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► To cite this version:

Thomas Lacoue-Labarthe, Sophie Martin, François Oberhänsli, Jean-Louis Teyssié, Ross Jeffree, et al.. Temperature and pCO₂ effect on the bioaccumulation of radionuclides and trace elements in the eggs of the common cuttlefish, *Sepia officinalis*. *Journal of Experimental Marine Biology and Ecology*, 2012, 413 (4), pp.45-49. 10.1016/j.jembe.2011.11.025 . hal-00659887

HAL Id: hal-00659887

<https://hal.science/hal-00659887>

Submitted on 14 Jan 2012

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1 **Temperature and $p\text{CO}_2$ effect on the bioaccumulation of radionuclides and**
2 **trace elements in the eggs of the common cuttlefish, *Sepia officinalis***

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38 **Abstract.** The increasing release of CO₂ by human activities leads to ocean acidification and
39 global warming. Both those consequences (*i.e.*, increase in seawater pCO₂ and temperature)
40 may drastically affect physiology of marine organisms. The effects of low pH and elevated
41 temperature on the bioaccumulation of radionuclides (²⁴¹Am, ¹³⁴Cs) and trace elements (⁶⁰Co,
42 ⁵⁴Mn or ⁷⁵Se) were studied during the embryonic development of the common cuttlefish *Sepia*
43 *officinalis*. The lowered accumulation of essential ⁶⁰Co and ⁵⁴Mn with decreasing pH was
44 larger at 16°C than at 19°C. Se was not detected in the embryo whereas it penetrated through
45 the eggshell, suggesting that an alternative pathway ensures the supply of this essential metal
46 for the embryo. ²⁴¹Am was totally retained by the eggshell irrespective of pH and temperature.
47 Finally, the amount of Cs found in the peri-vitelline fluid increased with decreasing pH likely
48 because of an enhanced swelling of the cuttlefish egg under elevated CO₂.

49

50 **Keywords:** cephalopod, ocean acidification, ocean warming, eggshell, embryonic
51 development, metal.

52

53 INTRODUCTION

54

55 Climate change due to the increasing atmospheric carbon dioxide (CO₂) concentrations in the
56 atmosphere is now considered to be a major threat to biodiversity as well as to the structure
57 and function of ecosystems (Mc Carthy, 2001). Although all the causes are still subjected to
58 debate (Sharp, 2003), it is widely accepted that human activities such as fossil fuel
59 combustion, cement production and change in land use (Hansen et al., 2007) drove the rising
60 global atmospheric concentration of CO₂ from a pre-industrial value of approximately 280
61 parts per million (ppm) to 384 ppm in 2007 (~387 ppm in 2009) (Anonymous, 2010).
62 Moreover, models project that the CO₂ concentration will double by midcentury (> 500 ppm)
63 and could reach between 730 and 1020 ppm in 2100 (IPCC, 2007). The accumulation of
64 greenhouse gases will have profound consequences among which the global warming.
65 Temperature is expected to increase on average by 3°C at the Earth's surface over the course
66 of this century (IPCC, 2007). Furthermore, the precipitation and evaporation rates that drive
67 the hydrologic cycle as well as the winds regimes will also be affected (Roessig et al., 2004).
68 Regarding the Ocean, similar trends are expected for sea surface temperature due to the
69 warming of the surface mixed layer (Levitus et al., 2005) and the sea level will rise mainly
70 because of the melting ice. Moreover, as the oceans are major carbon sink absorbing about
71 25% of anthropogenic CO₂, the sea surface CO₂ partial pressure ($p\text{CO}_2$) is also expected to
72 increase (Sabine et al., 2004). This causes major shifts in seawater carbonate chemistry and is
73 likely to reduce pH by 0.2-0.4 units, a phenomenon called "ocean acidification", over the
74 course of this century (Caldeira and Wickett, 2005).

75 In the northern Europe, the main scenario for climate change is milder, wetter and stormier
76 winters (IPCC, 2007). Summers could be warmer and dryer, although the climatic effect is
77 expected to be less pronounced than in winters (Jonsson and Jonsson, 2009). Simultaneously,

78 the increasing precipitations will directly and indirectly impact the marine coastal area,
79 causing a decrease of the salinity and amplifying the anthropogenic factors, such as hypoxic
80 events linked to increased agricultural runoff or contamination of the nearshore zones by
81 organic and inorganic contaminants released in the environment (Harley et al., 2006). Thus,
82 the marine biota will have to deal with interactions between increasing seawater temperature,
83 $p\text{CO}_2$, and well-known anthropogenic stressors such as the contamination by heavy metals.

84 In ecotoxicology and risk assessment studies, it is widely accepted that the early life stages
85 are the most sensitive to metallic contamination (*e.g.* Calabrese et al., 1973; Warnau et al.,
86 1996; Oral et al., 2010). Recently, a growing body of works also showed that embryos and
87 larvae were affected by hypercapnia (*e.g.* Kurihara et al., 2008; Dupont et al., 2008; Walther
88 et al., 2010) as their developing systemic acid-base regulation is not yet fully efficient
89 (Pörtner and Farrell, 2008). As the highly productive coastal zones serve as nursery grounds
90 for many marine species of ecological and commercial importance (Dunn et al., 1999;
91 Selleslagh and Amara, 2008), studying interaction between multiple stressors is critical to
92 project the future of coastal ecosystems (Harley et al., 2006; Dupont et al., 2010).

93 In Europe, the common cuttlefish *Sepia officinalis* is seasonally exploited in the English
94 Channel (Royer et al., 2006) when adults come in the shallow waters to mate and spawn their
95 eggs before massive die-offs (Boucaud-Camou and Boismery, 1991). The eggs laid in
96 shallow waters are subject to chronic exposure to various contaminants among which, heavy
97 metals that bioaccumulate in the cuttlefish during its ontogenic development (Miramand et
98 al., 2006; Villanueva and Bustamante, 2006). Experimental studies demonstrated that the
99 cuttlefish embryo is partially protected because of the selective permeability of the eggshell to
100 trace elements, depending on the developmental stage (Bustamante et al., 2002, 2004, 2006a;
101 Lacoue-Labarthe et al., 2008a, 2009a, 2010). Moreover, these shielding properties of the egg
102 capsule against Ag, Cd and Zn penetration and the metal accumulation efficiencies of the

103 embryo seemed to be affected when eggs develop under warmer and hypercapnic conditions
104 (Lacoue-Labarthe et al., 2009b).

105 This study addresses the combined effects of elevated temperature and $p\text{CO}_2$ on the
106 bioaccumulation of selected radionuclides and trace elements in the cuttlefish embryo.
107 Indeed, the development of nuclear facilities and fallout from nuclear weapon testing resulted
108 in the release of several medium- and long-lived radionuclides into aquatic environments such
109 as ^{241}Am , ^{60}Co , ^{134}Cs , ^{54}Mn or ^{75}Se (e.g. Warnau et al., 1999, Ke et al., 1999). Although these
110 radionuclides are generally considered as micropollutants in the oceans, they are known to
111 accumulate in marine organisms (e.g. Miramand and Guary, 1981; Bustamante et al. 2006b,
112 Lacoue-Labarthe et al., 2010; Metian et al., 2009, 2011). In addition, radionuclides are of
113 specific interest given the ecotoxicological concern regarding their stable isotopes. Indeed,
114 Co, Mn and Se are trace elements that have essential functions in physiology but also toxic
115 effect if they are excessively accumulated in organisms (e.g. Rainbow, 2002). In this study,
116 we therefore use radiotracers to investigate the bioaccumulation behaviour of these
117 corresponding stable elements that are present in marine waters (Warnau and Bustamante,
118 2007).

119

120 MATERIALS AND METHODS

121

122 Adult cuttlefish were collected by net-fishing off Monaco in April 2008 and maintained in
123 open-circuit tanks in the IAEA-MEL premises. After mating, the fertilized eggs that were laid
124 by each female were immediately separated to optimise their oxygenation. The eggs ($n = 300$)
125 were placed and kept in six 5-L plastic bottles (50 eggs per bottle; one bottle per treatment;
126 $0.45\mu\text{m}$ and UV-sterilized seawater; constantly aerated closed-circuit; salinity 38 p.s.u.;
127 light/dark cycle: 12h/12h) during the full development time in controlled conditions of
128 temperature and pH in a crossed (2 temperatures x 3 pH levels) experimental design. The pH
129 and temperature values were chosen consistent with those of realistic modelled scenarios of
130 climate change that would occur at the end of the century (Orr et al., 2005; Solomon et al.,
131 2007). Temperature was controlled in each bath to within $\pm 0.5^\circ\text{C}$ using temperature
132 controllers connected to 300 W submersible heaters. The pH was controlled in each bottle to
133 within ± 0.05 pH unit with a continuous pH-stat system (IKS, Karlsbad) that bubbled pure
134 CO_2 into the bottles that were continuously aerated with CO_2 -free air. The pH was maintained
135 at a mean (\pm SD) of 7.61 ± 0.11 , 7.84 ± 0.04 , and 8.09 ± 0.04 , at ambient temperature ($16.0 \pm$
136 0.1°C), and of 7.61 ± 0.08 , 7.84 ± 0.04 , and 8.09 ± 0.09 , at elevated temperature ($18.9 \pm$
137 0.3°C), corresponding to $p\text{CO}_2$ of 1399, 781, and 404 ppm at ambient temperature and 1440,
138 799, and 399 ppm at elevated temperature, respectively (Lacoue-Labarthe et al., 2009b). The
139 eggs were exposed to dissolved radiotracer: $0.27 \text{ kBq l}^{-1} {}^{241}\text{Am}$, $0.95 \text{ kBq l}^{-1} {}^{60}\text{Co}$, 1.18 kBq l^{-1}
140 ${}^{134}\text{Cs}$ (Amersham, UK) and $0.91 \text{ kBq l}^{-1} {}^{54}\text{Mn}$ and $1.04 \text{ kBq l}^{-1} {}^{75}\text{Se}$ (Isotope Product
141 Laboratory, USA). Briefly, seawater of each aquarium was spiked with typically $5 \mu\text{l}$ of
142 radioactive stock solution (${}^{241}\text{Am}$ dissolved in 1N HCl, ${}^{60}\text{Co}$ in 0.1N HCl, ${}^{134}\text{Cs}$ in H_2O , ${}^{54}\text{Mn}$
143 in 0.1N HCl and ${}^{75}\text{Se}$ in 0.1N HCl). Seawater and spikes were renewed daily during the first
144 week and then every second day to maintain good water quality and radiotracer

145 concentrations as constant as possible. The detailed experimental procedure has been
146 previously described in Lacoue-Labarthe et al. (2009b).

147 At hatching time (i.e., 62 d and 42 d at 16 and 19°C, respectively, when first hatching events
148 occurred), in each treatment, 10 eggs were fresh weighed and dissected and the radiotracer
149 activities were counted in the eggshell, the embryo and the perivitelline fluid (PVF). In
150 parallel, 10 newly hatched cuttlefish in each bottle were weighted and counted.

151 Radioactivities were measured using a high-resolution γ -spectrometry system consisting of
152 four coaxial Germanium (N- or P-type) detectors (EGNC 33-195-R, Canberra[®] and Eurysis[®])
153 connected to a multi-channel analyzer and a computer equipped with a spectra analysis
154 software (Interwinner[®] 6). Radioactivities were determined by comparison with standards of
155 known activity and of appropriate geometry. Measurements were corrected for counting
156 efficiency and physical radioactive decay. Counting times were adapted to obtain relative
157 propagated errors less than 5% (Metian et al., 2008).

158 Accumulation of radiotracers was expressed as concentration factors (CF), which is the ratio
159 between radiotracer activity in the egg or egg compartment (Bq g^{-1}) and the time-integrated
160 activity in seawater (Bq g^{-1}) (Metian et al., 2008). The distribution of metals was expressed a
161 % of the total activity load. Results are expressed as mean \pm SD.

162

163 **RESULTS**

164

165 At hatching time, only ^{60}Co , ^{134}Cs and ^{54}Mn were accumulated with sufficient activities to be
166 detected in the newly hatched juveniles, showing that ^{241}Am and ^{75}Se were not efficiently
167 accumulated during the embryonic development. The lower the pH of seawater where eggs
168 were incubating, the less ^{60}Co and ^{54}Mn were accumulated in the tissues of hatchlings (Fig. 1;
169 Table 1). Moreover, this effect tended to be reduced when the temperature increases, i.e. ^{60}Co
170 and ^{54}Mn CF values were 1.5 and 1.1 fold lower at pH 7.60 than at normal pH, when eggs
171 developed at 16 and 19°C, respectively. Finally, neither the temperature nor the pH has an
172 effect on the ^{134}Cs CF values in the hatchlings.

173

174 At the end of the embryonic development, radiotracer distribution was determined among the
175 different eggs compartments (Table 2). Regardless the temperature and $p\text{CO}_2$, over 80% of
176 both radionuclides and trace elements were found associated with the eggshell, with the
177 exception of ^{134}Cs that was found in variable proportions in the eggshell, the embryo and the
178 PVF. More precisely, the fraction of Cs in PVF was higher at lower pH levels than at normal
179 pH (KW test; $p < 0.05$), whereas the one in the eggshell was lower at pH_T 7.60 than at pH_T
180 8.10 (KW test; $p < 0.05$). No effect of the temperature was noted on the Cs fraction in PVF
181 (KW test; $p > 0.05$). Concerning Se, this element was slightly detected in the PVF but not in
182 the embryo.

183

184 **DISCUSSION**

185

186 Studying the impact of the ocean acidification on the metal accumulation on water-breathing
187 animals must raise attention on two aspects. On the one hand, the speciation of some metals

188 and their bioavailability could be affected as ^{241}Am that form strong complexes with
189 hydroxide and carbonate ions (Choppin, 2006), both of which being expected to decrease in
190 seawater (Feely, 2004). Other metals mainly found as free forms (Co^{2+} , Cs^+ , Mn^{2+}), or
191 oxydized state (SeO_4^{2-}) will be more available by only few percent with decreasing pH
192 (Byrne, 2002). Moreover, the increase in the concentration of H^+ could reduce by competition
193 for the binding site the metal adsorption on the eggshell surface or epithelia (Millero et al.,
194 2009). On the other hand, the acid-base balance and ionoregulation impairments (Pörtner,
195 2008) caused by hypercapnia could modify the active transport or passive diffusion of metals
196 through the epithelial membrane (Rainbow, 1997) and therefore the accumulation efficiencies
197 in the animal tissues.

198 As noticed in previous studies for different elements (Bustamante et al., 2002, 2004, 2006a;
199 Lacoue-Labarthe et al., 2008a, 2009a, 2010) the behaviour of trace elements towards
200 cuttlefish egg strongly varies according to the considered elements. Among the three element
201 found in the embryo, Co and Mn accumulation decreased with pH as previously observed
202 with Cd (Lacoue-Labarthe et al., 2009b). It is likely that both these “Class Borderline”
203 elements (Nieboer and Richardson, 1980) might be subjected to increased competition with
204 protons for the adsorption on binding sites on embryonic tissues. Nevertheless, regarding the
205 direct effect of temperature with similar CF at 19°C than at 16°C despite of a shorter exposure
206 duration, we assume that the presumably increased metabolic level of the embryo could
207 contribute to these essential element uptake efficiencies (Rainbow and White, 1990;
208 Barceloux, 1999). The seawater hypercapnia would in this case enhance the metabolic
209 depression of the late-stage embryo (Rosa and Seibel, 2008; Melzner et al., sbm) and limit
210 active metal accumulation. Contrasting to Co and Mn, Se passed slightly through the eggshell
211 but was not significantly accumulated by the embryo, suggesting that the need of this
212 essential metal (Bell et al., 1986) was entirely supplied by the maternal transfer through Se

213 incorporation in the yolk reserve (Lacoue-Labarthe et al., 2008b). More globally, these results
214 raise the question of the impact of a potential depletion in essential elements during the
215 sensitive early-life stages of this marine invertebrate.

216 Regarding the radionuclide behaviours, the absence of ^{241}Am neither in the embryo nor in the
217 PVF confirms that the eggshell, retaining more than 98% of the Am amount, totally prevents
218 against ^{241}Am penetration (Bustamante et al., 2006a; Lacoue-Labarthe et al., 2010) and that
219 pH and temperature variations did not affect these shielding properties although the
220 proportion of bioavailable Am increase with decreasing pH (Choppin, 2006). Oppositely, the
221 eggshell displayed an evident permeability toward ^{134}Cs (Bustamante et al., 2006a) as CF
222 values were > 10 in the embryo. Interestingly, decreasing pH and increasing temperature did
223 not affect Cs uptake in the embryo but enhanced the proportion associated with the PVF
224 compartment (Table 2). This probably results from the enhanced egg swelling illustrated by
225 higher egg weight under hypercapnic conditions and warm water (Lacoue-Labarthe et al.,
226 2009b). Therefore the fraction of ^{134}Cs , already described as a tracer of water movement
227 during the cuttlefish egg development (Lacoue-Labarthe et al., 2010), increased with the peri-
228 vitelline volume.

229 In conclusion, this study demonstrates that both temperature and pH have distinct effects on
230 the bioaccumulation in cuttlefish embryos. Both the chemical properties of the elements and
231 the physiological responses of the organism to ocean warming and acidification could account
232 for the observed effects. The impacts of pollutants in the context of the global change, such as
233 the potential depletion of essential elements in embryo, should to be further assessed in
234 marine organisms that use coastal areas to complete their life cycle.

235

236 **ACKNOWLEDGEMENTS**

237 TL-L was supported by a post-doctoral grant of the French Conseil Général de Charente
238 Maritime (17). The International Atomic Energy Agency is grateful to the Government of the
239 Principality of Monaco for the support provided to its Environment Laboratories. This work
240 was supported by the IAEA and LIENSs and is a contribution to the "European Project on
241 Ocean Acidification" (EPOCA) which received funding from the European Community's
242 Seventh Framework Programme (FP7/2007-2013) under grant agreement n° 211384.

243

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401 Table 1. *Sepia officinalis*. Two-way ANOVA parameters testing the effects of three pH (7.60, 7.85
 402 and 8.10) and two temperatures (16 and 19°C) on the weight of the eggs (from Lacoue-Labarthe et al
 403 [33]), and on the concentration factor (CF) of ⁶⁰Co, ¹³⁴Cs and ⁵⁴Mn in the hatchlings at the end of the
 404 embryonic development (see Figures 1).

405

Parameter	pH			Temperature			pH X Temperature		
	<i>df</i>	MS	F	<i>df</i>	MS	F	<i>df</i>	MS	F
Egg weight	2	0.587	8.8 ***	1	1.486	22.2 ***	2	0.018	0.3 ns
⁶⁰ Co CF	2	20424	17.3 ***	1	454	0.7 ns	2	7465	6.3 **
¹³⁴ Cs CF	2	168	0.9 ns	1	345	1.8 ns	2	138	0.7 ns
⁵⁴ Mn CF	2	2152	17.3 ***	1	1316	10.6 **	2	2230	18.0 ***

406 *df* = degree of freedom; MS = mean squares. Probability levels for significant effects: p < 0.001 (***), p < 0.01 (**), p <
 407 0.05 (*), p < 0.1 (°); ns = non significant.

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409 Table 2. *Sepia officinalis*. Distribution ^{241}Am , ^{60}Co , ^{134}Cs , ^{54}Mn and ^{75}Se expressed as % among the
 410 different egg compartments, at the end of development (after 62 d and 42 d of incubation at 16°C and
 411 19°C, respectively) following three different pH levels at two different temperatures.
 412

Experiment	16°C			19°C		
	7.6	7.85	8.1	7.6	7.85	8.1
(a) ^{241}Am						
Eggshell	98.8 ± 0.2	98.8 ± 0.7	98.2 ± 1.2	96.9 ± 2.2	98.8 ± 0.8	98.3 ± 1.0
Embryo	< 1	< 1	< 1	< 1	< 1	< 1
PVF	< 1	< 1	< 1	< 1	< 1	< 1
(b) ^{60}Co						
Eggshell	88.4 ± 0.9	87.2 ± 2.1	90.5 ± 2.6	83.1 ± 1.8	89.7 ± 1.0	91.4 ± 0.4
Embryo	10.2 ± 0.8	11.5 ± 2.1	8.5 ± 2.6	15.0 ± 1.3	9.0 ± 0.9	8.4 ± 0.5
PVF	1.4 ± 0.1	1.3 ± 0.2	0.9 ± 0.2	1.9 ± 0.5	1.3 ± 0.4	0.2 ± 0.1
(c) ^{134}Cs						
Eggshell	27.8 ± 11.7	33.0 ± 8.0	52.7 ± 17.6	19.2 ± 3.1	24.9 ± 14.1	48.6 ± 19.5
Embryo	30.6 ± 14.9	34.0 ± 7.1	30.3 ± 14.6	39.4 ± 5.0	36.0 ± 14.3	42.5 ± 18.1
PVF	41.6 ± 8.9	33.0 ± 1.7	17.0 ± 4.9	41.3 ± 5.7	39.1 ± 6.1	8.9 ± 2.1
(d) ^{54}Mn						
Eggshell	88.9 ± 1.3	88.8 ± 0.3	88.0 ± 2.6	84.8 ± 2.1	90.2 ± 1.7	83.4 ± 1.9
Embryo	8.7 ± 0.8	8.9 ± 0.5	10.0 ± 3.0	12.8 ± 2.0	8.1 ± 1.3	15.5 ± 1.7
PVF	2.4 ± 1.0	2.3 ± 0.4	2.0 ± 0.7	2.4 ± 0.5	1.7 ± 0.4	1.1 ± 0.4
(e) ^{75}Se						
Eggshell	97.2 ± 0.4	97.5 ± 0.4	98.3 ± 0.3	95.0 ± 1.3	96.1 ± 0.4	98.0 ± 0.6
Embryo	< 1	< 1	< 1	< 1	< 1	< 1
PVF	2.0 ± 0.2	2.0 ± 0.5	1.1 ± 0.3	3.6 ± 0.8	2.7 ± 0.3	1.0 ± 0.4

413 PVF: perivitelline fluid. Letters denote statistically significant differences (Kruskall-Wallis test; $p <$
 414 0.05) between pHs for each temperature.
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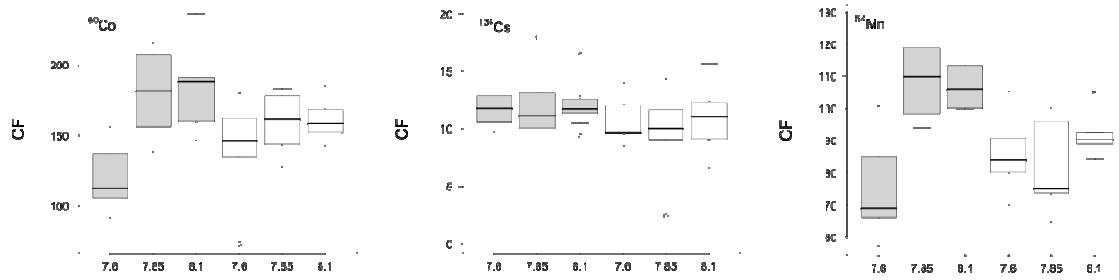


Figure 1. *Sepia officinalis*. Concentration factors of ^{60}Co , ^{134}Cs and ^{54}Mn (CF; n = 10), in the newly hatched juvenile exposed at three different pH at 16°C (grey) and 19°C (white) during the whole development time, *i.e.* 62 d and 42 d, respectively. ^{241}Am and ^{75}Se activities levels were under detection limits. Results of the statistical analysis were reported on the Table 1.