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Marine Environmental Research Biomarker induction in tropical fish species on the Northwest Shelf of Australia by produced formation water Shiqian Zhu¹, Susan Codi King², and Mary L. Haasch^{1*} Environmental Toxicology Research Program, National Center for Natural Products Research, Pharmacology Department, School of Pharmacy, The University of Mississippi, University, MS. 386771; Australian Institute of Marine Science, PMB#3 Townsville, Qld 4810, Australia² • **Contact Information** Mary L. Haasch The University of Mississippi School of Pharmacy, NCNPR, ETRP 347 Faser University, MS 38677-1848 T: 662-915-5958 F: 662-915-1285 E: mlhaasch@olemiss.edu *Current Address USEPA, MED NRC Research Associate, Senior Scientist 6201 Congdon Blvd. Duluth, MN 55804 T: 218-529-5172 F: 218-529-5003 E: Haasch.Mary@epa. gov

41 ABSTRACT

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43 Normal operation of oil well platforms results in the discharge of produced formation 44 water (PFW). The expression of CYP1A, CYP2M1- and 2K1-like proteins was examined for use as possible biomarkers of PFW exposure. A pilot study on the 45 46 Northwest Shelf of Australia had indicated that PFW contamination possibly 47 contributes to induction of CYP1A-like proteins in Gold-Spotted Trevally 48 (*Carangoides fulvoguttatus*). The pilot study samples were re-examined for CYP1A, and, in addition, CYP2K1/2M1-like proteins. In a subsequent caged-fish study in the 49 same location a second species, Stripey seaperch (Lutjanus carponotatus), caught at a 50 51 clean site, were distributed to three caging sites in a PFW gradient from the Harriet A 52 production platform: A (near-field), B (far-field) and C (a non-impacted reference 53 site). Fish were sampled at time (T) T=0, T=3 and T=10 days. Significant increases of 54 CYP1A, one CYP2K1- and two CYP2M1-like proteins were noted at Site A at T=10 55 d. For another CYP2K1-like protein, a significant increase was observed at Site A 56 only at T=3 d. These results support a previous study indicating that CYP1A protein is 57 sensitive to PFW exposure. Importantly, statistically significant environmental induction of both CYP2M1- and CYP2K1-like proteins in tropical fish due to PFW 58 59 exposure had not previously been described and induction of enzymes in the CYP2 family suggest new biomarkers for PFW. In addition, the novel response of one 60 CYP2K-like protein requires further verification, but offers promise for improved 61 62 monitoring of sub-lethal responses in marine organisms. 63

- *Keywords:* CYP1A, CYP2K1, CYP2M1, produced formation water, tropical fish,
 Australia, biomarkers
- 66 67

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69 1. Introduction

Produced formation water (PFW), the oily water usually discharged from a platform after separation from the oil, is made up from formation water (water associated with the oil in the reservoir) and potentially includes water which was injected into the reservoir to maintain pressure for oil production (Holdway, 2002). As oil fields age, the amount of PFW increases as the reservoir fills with injected water (Henderson et al., 1999).

When discharged, the PFW contains a large variety of naturally occurring materials from the formation, which include residual volatile and non-volatile hydrocarbons not removed by the separation regime, and also contains chemicals added to the production stream (Burns et al., 1999). Therefore, the composition of the discharged PFW is very complex. Almost all offshore oilfields produce formation water that can have significant environmental effects if not handled properly. Australia's oil and gas production activity is predominantly on the Northwest Shelf of Western Australia, off

83 the continental margins between the North West Cape and Dampier (Swan et al., 84 1994). There has been excellent work on the toxicology of PFW (Holdway, 2002; 85 Neff, 2002; Neff et al., 2006). Unfortunately, little is known of the effects of PFW on the tropical marine environment, especially for tropical reef fish. One study 86 87 demonstrated minor toxicity of PFW from Harriet A oil platform to corals and their symbiotic algae (Jones and Heyward, 2003). Another study illustrated a preliminary 88 89 toxicological assessment of PFW including analytical detection and acute toxicity 90 bioassay using European sea bass larvae (Mariani et al., 2004). An independent 91 scientific review commissioned by the Australian Petroleum Production and 92 Exploration Association (APPEA) and the Energy Resource Development 93 Corporation (ERDC) highlighted the need to conduct environmental assessments of 94 industry practices in shallow coastal ecosystems around Australia.

95 In 1998, the Australian Institute of Marine Science (AIMS) conducted a pilot study to evaluate the use of sub-lethal stress indicators in fish for assessing exposure to 96 97 organic contaminants associated with PFW discharged from an oil and gas production platform on the Northwest Shelf of Australia. The pilot study results confirmed that 98 99 within 1000 m from the Harriet A platform there was potential for biological effects in 100 fish populations exposed to PFW effluent and that further investigation was warranted 101 at this site (Codi King et al., 2005a). For this study, these samples were re-examined 102 for CYP1A using more universally repeatable conditions, and, in addition, the 103 response of CYP2K1/2M1-like proteins was determined. In May 2003, AIMS 104 conducted a comprehensive biological and chemical assessment of PFW effluent at 105 the Harriet A platform. This more comprehensive study utilized a controlled system in 106 which fish from a reference site were collected and placed in cages, and the cages 107 were located in a gradient line moving away from the point source of PFW being 108 discharged at Harriet A. The work described here is a subset of the suite of biomarkers 109 conducted for the 2003 study and presents in detail the cytochrome protein 110 biomarkers in an ecologically relevant reef species, Stripey seapearch (Lutjanus 111 carponotatus) for the evaluation of the potential effects of PFW discharge at the 112 Harriet A Platform on the Northwest Shelf of Australia.

113 2. Materials and Methods

114 **2.1 Pilot study: study site and fish collection**

115 Harriet A, an oil and gas production platform located 135 km west of Karratha on the 116 Northwest Shelf of Australia, was the main focus of this study (Figure 1). The 117 hydrocarbon concentration of PFW effluent in the sea-surface micro-layer, water 118 column, sediments and biota, at this location, have been well documented (Burns and 119 Codi, 1999) The site selections for the pilot study have previously been described in 120 Codi King et al. (2005a). Gold-spotted Trevally was chosen for the pilot study 121 because of its importance both commercially and recreationally to the NW Shelf of 122 Australia. A total of 41 Gold-spotted Trevally were collected from each of three sites: 123 Harriet A (n=23, Mean + SD; 3817 + 1532 g total weight, 654 + 103 mm total length, age 1-5 yr); Harriet C (n=6, Mean + SD; 3608 + 1503 g total weight, 654 + 103 mm 124

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125 126 127 128 129 130 131	total length, age 2-6 yr); and reference site off the Montebello Islands (n=12, Mean \pm SD; 3992 \pm 1122 g total weight, 660 \pm 75 mm total length, age 2-5 yr). The data demonstrated there were no significant differences in total weight, total length and age of fish between the three study sites ($P < 0.05$). All fish were processed as previously described in Codi King et al. (2005a). Briefly, livers were removed, frozen in liquid nitrogen and transported to AIMS via a liquid nitrogen dry shipper remaining stored in liquid nitrogen until thawed for microsomal preparation (see Section 2.3).

132 **2.2 Caged fish study: Study site and fish collection**

The PFW plume from Harriet A forms a fine surface slick that moves in a NNW 133 134 direction on the ebb tides and SSE direction on the flood tide. The PFW hydrocarbons 135 in the sea-surface microlayer were detectable to a distance of 1 km. The lighter MW 136 components (naphthalene and homologs) were in the dissolved phase while the 137 heavier MW hydrocarbons (phenanthrene and homologs) were associated with the 138 particulate phase. The particulate phase was removed from the water column within a 139 distance of 0.9 - 2 km of Harriet A (Burns and Codi, 1999; Holdway and Heggie, 140 1998). Therefore for the caged fish study, the cages were situated along the gradient 141 of PFW formed as the plume moves away from the discharge site. Site A (within 200 142 m NNW of Harriet A) was the near-field site; Site B (~ 1000 m NNW of Harriet A) was the far-field site and Site C (~20 km NNW of Harriet A; in the Montebello 143 144 Islands) was chosen as the reference site due to the lack of known contamination 145 (Figure 2). The field component of this project was conducted aboard the AIMS RV 146 Cape Ferguson from May 16-31, 2003. Initially, the target species was Red Emperor 147 (Lutjanus sebae), since it could be caged successfully (Mike Cappo, pers. comm.). On 148 the first day of collection; however, only 5 juvenile L. sebae were caught and another 149 species, Stripey seaperch (Lutjanus carponotatus) were being trapped at a much 150 higher proportion (10:1). L. carponotatus is an important recreational fishery in 151 Dampier Harbor and preliminary biomarker research with this species from the 152 Harriet A pilot study (Codi et al., 2001) suggested that it would suffice as a substitute 153 species for L. sebae. From May 16-19 2003, an estimated 130 L. carponotatus were 154 collected by fish traps along the Dampier Harbor LNG pipeline and maintained in a 155 1000-L holding tank with running seawater and air. On May 21, all fish were placed 156 in cages with 2 cages per station, 15 fish per cage at each of 3 sites (Figure 2). Fish 157 were fed individually quick frozen (IQF) pilchards of the highest quality (Western 158 Australia Bait Supply, Fremantle), collected from clean local waters off of Albany, 159 WA. Food canisters were replenished every two days during the field exposure. Fish were sampled at T=0, 3 and 10 d. 160

161 A total of 50 Stripey seaperch were analyzed for CYP1A, CYP2M- and CYP2K-like 162 proteins from the holding tank (T=0), and T=3 and T=10 for cage Sites A, B, and C. 163 There were no significant differences for total weight, total length and age between 164 the study sites (P < 0.05; Table 1). All fish were handled and processed in a similar 165 manner as stated for the pilot study.

166 2.3. Microsomal preparation and Western blot

Hepatic microsomal samples were prepared at the Australian Institute of Marine
Science, Townsville, Qld (AIMS) following methods previously described (Haasch,
2002) and shipped on dry ice to The University of Mississippi, University, MS.
Hepatic microsomal samples were stored at -80°C until analyzed. Protein content of
the microsomal fraction was determined by Bio-Rad DC Protein Assay (Hercules,
CA) using bovine serum albumin (BSA) as a standard. Microsomal proteins (10 µg

173 per well) were separated using pre-cast Criterion SDS-PAGE gels (10%T, 2.7%C 174 resolving gel; 4%T, 2.7%C stacking gel; Bio-Rad) at 200V for 1 h with the Bio-Rad 175 Criterion apparatus. Proteins were then electrotransferred (100V for 0.75 h) to a 176 PVDF membrane (FluoroTrans W, $0.2 \mu m$; VWR) together with one lane containing 177 MagicMark protein standards (contain proteins with IgG binding site; Invitrogen, Carlsbad, CA) detectable by chemiluminescence, for molecular weight determination 178 179 and for evaluation of transfer efficiency. Sample loading buffer also contained 180 Pyronin Y which is visible on the membrane and marks the dye front. Detection of 181 proteins of interest was accomplished by using one of the primary polyclonal 182 antibodies; rabbit anti-fish CYP1A (two synthetic peptides used as immunogens; 183 BioSense, Bergen, NO distributed by Cayman Chemical, Ann Arbor, MI), rabbit 184 anti-trout CYP2K1, rabbit anti-trout CYP2M1 (both gifts from Donald R. Buhler, 185 PhD, Oregon State University, Corvallis, OR). Proteins of interest were visualized 186 with goat anti-rabbit alkaline phosphatase conjugate (Bio-Rad) and the Western 187 Breeze Novex Western Blot Immunodetection Kit (Invitrogen) according to the 188 manufacturer's directions using Western Lightening CDP-Star Chemiluminescent 189 reagent (Perkin Elmer Life Sciences, Boston, MA). Image analysis was accomplished 190 using the VersaDoc 3000 Imaging System and QuantityOne analysis software 191 (Bio-Rad). A summary of CYP1A results of the pilot study samples was previously 192 presented (Codi King et al., 2005a) in which the samples were analyzed with a 193 different primary antibody (anti-trout CYP1A1277-294/KLH) under different SDS-PAGE 194 conditions (mini-Protean II, individually poured gels in the previous study) and with different methods for quantification (film, scanning densitometry and NIH Image® in 195 the previous study). For this study, a commercially available antibody was used and 196 197 the number of samples that could be compared on one gel was increased improving 198 the statistical analyses.

199 2.4 Densitometry and Statistical Analysis

200 The optical density of the samples is given as a Trace Quantity (QTY; Bio-Rad 201 QuantityOne) or the quantity of a band as measured by the area under its intensity 202 profile curve. Trace QTY units are intensity ×area. For standardization, one sample in 203 triplicate was loaded on each gel. The mean Trace QTY of this sample was used as a 204 gel factor for standardization between gels. Trace QTY results are presented as an 205 Arbitrary Optical Density (without gray-scale calibration). Statistical analysis was 206 performed using GraphPad Prism. For the pilot study, a two-way ANOVA was 207 performed using both gender and site as covariates. The results indicated that for each 208 individual isozyme site was a significant factor, but not gender, and there were no 209 interactions between these two factors. Therefore, for the one-way ANOVA the males 210 and females were pooled, and the ANOVA results were analyzed by Turkey's multiple 211 comparison test (P < 0.05, 0.01 and 0.001). Similar analysis of covariates allowed the 212 use of the same statistical approach for the caged fish study.

- 213 3. Results
- 214 3.1 Pilot study

The results of the immunodetection of the CYP1A-, CYP2M1-, and CYP2K1-like 215 proteins indicated that proteins of the appropriate molecular weight range for each 216 217 isozyme were detectable in Gold-Spotted Trevally. Most Gold-Spotted Trevally 218 showed one immunodetectable band of CYP1A-, CYP2M1- and CYP2K1- like 219 proteins; some samples had a second CYP2M1- like protein. Because this second 220 protein was not present in each sample, only the first band of lower molecular weight 221 was quantified (Figure 3). Similar to the previous study (Codi King et al., 2005a), CYP1A-like protein levels in Gold-Spotted Trevally from the near-field (Harriet A) 222 223 site were significantly higher than those from the far-field (Harriet C) site and the 224 reference site. Although not significant, Gold-Spotted Trevally also exhibited a trend 225 of increased CYP2M1-like protein levels at Harriet A as compared to Harriet C and 226 the reference site (Figure 3).

For CYP2K1-like protein, a significant increase was observed at the Harriet C site compared to the reference site (Figure 3). The induction of CYP2K1-like proteins at the Harriet C site in Trevally may be due to possible exposure to hydrocarbons from natural oil seepages found in this region, or due to the movement of fish between platforms. The elimination of these variables was one of the main reasons for undertaking the subsequent controlled field study using caged fish.

233 3.2. Caged fish study

234 Immunodetection of CYP1A-like proteins indicated that proteins of the appropriate 235 molecular weight range were detectable in Stripey seaperch at all three sites and over 236 the time of the exposure (Figure 4a). The results for the CYP1A-like proteins (Figure 237 4b) demonstrated that by Day 3 of exposure both Site A and Site B had higher mean 238 values than Site C, albeit not significant. By Day 10, fish collected at Site A had 239 significantly higher induction of CYP1A-like proteins as compared to the fish at Site 240 B and Site C. Comparisons also demonstrated that Site A fish after 10 days of 241 exposure had significantly induced levels of CYP1A-like proteins as compared to the 242 holding tank fish (T=0).

Figure 5a presents the optical densities for the species-specific pattern of 243 244 CYP2M1-like proteins determined in Stripey seaperch, with two distinct appropriate 245 molecular weight proteins in most samples; CYP2M1-like Protein 1 (52.8 kDa) and 246 CYP2M1-like Protein 2 (49.0 kDa). Immunodetection of CYP2M1-like proteins 247 demonstrated a similar pattern of expression for both Protein 1 and Protein 2 in 248 Stripey seaperch with significant induction occurring at Day 10 for both proteins 249 (Figures 5b and 5c) and at Day 3 the same increasing but not significant trend. For 250 Site A both Day 3 and Day 10 exposed Stripey seaperch had significantly higher 251 levels of CYP2M1-like proteins than the holding tank fish. It is interesting that Site C 252 Day 3 exposed fish also demonstrated site-specific differences from the holding tank 253 fish.

Figure 6a presents the optical densities for the species-specific pattern of CYP2K1-like proteins determined in Stripey seaperch, again showing two distinct appropriate molecular weight proteins; CYP2K1-like Protein 1 (52.5 kDa) and

257 CYP2K1-like Protein 2 (48.2 kDa). Unlike the two CYP2M1-like proteins, the 258 induction patterns for the two CYP2K1-like proteins were dissimilar. For 259 CYP2K1-like Protein 1, levels (Figure 4b) were significantly elevated for Site A Day 260 3 as compared to fish collected on Day 3 from Site B and Site C. At Day 10 there 261 were no site-specific differences in expression, although Site A Day 10 and Site A Day 3 Stripey seaperch had significantly higher CYP2K1-like Protein 1 levels than 262 263 the holding tank fish. For CYP2K1-like Protein 2 levels (Figure 4c) the induction 264 trend was similar to the pattern observed for CYP1A- and both CYP2M1-like Protein 265 1 and Protein 2 in that by Day 10, Site A fish demonstrated the highest mean level of 266 induction. The same site-specific differences were also observed with Site A Day 10 267 fish and Day 3 fish having significantly higher CYP2K1-like Protein 2 induction 268 levels than the holding tank fish.

269 4. Discussion

270 Fish are important organisms for monitoring environmental contaminants because of 271 their position in the food web, their lifestyle, their relative abundance and their 272 adaptable physiology. Fish CYP1A induction is an extremely useful biomarker of 273 chemical exposure in the environment (Flammarion et al., 2002), particularly for 274 persistent pesticides (Levine et al., 1999), polychlorinated biphenyls (Burton et al., 275 2002; Haasch et al., 1993), dioxins (Zodrow et al., 2004; Smith, 2003), pulp mill 276 effluents (Oikari et al., 2002), xenoestrogens (Katchamart et al., 2002), spilled crude 277 oil (Jewett et al., 2002), and polycyclic aromatic hydrocarbons (PAHs) (Stagg et al., 278 2000). The induction of CYP1A has implications for human fish consumption, as well 279 as for the health status of aquatic organisms (Arinc et al., 2000).

280 The induction of CYP2 family proteins in bluegill and catfish due to exposure to 281 peroxisome proliferating agents (PPAs; Haasch, 1996) was the first description of an 282 inducible P450 in fish other than CYP1A. The CYP2M1- and 2K1-like proteins in 283 bluegill and catfish were later shown to be lauric acid hydroxylases (Haasch et al., 284 1998). Bivalves exposed to petroleum water-accommodated fraction (WAFs) produce peroxisome proliferation (Cajaraville et al., 1997). Therefore, petroleum hydrocarbon, 285 286 as the major contaminant in PFW, may be considered a PPA, even though the 287 mechanism is unknown. These findings provided an indication that a petroleum 288 hydrocarbon-initiated peroxisome proliferation response may occur in lower 289 vertebrates similar to the model PPA-mediated induction of CYP2M1/2K1 (Ackers et 290 al., 2000; Johnson et al., 1996).

291 These two CYP2 family P450 isozymes have been most thoroughly characterized in 292 rainbow trout, (Oncorhynchus_mykiss; Buhler and Wang-Buhler, 1998). The molecular 293 weight of CYP2K1 in rainbow trout is about 54 kDa, and has been shown to catalyze 294 the $(\omega-1)$ and $(\omega-2)$ hydroxylation of lauric acid and some longer chain fatty acids; the 295 molecular weight of CYP2M1 in rainbow trout is about 50 kDa, and this isozyme is 296 the only known P450 responsible for laurate hydroxylation at the (ω -6) position. Both 297 isozymes also catalyze the hydroxylation of estradiol, testosterone and progesterone 298 (Buhler and Wang-Buhler, 1998). It follows that induction of CYP2K1- and

299 CYP2M1-like proteins in fish may have effects on fatty acid and steroid homeostasis, 300 with possible links to developmental and reproductive toxicity as well as 301 carcinogenesis (Biegel et al., 1995; Vanden Heuvel, 1996; Vanden Heuvel et al., 1991). 302 A variety of structurally diverse xenobiotics and pharmaceuticals are known to 303 produce peroxisome proliferation and concordant fatty acid metabolizing enzyme 304 induction in mammals, therefore, the monitoring of CYP2 family isozymes in fish 305 may have implications for human health and has the potential to become an important 306 indicator of environmental health.

307 The immunodetection results indicated one highly immunoreactive protein for CYP1A and multiple immunoreactive proteins for the CYP2-like proteins. Some 308 309 cross-reactivity between CYP2-like proteins is expected with the polyclonal 310 anti-CYP2K1 and anti-CYP2M1 antibodies (some affinity for the same protein 311 epitopes) as these antibodies have cross-reactivity to both rainbow trout CYP2K1 and 312 CYP2M1 (Buhler and Wang-Buhler, 1998). Further evidence for cross-reactivity was 313 evident in samples that also included one or more minor proteins of intermediate size 314 that were specific to individuals and not consistently represented across all 315 individuals. The CYP2 family is known to be fairly complex in number of isozymes 316 and expression levels (Buhler and Wang-Buhler, 1998). Multiple cross-reactivity was 317 not unexpected and did not detract from the relatively clear induction of two proteins 318 using the anti-CYP2 antibodies. The proteins were designated CYP2M1- and 319 CYP2K1-like proteins because appropriate standards for the tropical fish species do 320 not yet exist and were beyond the scope of this study. Nevertheless, it is evident that 321 the CYP2 family antibodies recognize constitutively present and inducible proteins 322 that could potentially be used as biomarkers. Furthermore, it is possible, with 323 appropriate sequence information, for commercially-available anti-CYP2 family antibodies to be developed. 324

325 The pilot study CYP1A results supported the findings of an earlier analysis of those 326 samples using a different methodology indicating induction of CYP1A (Codi King et 327 al., 2005a). Unlike the previous analysis, the current study did not indicate 328 gender-specific responses. Multiple differences between the two analyses could 329 contribute to the discrepancy including the use of a different primary antibody 330 preparation. Similar to the previous analysis, site-specific responses were detected. 331 Analysis of the pilot study samples for CYP2M1- and CYP2K1-like proteins 332 indicated possible site-specific induction and prompted the further examination of 333 these proteins in the caged fish study.

Studies of environmental contaminant-mediated induction of the cytochromes P450 in tropical fish are lacking (Codi King et al., 2005a; Vrolijk et al., 1994). Both the pilot and caged fish field studies indicated significant results pertaining to the induction of cytochromes P450 in tropical fish species by PFW. First, the induction of CYP1A in Gold-Spotted Trevally and Stripey seaperch at the near-field site indicated that even though PFW discharge limits are being met (Codi King et al., 2005b), exposure to CYP1A inducing chemicals within the diluted PFW produced a significant induction

341 effect. Numerous studies have demonstrated that chronic low level exposure of fish 342 and especially larvae to PAHs can have a significant impact and can produce 343 long-term, adverse health and reproductive effects (Incardona et al., 2004; Stegeman 344 et al., 2001). These adverse effects are most likely to be more serious to those 345 organisms attached to the platform or living under or near the platform in a restricted 346 range but could also be detrimental to organisms cyclically exposed to pools of PFW. 347 Second, the significant induction of the CYP2K1/2M1-like proteins in both the 348 Gold-Spotted Trevally and the Stripey seaperch is the first documented significant 349 environmental induction in fish of any cytochrome P450 isozyme other than 350 CYP1A1. The similar induction pattern of CYP1A1 and CYP2K1-like Protein 2 and 351 CYP2M1-like Protein 1 and 2 may indicate that all of these isozymes are induced by 352 the same chemical class contained within the PFW. While this finding alone is a 353 reason for further study, it is perhaps even more interesting that there is an apparent 354 response difference for induction of particular CYP2K1/2M1-like proteins. Only the 355 CYP2K1-like Protein 1 was induced at the near-field site (Site A) on Day 3, but not 356 on Day 10. The apparently different induction of the CYP2K1 Protein 1 compared to 357 the other CYP isozymes potentially represents a unique chance to develop another 358 bioindicator of PFW exposure. Although the metabolic activities of the CYP2M1 and 359 CYP2K1 isozymes in Stripey seaperch are not known, it is reasonable to assume that 360 the fatty acid and reproductive steroid hormone substrates would differ among the different proteins. Induction of different isozymes may therefore have different short-361 362 and long-term effects on the exposed organisms.

363 In support of the previous findings, the results indicate that CYP1A protein is 364 sensitive to PFW exposure. Importantly, statistically significant environmental 365 induction of both CYP2M1- and CYP2K1-like proteins in tropical fish due to PFW exposure had not previously been described and CYP2 family induction may 366 367 represent possible new biomarkers (other than CYP1A) associated with PFW. In 368 addition, the novel response of one CYP2K-like protein requires further verification 369 but offers promise for improved monitoring of sub-lethal responses in marine 370 organisms.

371 References

- 372
- Ackers, J.T., Johnston, M.F., and Haasch, M.L. (2000). Immunodetection of hepatic peroxisomal PMP70 as an indicator of peroxisomal proliferation in the mummichog, *Fundulus heteroclitus*. Marine Environmental Research, 50, 361-365.

Arinc, E., Sen, A., and Bozcaarmutlu, A. (2000). Cytochrome P4501A and associated mixed-function oxidase induction in fish as a biomarker for toxic carcinogenic pollutants in the aquatic environment. Pure and Applied Chemistry, 72, 985-994.

- Biegel, L.B., Liu, R.C., Hurtt, M.E., and Cook, J.C. (1995). Effects of ammonium perfluorooctanoate on Leydig cell function: in vitro, in vivo, and ex vivo studies. Toxicology Applied Pharmacology, 134, 18-25.
- Buhler, D.R., and Wang-Buhler, J.L. (1998). Rainbow trout cytochrome P450s: purification, molecular aspects, metabolic activity, induction and role in environmental monitoring. Comparative Biochemistry and Physiology. Part C, Pharmacology, Toxicology and Endocrinology, 121, 107-137.
- Burns, K.A., and Codi, S. (1999). Non-volatile Hydrocarbon Chemistry Studies Around a Production Platform on Australia's Northwest Shelf. Estuarine, Coastal and Shelf Science, 49, 853-876.
- Burns, K.A., Codi, S., Furnas, M., Heggie, D., Holdway, D., King, B., and McAllister, F. (1999). Dispersion and fate of produced formation water constituents in an Australian Northwest Shelf shallow water ecosystem. Marine Pollution Bulletin, 38, 593-603.
- Burton, J.E., Dorociak, I.R., Schwedler, T.E., and Rice, C.D. (2002). Circulating lysozyme and hepatic CYP1A activities during a chronic dietary exposure to tributyltin (TBT) and 3,3',4,4',5-pentachlorobiphenyl (PCB 126) mixtures in channel catfish, *Ictalurus punctatus*. Journal of Toxicology and Environmental Health, Part A, 65, 589-602.
- Cajaraville, M.P., Orbea, A., Marigomez, I., and Cancio, I. (1997). Peroxisome proliferation in the digestive epithelium of mussels exposed to the water accommodated fraction of three oils. Comparative Biochemistry and Physiology, C: Pharmacology, Toxicology and Endocrinology, 117C, 233-242.
- Codi King, S., Johnson, J.E., Haasch, M.L., Ryan, D.A.J., Ahokas, J.T., and Burns, K.A. (2005a). Summary results from a pilot study conducted around an oil production platform on the Northwest Shelf of Australia. Marine Pollution Bulletin 50, 1163-1172.
- Codi King, S., Bennett, C., Haasch, M.L., Zhu, S., Mondon, J., Mueller, J, McLean, C., Humphrey, C., and Shaw, M. (2005b). Chemical and biological assessment of produced formation water (PFW) effluent from Harriet A Platform, on a tropical reef species, Stripey searperch (*Lutjanus carponotatus*). Final Report for Apache Energy Pty. Australian Institute of Marine Science. 125 pgs.
- Codi, S., Burns, K.A., Johnson, J.E., Ramsay, M., Ryan, D.J.A., and Haasch, M.L. (2001). A pilot study conducted around Harriet A petroleum production on the Northwest Shelf of Australia: Assessment of sub-lethal stress indicators in fish as potential indicators of petroleum hydrocarbon exposure. Final report to Apache Energy Pty Ltd., 73 pgs.
- Flammarion, P., Devaux, A., Nehls, S., Migeon, B., Noury, P., and Garric, J. (2002). Multibiomarker Responses in Fish from the Moselle River (France). Ecotoxicology and Environmental Safety, 51, 145-153.

- Haasch, M.L. (1996). Induction of anti-trout lauric acid hydroxylase immunoreactive proteins by peroxisome proliferators in bluegill and catfish. Marine Environmental Research, 42, 287-291.
- Haasch, M.L. (2002). Effects of vehicle, diet and gender on the expression of PMP70- and CYP2K1/2M1-like proteins in the mummichog. Marine Environmental Research, 54, 297-301.
- Haasch, M.L., Henderson, M.C., and Buhler, D.R. (1998). Induction of lauric acid hydroxylase activity in catfish and bluegill by peroxisome proliferating agents. Comparative Biochemistry and Physiology, Part C: Pharmacology, Toxicology and Endocrinology, 121C, 297-303.
- Haasch, M.L., Prince, R., Wejksnora, P.J., Cooper, K.R., and Lech, J.J. (1993). Caged and wild fish: Induction of hepatic cytochrome P-450 (CYP1A1) as an environmental biomonitor. Environmental Toxicology Chemistry, 12, 885-895.
- Henderson, S.B., Grigson, S.J.W., Johnson, P., and Roddie, B.D. (1999). Potential impact of production chemicals on the toxicity of produced water discharges from North Sea oil platforms. Marine Pollution Bulletin, 38, 1141-1151.
- Holdway, D., and Heggie, D.T. (1998). Tracking produced formation water discharge from a petroleum production platform to the North West shelf. APPEA Journal, 38, 665-680.
- Holdway, D.A. (2002). The acute and chronic effects of wastes associated with offshore oil and gas production on temperate and tropical marine ecological processes. Marine Pollution Bulletin, 44, 185-203.
- Incardona, J.P., Collier, T.K., and Scholz, N.L. (2004). Defects in cardiac function precede morphological abnormalities in fish embryos exposed to polycyclic aromatic hydrocarbons. Toxicology Applied Pharmacology, 196, 191-205.
- Jewett, S.C., Dean, T.A., Woodin, B.R., Hoberg, M.K., and Stegeman, J.J. (2002). Exposure to hydrocarbons 10 years after the Exxon Valdez oil spill: evidence from cytochrome P4501A expression and biliary FACs in nearshore demersal fishes. Marine Environmental Research, 54, 21-48.
- Johnson, E.F., Palmer, C.N., Griffin, K.J., and Hsu, M.H. (1996). Role of the peroxisome proliferator-activated receptor in cytochrome P450 4A gene regulation. FASEB Journal, 10, 1241-1248.
- Jones, R.J., and Heyward, A.J. (2003). The effects of produced formation water (PFW) on coral and isolated symbiotic dinoflagellates of coral. Marine and Freshwater Research, 54, 153-162.
- Katchamart, S., Miranda, C.L., Henderson, M.C., Pereira, C.B. and Buhler, D.R. (2002). Effect of xenoestrogen exposure on the expression of cytochrome P 450 isoforms in rainbow trout liver. Environmental Toxicology and Chemistry, 21, 2445-2451.
- Levine, S.L., Oris, J.T. and Denison, M.S. (1999). Modulation of CYP1A expression in rainbow trout by a technical grade formulation of propiconazole. Environmental Toxicology and Chemistry, 18, 2565-2573.
- Mariani, L., Manfra, L., Maggi, C., Savorelli, F., Di Mento, R., and Cicero, A.M. (2004). Produced formation waters: a preliminary study on chemical characterization and acute toxicity by using fish larvae (*Dicentrarchus labrax* L., 1758). Fresenius Environmental Bulletin 13, 1427-1432.

- Neff, J.M. (2002). Bioaccumulation in Marine Organisms: Effect of Contaminants from Oil Well Produced Water, Elsevier Science. 452 pgs.
- Neff, J.M., Johnsen, S., Frost, T.K., Utvik, T.I., and Durell, G.S. (2006). Oil well produced water discharges to the North Sea. Part II: comparison of deployed mussels (*Mytilus edulis*) and the DREAM model to predict ecological risk. Marine Environmental Research 62, 224-246.
- Oikari, A., Fragoso, N., Leppanen, H., Chan, T., and Hodson, P.V. (2002). Bioavailability to juvenile rainbow trout (*Oncorhynchus mykiss*) of retene and other mixed-function oxygenase-active compounds from sediments. Environmental Toxicology and Chemistry, 21, 121-128.
- Smith, P.N. (2003). Environmental polychlorinated biphenyl exposure and cytochromes P450 in raccoons (*Procyon lotor*). Environmental Toxicology and Chemistry, 22, 417-423.
- Stagg, R.M., Rusin, J., McPhail, M.E., McIntosh, A.D., Moffat, C.F., and Craft, J.A. (2000). Effects of polycyclic aromatic hydrocarbons on expression of CYP1A in salmon (*Salmo salar*) following experimental exposure and after the Braer oil spill. Environmental Toxicology and Chemistry, 19, 2797-2805.
- Stegeman, J.J., Schlezinger, J.J., Craddock, J.E., and Tillitt, D.E. (2001). Cytochrome P450 1A expression in midwater fishes: potential effects of chemical contaminants in remote oceanic zones. Environmental Science and Technology, 35, 54-62.
- Swan, J.M., Neff, J.M. and Young, P.C. (1994). Environmental Implications of Offshore Oil
 & Gas Development in Australia: The findings of an independent scientific review.,
 Australian Petroleum Production & Exploration Association, Sydney. 125 pgs.
- Vanden Heuvel, J.P. (1996). Perfluorodecanoic acid as a useful pharmacologic tool for the study of peroxisome proliferation. General Pharmacology, 27, 1123-1129.
- Vanden Heuvel, J.P., Kuslikis, B.I., Shrago, E., and Peterson, R.E. (1991). Inhibition of long-chain acyl-CoA synthetase by the peroxisome proliferator perfluorodecanoic acid in rat hepatocytes. Biochemical Pharmacology, 42, 295-302.
- Vrolijk, N., Targett, M., Woodin, B.R., and Stegeman, J.J. (1994). Toxicological and ecological implications of biotransformation enzymes in the tropical teleost *Chaetodon capistratus* Marine Biology, 119, 151-158.
- Zodrow, J.M., Stegeman, J.J., and Tanguay, R.L. (2004.) Histological analysis of acute toxicity of 2,3,7,8-tetrachlorodibenzo-p-dioxin (TCDD) in zebrafish. Aquatic Toxicology, 66, 25-38.

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383 Table 1: Characteristics of Stripey seaperch used in the caged fish study. All of the

384 fish were of intermediate reproductive status (sex-stage) and had plasma vitellogenin

385 levels that were below the detection limit.

Holding Tank or Cage Site	Time (days)	Sex Ratio (Female/Male)	Body Weight [*]	Body Length [*]	Age [*] (vr).∕
ouge site	(uujs)	(I children in the children in	(8)	(1111)	()1)
Holding Tank	0	(4/7)	271 <u>+</u> 85	261 <u>+</u> 22	5 <u>+</u> 2
	3	(4/4)	292 <u>+</u> 83	265 <u>+</u> 30	6 <u>+</u> 4
Reference Site C	10	(4/4)	284 <u>+</u> 52	257 <u>+</u> 19	4 <u>+</u> 2
	3	(0/5)	273 <u>+</u> 123	260 <u>+</u> 34	5 <u>+</u> 3
Far-Field Site B	10	(0/5)	324 <u>+</u> 92	269 <u>+</u> 21	3 <u>+</u> 2
Norr Field Side A	3	(3/5)	371 <u>+</u> 88	285 <u>+</u> 20	5 <u>+</u> 3
Near-Field Site A	10	(1/4)	293 <u>+</u> 79	262 <u>+</u> 28	5 <u>+</u> 2

386 *Values are mean ± SD for male and female combined.

Colored Colore

387





390 triangles, off the Montebello Islands.





Figure 2. Site locations of the caged fish study: (Site A ~200 m, Site B ~1000 m and Site C ~20 km)
relative to the Harriet A PFW discharge.

395

396	Figure 3. Pilot study: the species specific pattern of CYP1A-, CYP2M1-, and CYP2K1- like proteins
397	determined in Gold-Spotted Trevally (shown for comparison of isozyme expression not for illustration
398	of site effects), and graphical representation of the optical densities (Mean ± SD) determined for each
399	site. The molecular weight range of the proteins is typical of cytochromes P450 and is between 45 and
400	60 kDa for all three isozymes. Numbers above the bars indicate the number of female and male fish
401	(e.g., 6/2, female/male) represented in the average densitometry values. No statistical difference was
402	observed between genders.

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405 Figure 4 a-b. Caged fish study: the species-specific pattern of CYP1A-like proteins determined in

406 Stripey seaperch. (a) Image of the chemiluminescence detection of CYP1A-like proteins with a

407 single distinct band (56.1 kDa) in the appropriate molecular weight range (shown for comparison

408 of isozyme expression not for illustration of site effects), and (b) graphical representation of the

409 optical densities (Mean ± SD) determined in Stripey seaperch at each site over time. Numbers

- 410 above the bars indicate the number of female and male fish (e.g., 4/7, female/male) represented in
- 411 the average densitometry values. No statistical difference was observed between genders.



413	Figure 5 a-c. Caged fish study: the species specific pattern of CYP2M1-like proteins determined in
414	Stripey seaperch. (a) Image of the chemiluminescence detection of CYP2M1-like proteins with two
415	distinct bands of protein (shown for comparison of isozyme expression not for illustration of site
416	effects), (b) the graphical representation of optical densities (Mean ± SD) determined for Protein 1 and
417	(c) for Protein 2. Numbers above the bars indicate the number of female and male fish (e.g., 4/7,
418	female/male) represented in the average densitometry values. No statistical difference was observed
419	between genders.



420

Figure 6 a-c. Caged fish study: the species specific pattern of CYP2K1-like proteins determined in Stripey seaperch. (a) Image of the chemiluminescence detection of CYP2K1-like proteins with two distinct bands of protein (shown for comparison of isozyme expression not for illustration of site effects), (b) the graphical representation of optical densities (Mean ± SD) determined for Protein 1 and (c) for Protein 2. Numbers above the bars indicate the number of female and male fish (e.g., 4/7, female/male) represented in the average densitometry values. No statistical difference was observed between genders.

