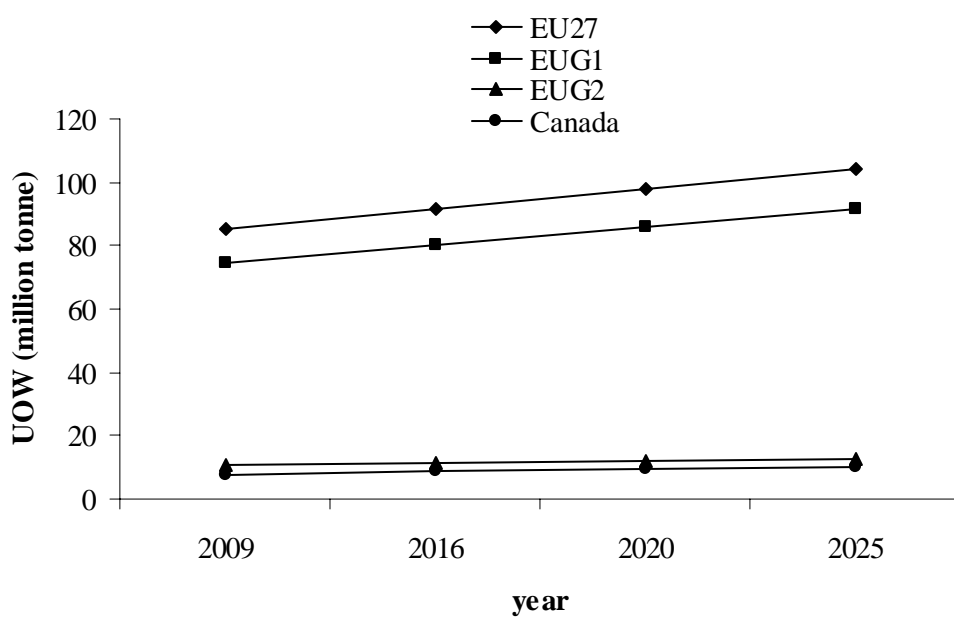


Fig. 1: Correlation of urban population (% UP) with per capita gross domestic product (GDP) according to UNPD (2007) and UNSD (2008) for (a) Europe; and (b) The Americas. (c) correlation of per capita urban organic waste (UOW) and municipal solid waste (MSW) production with per capita gross domestic production (GDP) according to OECD (2006-2007); Eurostat (2008); Sufian & Bala (2007); Government of India (2008); Kanbour (1997); WHO (1995); Al-Yousfi (2003); Alamgir & Ahsan (2007); and World Resources (1998-99).



(a)



(b)

Fig. 2: Urban solid waste generation in the European Union (EU) and Canada on an annual basis (a) Municipal Solid Waste (MSW); and (b) Urban Organic Waste (UOW).

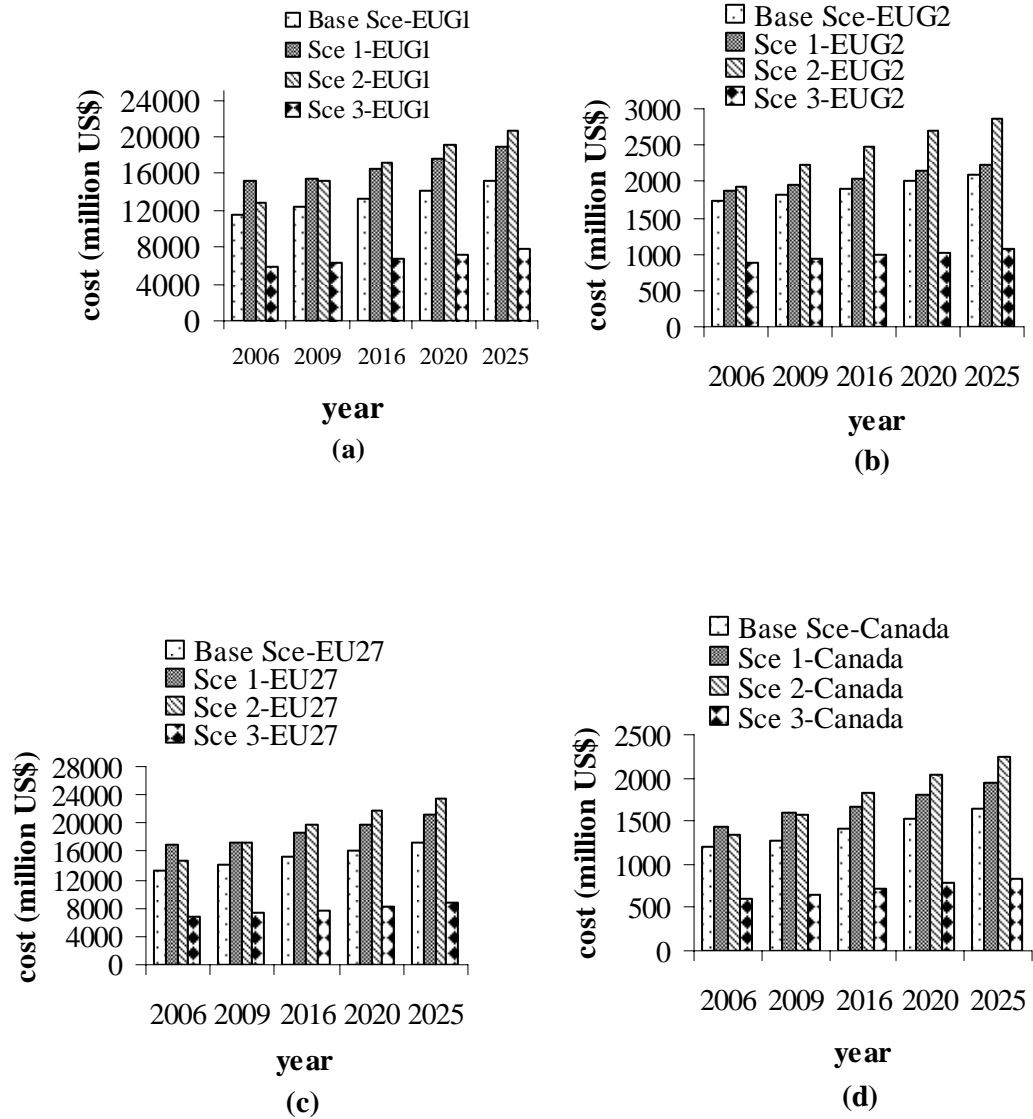


Fig. 3: Cost of urban organic waste (UOW) handling and treatment in consideration of various scenarios for (a) The European Union Group one (EUG1 with annual GDP over \$25000 US); (b) The European Union Group two (EUG2 with annual GDP under \$25000 US); (c) 27 countries of The European Union (EU27 or EUG1 + EUG2); and (d) Canada.

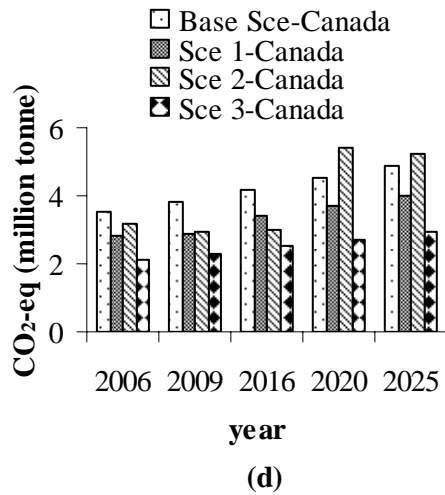
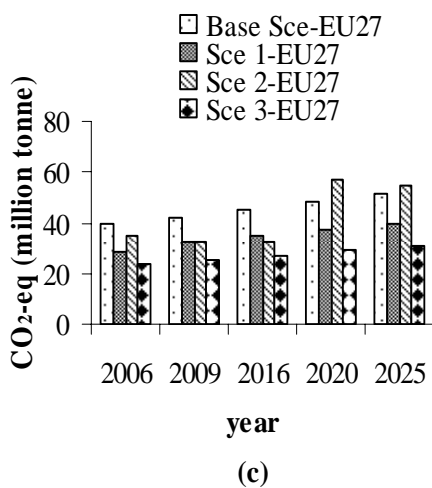
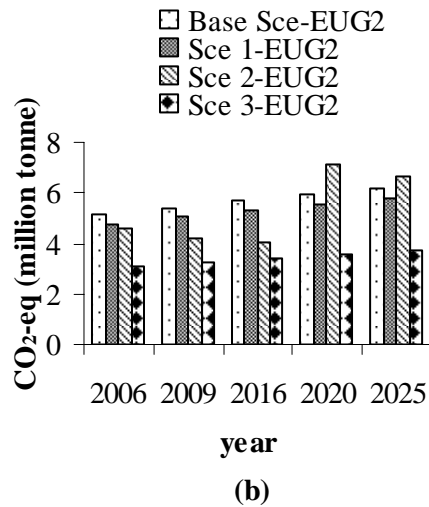
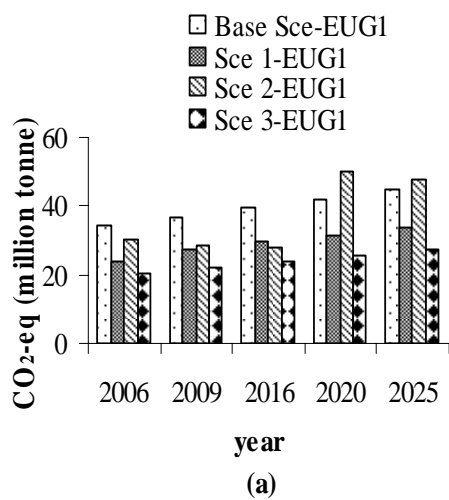
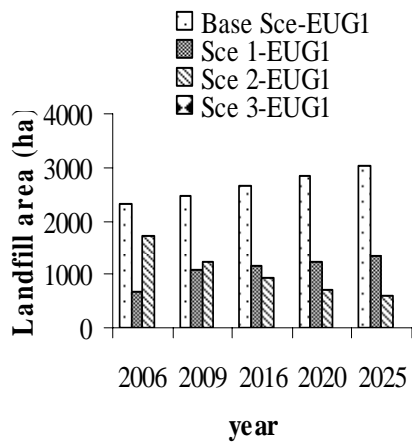
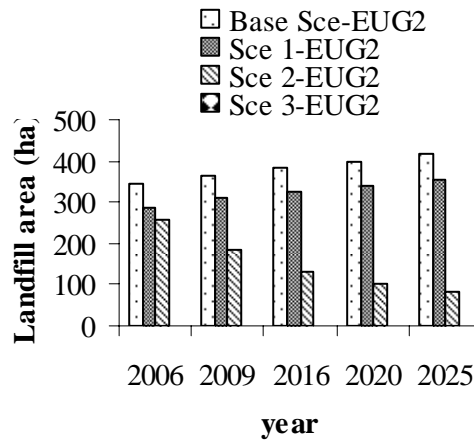


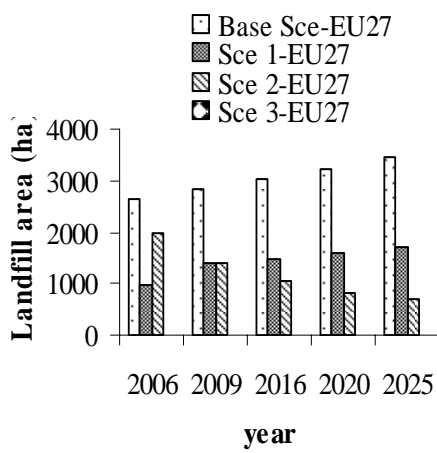
Fig. 4: Computed Greenhouse Gases (CH_4 , CO_2 & N_2O) emissions from UOW treatment in consideration of various scenarios for (a) The European Union Group one (EUG1 with annual GDP over \$25000 US); (b) The European Union Group two (EUG2 with annual GDP under \$25000 US); (c) 27 countries of The European Union (EU27 or EUG1 + EUG2); and (d) Canada.



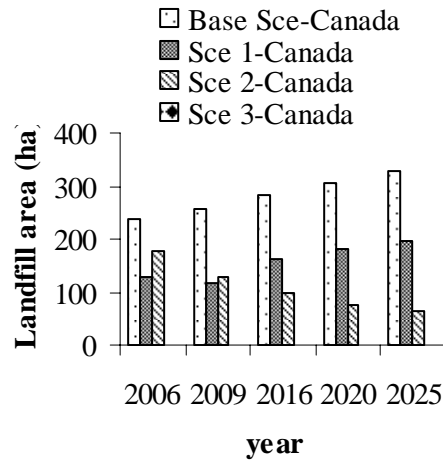
(a)



(b)



(c)



(d)

Fig. 5: Computed landfill area required for urban organic waste (UOW) landfilling in consideration of various scenarios for (a) The European Union Group one (EUG1 with annual GDP over \$25000 US); (b) The European Union Group two (EUG2 with annual GDP under \$25000 US); (c) 27 countries of The European Union (EU27 or EUG1 + EUG2); and (d) Canada.

Table 1: Legislation to divert MSW from landfill in some EU countries and provinces of Canada in 2006.

Country	Diversion	Legislation for diversion
Germany	99 %	<ul style="list-style-type: none"> - Waste Disposal Act (1972): closure of uncontrolled landfills and introduction of strict and costly regulations for new and larger landfills. - Waste Avoidance and Management Act (1986) and German Packaging Ordinance (1991): obligatory packaging recovery by producers and users. - Technical Guidelines for Hazardous Wastes (1991) and Solid Household Wastes and Wastes Similar to Household Wastes (1993): more stringent and costly regulations on landfills and incinerators; realisation that organic wastes create most leachate and gas issues. - Ordinance on Biowaste (1998): control for the sanitary treatment of organic wastes. - Renewable Energy Law (2000): subsidizing of biogas production from organic wastes and municipal support for home composting.
The Netherlands	98 %	<ul style="list-style-type: none"> - National Environmental Policy (1988): brought awareness for recycling with slow progress in terms of mass recycled. - Regulation Air Emissions on Incineration (1989): strict and costly regulations imposed on incinerators. - Regulation on the anaerobic digestion of UOW: subsidization of OW to produce electricity.
France	56 %	<ul style="list-style-type: none"> - Directive aimed at MSW (2005): has producers share the cost of waste management. - Action Plan for 2009-2012: aims at increasing recycling, by applying a higher tax on waste management and requiring manufacturers to help finance packaging management. - Incineration: air quality policies improved from 1995 to 2006; early softer policies allowing the incineration of 47% of all MSW to generally produce energy.
United Kingdom	22 %	<ul style="list-style-type: none"> - Hazardous Waste Directive (1991), Packaging Waste Directive (1994), White paper <i>Making Waste Work</i> (1995), Landfill Directive (1999): UK fails to draw 1999 waste management plan. - Waste Strategy for England and Wales (2000): making local authorities responsible for waste management has limited impact on landfilled MSW. - Renewable Obligation Order (2002): sets a 2010 target for the UK to generate 10 % of its electricity from renewable sources with power to be purchased by the Non-Fossil Purchasing Agency (NFPA).
Canada British Columbia	32%	<ul style="list-style-type: none"> - Emission Criteria for Municipal Solid Waste Incinerators (1991): imposed emission standards and gives all incinerators 5 years to comply. - Landfill Criteria for Municipal Solid Wastes (1993) and Guidelines for Environmental Monitoring (1996): controls and sets standards for the management of landfills; landfill operators have to show that they meet performance criteria. - Regional District Solid Waste Management plan Guidelines (1994): each district sets its plan and several have banned the disposal by landfill of recyclable, hazardous and organic materials. - Recycling Regulation - Product Stewardship (revised 2009): industry and the consumer are responsible for the disposal of packaging and containers.
Nova Scotia	41 %	<ul style="list-style-type: none"> - Nova Scotia Environment Act (1994): proposed to develop a Solid Waste Resource Development Strategy with the 55 municipalities, introduced strict regulations on new landfills and required existing landfills to conform by 2006. - Environment Goals and Sustainable Prosperity Act (1996): committed to recycling 50% of all municipal solid wastes; ban on disposal of all containers and packaging, refillable or recyclable beverage containers, and all organics which can be composted.
Ontario	19 %	<ul style="list-style-type: none"> - Waste Diversion Act (2002) giving Board the responsibility to develop and implement waste diversion programs: reviewed by public consultation to recommend path towards zero waste.
Quebec	27 %	<ul style="list-style-type: none"> - Bylaw on landfill tax (2006): a landfill tax of \$10 ton⁻¹. - Bylaw on the disposal by landfill and incineration of municipal solid wastes (2006): revision with strict environmental requirements whereas before, in many instances a landfill liner was not even required. - Law and several bylaws on recovery and recycling (1990): creation of Recy-Quebec to manage the recycling of glass, paper and metal.

Note: regulations on compost quality are excluded. References : Eurostat (2009); Statistics Canada (2008); CCC (2010); MEEDDM (2009); LAP (2003); Zhang et al. (2009); Schnurer (2000); Bergs (2005); van der Sloot (1996); Gervais (2002).

Table 2: Population and waste generation for countries categorized by GDP per capita (US\$) in 2005.

Category	Countries	GDP ^a per capita US\$	TP ^b (‘1000)	UP ^b (‘1000)	UP %	MSW ^c generated million tonne	UOW ^d generated million tonne	MSW ^c landfilled million tonne
EU Group 1 (EUG1) GDP> 25000	Luxembourg	64102	457	378	82.8	0.32	0.14	0.06
	Ireland	39040	4143	2507	60.5	3.07	0.77	1.84
	Denmark	34298	5417	4653	85.9	3.99	1.16	0.21
	Netherlands	34289	16328	13095	80.2	10.19	3.57	0.18
	Austria	33299	8292	5514	66.5	5.13	1.80	0.66
	UK	33125	60245	54040	89.7	35.18	13.13*	22.65
	Belgium	32955	10398	10117	97.3	4.95	1.93	0.38
	Sweden	32750	9038	7619	84.3	4.36	2.12*	0.21
	Finland	32467	5246	3274	62.4	2.49	1.13*	1.48
	France	31846	60991	46780	76.7	33.06	10.58	11.89
	Germany	29913	82652	60667	73.4	46.62	6.53	3.97
	Italy	29189	58646	39645	67.6	31.79	9.22	17.30
	Spain	26792	43397	33285	76.7	25.91	12.69	12.67
Total/Average		34928	365250	281574	77	207	65	74
EU Group 2 (EUG2) GDP <25000	Greece	23386	11100	6704	60.4	4.86	1.40*	4.30
	Slovenia	22294	1999	990	49.5	0.85	0.24*	0.66
	Portugal	21168	10528	6064	57.6	4.70	1.60	2.93
	Czech Republic	20931	10192	7491	73.5	2.95	0.90*	2.13
	Cyprus	19724	836	579	69.3	0.62	0.11*	0.55
	Malta	19239	403	377	93.6	0.25	0.05*	0.22
	Hungary	18257	10086	6687	66.3	4.64	1.35	3.85
	Slovakia	15991	5387	3027	56.2	1.56	0.44*	1.23
	Estonia	15990	1344	933	69.4	0.59	0.10*	0.37
	Lithuania	14538	3425	2281	66.6	1.29	0.21*	1.17
	Poland	14156	38196	23491	61.5	9.36	3.08*	8.63
	Latvia	13692	2302	1565	68	0.71	0.15*	0.56
	Bulgaria	9220	7745	5437	70.2	3.68	0.42*	3.14
Romania	9067	21628	11614	53.7	8.15	1.17*	6.40	
Total/Average		16975	125171	77241	65	44	11	36
North Americas	US	41410	299846	242276	80.8	222.90	55.73	121.03
	Canada	33400	32271	25849	80.1	13.40	3.22	8.05 ^e
	Mexico	10689	104266	79555	76.3	36.00	18.36	34.81 ^f

GDP - Gross domestic product; UP - Urban population; TP - Total population; MSW - Municipal solid waste; OW - Organic waste; ^aUN (2007); ^bWorld Population Prospects (2007); ^cEurostat (2008); ^dOECD (2006/2007); *Estimated from Adhikari et al. (2006) with 2005 UOW and GDP data; ^e2004; ^f2006; MSW generated and landfilled data for North America from OECD (2006/2007).

Table 3: Municipal solid waste (MSW) composition for the EU and North America in 2005.

Country	TMSWG million tonne	OW %	P& PB %	T&O %	GMP %	OW million tonne	P&PB million tonne	T&O million tonne	GMP million tonne
Luxembourg	0.32	45	22	16	16.8	0.14	0.07	0.05	0.05
Ireland	3.07	25	31	23	20	0.77	0.95	0.71	0.61
Denmark	3.99	29	27	32	11.8	1.16	1.08	1.28	0.47
Netherlands	10.19	35	26	12	27	3.57	2.65	1.22	2.75
Austria	5.13	35	22	19	24	1.80	1.13	0.98	1.23
Belgium	4.95	39	17	29	15	1.93	0.84	1.44	0.74
France	33.06	32	20	26	22	10.58	6.61	8.59	7.27
Germany	46.62	14	34	12	39	6.53	15.85	5.59	18.18
Italy	31.79	29	28	22	20	9.22	8.90	6.99	6.36
Spain	25.91	49	21	7	24	12.69	5.44	1.81	6.22
Hungary	4.64	29	15	35	21	1.35	0.70	1.62	0.97
Portugal	4.70	34	21	23	22	1.60	0.99	1.08	1.03
Slovakia	1.56	38	13	31	18	0.59	0.20	0.48	0.28
USA	222.90	25	34	16	25	55.73	75.79	35.66	55.73
Canada	13.40	24	47	8	22	3.22	6.30	1.07	2.95
Mexico	36.00	51	15	16	15	18.36	5.40	5.76	5.40

Source: OECD (2006/2007); Eurostat (2008)

TMSWG - Total municipal solid waste generated; OW - Organic waste;

P&PB - Paper and cardboard; T&O - Textile and others; GMP - Glass, metal and plastics.

Table 4: Composting of urban organic waste (UOW) in some of the EU countries and Canada.

Country	Composting	Composting	Composting	Composting
	million tonne	%	million tonne	%
	1995		2005	
Luxembourg	0.02	7.04	0.06 ^b	19.28 ^b
Ireland	-	-	0.03 ^a	1.25 ^a
Denmark	0.32	10.69	0.55 ^b	15.28 ^b
Netherlands	2.01	23.72	2.38 ^c	23.49 ^c
Austria	0.94	26.56	2.06 ^c	44.7 ^c
Belgium	0.27	5.95	1.05 ^b	22.76 ^b
France	2.53	9.14	4.87	14.33
Germany	5.59	10.96	8.30 ^c	17.14 ^c
Italy	0.12	0.45	10.55	33.30
Spain	2.09	10.38	7.43 ^c	32.69 ^c
Slovenia	0.00	0.00	0.01 ^a	1.15 ^a
Portugal	0.50	12.90	0.31	6.26
Hungary	-	-	0.05 ^b	1.07 ^b
Canada	-	-	1.66 ^c	12.50 ^c

Source: OECD (2006/2007); Eurostat (2008)

UOW - Urban organic waste

^a2002; ^b2003; ^c2004.

Table 5: Municipal Solid Waste (MSW) composting facilities in some EU countries and Canada.

Country	No of composting facilities	Description	References
France	119	65 mixed MSW and 54 source-separated urban organic waste (UOW) composting facilities.	ECN (2009)
Germany	700-900	Source separated UOW composting is the main focus in Germany for quality assurance.	ECN (2009)
Finland	20	In-vessel composting facilities with a capacity ranging from 5 to 35 million kg.	ECN (2009)
Canada	54	Composting of food waste from institutions, commercials & industries (IC&I) and residential sectors	Antler (2009)

Table 6: Various scenarios for Urban Organic Waste (UOW) treatment strategies.

Scenarios	LF %	CCF %	CCC %	HC %
Base Scenario (Base Sce)	100	0	0	0
Scenario one (Sce 1)				
EUG1	42	57	0	1
EUG2	84	15	0	1
Canada	56	43	0	1
Scenario two (Sce 2)				
Year				
2006	75	25	0	0
2009	50	50	0	0
2016	35	65	0	0
2020	25	75	0	0
2025	20	80	0	0
Scenario three (Sce 3)	0	10	30	60

LF - Landfilling; CCF - Centralized composting facility; CCC - Community composting Centre; HC - Home composters.

Table 7: Cost of landfilling and composting systems for urban organic waste (UOW).

Costs (\$ US tonne ⁻¹ wet UOW)	LF ^a	CCF ^b	CCC	HC
Collection	115	153	0	0
Capital costs				
Land acquisition	2	4	0	0
^h Civil works	19	17	20 ^c	0
Equipments	0	18	64 ^d	0
Site assessment/restoration	2	0	0	0
After care	7	0	0	0
Compost bin	0	0	0	31 ^f
Sub-total	30	39	84	31
Variable costs				
Annual maintenance	0	7	0	0
ⁱ Manpower/operation	18	23	28 ^e	0
Fuel and disposal of rejects	0	19	0	0
Training/Promotion	0	0	6	11 ^g
^j Gas capturing	2	0		
Sub-total	20	49	34	11
Total	165	241	118	42

LF - Landfilling; CCF - Centralized Composting Facility; CCC - Community Composting Center;

HC - Home Composter; UOW - Urban Organic Waste;

^{a&b} values adjusted using an annual inflation rate of 2.5% from 2002 to 2009 and an exchange of \$1.00 US = 0.72 € mid-month average for 2009 (Eunomia 2002; Bank of Canada 1995-2010); for EU, collection cost averaged from a range of \$50 to \$210 US tonne⁻¹; CCF increased by 33 % because of double collection at a lower frequency for MSW other than UOW; other costs for LF and CCF were based on UK and Italian values;

^c Site preparation and installation costs for CCC of \$40000 US (Eco- Quartier 2009);

^d CCC costing \$30000 US for 11.5 m³ total capacity (Eco-Quartier 2009) serving 360 persons generating 0.63 kg UOW person⁻¹day⁻¹ (OECD 2006/2007);

^e Annual management cost of \$5700 US;

^f Wooden 400L composting bin at \$200 US (Recycle works Ltd 2010) serving 4 persons per household generating 0.63 kg UOW person⁻¹day⁻¹ (OECD 2006/2007);

^g Promotion of bins and education of producer (Nash 1992);

^h Civil works for LF including capital investment and interest;

ⁱ Manpower/operation for labour, power and equipment maintenance;

^j Gas capturing cost consisting of vertical extraction wells at \$246 US m⁻¹ drilling and installation cost with a 25 m influence radius, and at \$525 US per wellhead assembly with piping and valves (CIWMB 2008), based on a landfill 15 m deep with an average waste density of 380 kg m⁻³ (Peavy et al. 1985; Bhide 1994).

Table 8: Proposed community and home composting in Paris, France and Toronto, Canada.

Description	Paris		Toronto	
	^a TP million(2005):	10.4	^a TP million(2006):	5.671
	^a Total area (km ²):	3043	^a Total area (km ²):	2279
	% NUP:	22	% NUP:	22
	Population km ⁻² :	3400	Population km ⁻² :	2500
	^b UOWperson ⁻¹ day ⁻¹ :	0.63	UOWperson ⁻¹ day ⁻¹ :	0.63
	CCC	HC	CCC	HC
^c composting centre km ⁻²	2	-	1	-
In-vessel composters centre ⁻¹	3	-	2	-
In-vessel composter size	6 m long x 1.8 m diameter	-	6 m long x 1.8 m diameter	-
^d Compost bins km ⁻²	-	255	-	438
Capacity of home composter bin ⁻¹	-	400L	-	400L
Composting costs	\$118 US tonne ⁻¹	\$42 US tonne ⁻¹	\$118 US tonne ⁻¹	\$42 US tonne ⁻¹
Total UOW composted yr ⁻¹ km ⁻²	548 tonne	235 tonne	173 tonne	403 tonne
Cost saving compared to landfilling	\$47 US tonne ⁻¹	\$123 US tonne ⁻¹	\$47 US tonne ⁻¹	\$123 US tonne ⁻¹
Cost saving compared to CCF	\$123 US tonne ⁻¹	\$199 US tonne ⁻¹	\$123 US tonne ⁻¹	\$199 US tonne ⁻¹

TP - Total population; NUP - National urban population; UOW - Urban Organic Waste;

CCC - Community composting centre; HC - Home composter;

The density and capacity of CCC and HC corresponds to the waste generated by the population.

^a Demographia (2009);

^b OECD (2006/2007) and assuming same for Toronto;

^c Assuming 70 % and 30 % population live in multi-storey buildings in Paris and Toronto, respectively; ^d Assuming 30 % and 70 % people live in single unit house in Paris and Toronto, respectively with 4 persons household⁻¹.