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Coherent response of lakes in Ontario, Canada to reductions in sulphur deposition: the effects of climate on sulphate concentrations

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Abstract

Sulphate deposition in south-central Ontario declined between 1976 and 2000 by more than 50%, whereas lake sulphate (SO_4^{2-}) concentrations decreased by, on average, only half as much. To investigate the factors that controlled this slower than expected response, the temporal patterns in lake SO_4^{2-} concentrations were compared with patterns in both deposition and climate, since climate has a major influence on the hydrological cycle in this part of the continent. To do this, the temporal coherence in SO_4^{2-} concentrations between 9 lake basins was estimated using the intraclass correlation from a repeated-measures analysis of variance and two subsets of lakes were found (six in one group, four in the other), each with lakes having synchronous patterns. One subset (4 lakes) included the 3 with the longest water replenishment times (>3.4 yr) which are expected to respond to decreases in SO_4^{2-} deposition more slowly. However, the average pattern reflecting the temporal changes of each of the two subsets was very similar. The response of both subsets of lakes to the decreasing SO_4^{2-} deposition over two decades was independent of the degree of acidification or sensitivity to acidification of the lakes. In a determination of which factors best predicted each of those two subsets' SO_4^{2-} time series, good predictive models were produced by regional/global-scale climate indices, specifically the Southern Oscillation Index (SOI) describing the El Niño Southern Oscillation (ENSO) and the North Atlantic Oscillation Index (NAOI), as well as by SO_4^{2-} deposition indices. When the predictor variables were combined, models which described the long-term changes in lake SO_4^{2-} concentration best included the SOI, the NAOI and SO_4^{2-} deposition. Thus, large-scale climate factors play a major role in determining the response of aquatic systems to changes in SO_4^{2-} deposition, perhaps through their influence on lake and/or catchment processes that effectively delay recovery.

Keywords: Atmospheric deposition, lake recovery, temporal trends, climate, temporal coherence

Introduction

Sulphur emission control programmes, some of which were implemented as early as the 1970s, have reduced both SO_2 emissions and S deposition in most regions of Europe, the United States (Hedin *et al.*, 1987; Driscoll *et al.*, 1995) and Canada (Dillon *et al.*, 1988; Sirois and Summer, 1989; Bouchard, 1997). Furthermore, these declining SO_2 emissions and S deposition have resulted generally in lower concentrations of SO_4^{2-} (sulphate) in surface waters, the simplest and most direct measure of lake recovery, in both the United States (Driscoll and van Dreason, 1993; Newell, 1993; Stoddard and Kellogg, 1993; Webster *et al.*, 1993;

Webster and Brezonik, 1995; Stoddard *et al.*, 2003) and Canada (Dillon and LaZerte, 1992; Clair *et al.*, 1995; Jeffries *et al.*, 2000; Bouchard, 1997).

However, the reductions in surface water SO_4^{2-} concentrations have often been much less than those anticipated, following the reduction in atmospheric S deposition (Driscoll *et al.*, 1995; Dillon *et al.*, 1997; Dillon and Evans, 2001; Löfgren *et al.*, 2001). Furthermore, other important measures of chemical recovery, namely increases in lake pH and acid neutralising capacity, ANC, have provided conflicting evidence. In a few cases (Kahl *et al.*, 1993; Webster and Brezonik, 1995; Stoddard *et al.*, 1998),

they have provided limited evidence of chemical recovery but in many others (Clair *et al.*, 1995; Houle *et al.*, 1997; Dillon *et al.*, 1997), there has been little or no change in these parameters despite declining S deposition.

The apparent lack of proportional response in lake SO_4^{2-} concentrations in certain areas has been attributed to several mechanisms, including 1) re-oxidation of reduced S in the lakes' catchments followed by increasing flux to the lakes (Dillon and LaZerte, 1992; Schindler *et al.*, 1996; Dillon *et al.*, 1997; Dillon and Evans, 2001), 2) desorption of previously adsorbed SO_4^{2-} in the soil (Driscoll *et al.*, 1995; Löfgren *et al.*, 2001; Alewell, 2001; Wilander, 2001; Hill *et al.*, 2002) and/or 3) mineralisation and oxidation of organically bound S (Löfgren *et al.*, 2001; Mitchell *et al.*, 2001) from the catchments.

Superimposed on any long-term changes that might have occurred in the chemistry of the soils, precipitation and surface waters as a result of decreasing SO_2 emissions, is the influence of natural year-to-year and between-season variations in weather, particularly temperature and precipitation. These fluctuations induce variability in surface water chemistry (Monteith *et al.*, 2000) and may obscure the detection of trends that are expected as a result of reductions in anthropogenic SO_4^{2-} deposition. For example, if the time frame for the analysis of temporal response is short relative to the lag time of the lake or to weather variations, then natural fluctuations in lake water chemistry may limit the ability to detect patterns in many lake water quality parameters and to relate these patterns to decreased loading of acid pollutants (Bouchard, 1997). On the other hand, a longer term data set may obscure shorter trends occurring within the series. If the trends are not unidirectional, then the detection of trends is dependent on the time 'window' used in the analysis and can even make a significant difference regarding the detection of the presence and the direction of a trend (Clair *et al.*, 2002). For example, Clair *et al.* (1992) found significant decreases in SO_4^{2-} concentrations in lakes in Nova Scotia, Canada, for the period 1983-1989 but, with the addition of two more years of data, these trends were mostly reversed (Clair *et al.*, 1995).

The intra- and inter-annual fluctuations in weather are further complicated by the effects of longer term weather patterns i.e. climate and climate change. In North America, droughts in south and central Ontario (Dillon and LaZerte, 1992; Dillon *et al.*, 1997; Devito *et al.*, 1999), in northwestern Ontario (Schindler *et al.*, 1996) and in the upper midwest US (Webster and Brezonik, 1995), were hypothesised to cause re-oxidation of reduced S stored in wetlands, followed by subsequent increases in SO_4^{2-} net export in wet periods to downstream lakes. It was also suggested that the frequency of these droughts was

increasing (Schindler *et al.*, 1996).

Climate has long been recognised as an external factor that 'drives' the behaviour of lake and catchment biogeochemistry and often has a greater influence on the physical, chemical and biological properties of the lakes than do in-lake processes, catchment characteristics, or landscape position (Magnuson *et al.*, 1990; Webster *et al.*, 2000; George *et al.*, 2000; Baron and Caine, 2000). Few studies addressed the impact of known long-term climatic patterns or repeated events on surface waters although Anderson *et al.* (1996) related the dates of ice breakup in 20 Wisconsin lakes to the warm phase of El Niño Southern Oscillation (ENSO) episodes and George *et al.* (2000) found that winter temperatures in several Windermere catchment lakes were positively correlated to the North Atlantic Oscillation (NAO). These relationships reflect physical rather than chemical properties of the lakes. In addition, experimental manipulations at the ecosystem scale, such as the RAIN and CLIMEX projects (Wright and Jenkins, 2001), have shown that climate is a confounding factor in recovery from acidification.

Recently Evans *et al.* (2001) linked fluctuations in marine ion (Cl^- , Na^+ , Mg^{2+}) concentrations to the NAO; an inverse relationship between non-marine SO_4^{2-} and Cl concentrations at five of their lake sites in Great Britain suggested that soil retention and release of SO_4^{2-} also may be a 'sea-salt' related process, that might impede the detection of long term trends in lake recovery. In North America, Dillon *et al.* (1997) and Dillon and Evans (2001) reported a connection between S redox reactions and ENSO episodes. They found that elevated stream SO_4^{2-} concentrations and export rates in south-central Ontario occurred in the autumn of years with prolonged severe drought, and that these droughts occurred in the years following strong El Niño events, i.e. when the SOI was strongly negative. A link between precipitation patterns in the North American Great Lakes basin and the two extreme phases of the SOI, i.e. El Niño and La Niña, was established by Shabbar *et al.* (1997) who reported a distinct pattern of negative precipitation anomalies in the region during the first winter following the onset of El Niño events.

In this paper, the temporal coherence in SO_4^{2-} concentrations has been examined in a set of 8 lakes (9 basins in total) in south-central Ontario that have experienced significant reductions in sulphate deposition over the past two decades. Lake SO_4^{2-} concentration was chosen as a simple measure of the lakes' recovery following decreases in S deposition. Because the change in SO_4^{2-} concentration in these lakes has been substantially less than anticipated, the relationships between the observed patterns and external factors have been investigated, particularly climate, which

may have affected lake recovery. In a subsequent paper, changes in pH, alkalinity and base cation concentrations, parameters that are relevant to biological recovery, are evaluated.

Study area

The study lakes are located in the District of Muskoka and the County of Haliburton in central Ontario on a southern extension of the Precambrian Shield, near the southern limit of the Boreal ecozone (Fig. 1). Because the bedrock comprises sparingly soluble silicate minerals, and because the overburden is very shallow (typically less than 1 m, often less than 0.3 m), this region is very sensitive to acid deposition. Furthermore, it has received levels of S deposition that are among the highest in eastern Canada, with deposition levels of between 80 and 100 meq m⁻² yr⁻¹ at the start of this study. Bedrock in the region is primarily

granitised biotite and hornblende gneiss, with lesser areas of amphibolite and schist. The surface geology of the upland areas is dominated by basal tills and sand deposits; these are usually covered by peat in the lowland areas. Upland soils are primarily orthic humo-ferric and ferro-humic podzols and brunisols, while the lowland areas have humisols and humic mesisols. Bedrock outcrops are found in most catchments and small, shallow beaver ponds are found in half the catchments. Several of the catchments were partially cleared c. 100–130 years ago but the forests have regrown after farming was abandoned in the early 1900s. The economy of the region is controlled largely by activities related to recreation and tourism, with shoreline development extensive on two of the lakes (Harp, Dickie), light on three (Blue Chalk, Chub, Red Chalk) and absent on three (Heney, Plastic, Crosson). Selective forest harvesting has been practised periodically at several of the sites but there has been little in the past several decades. Upland

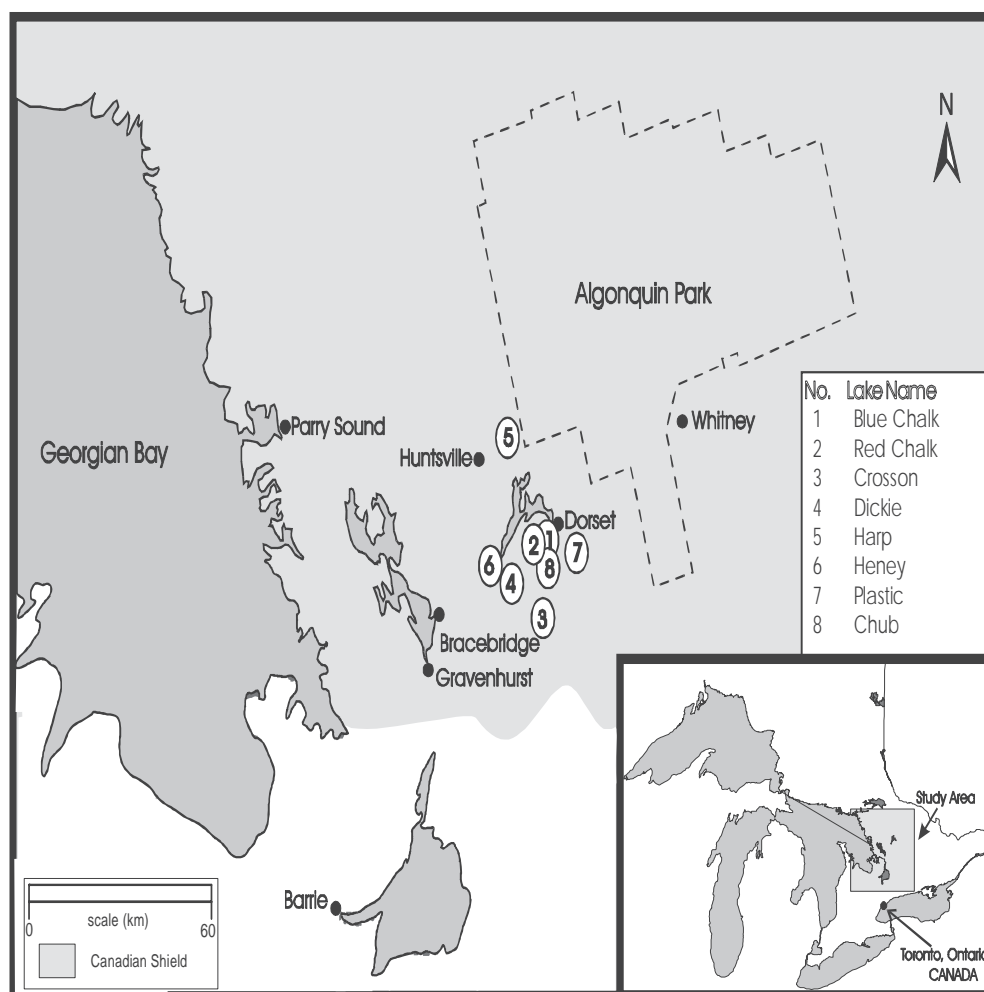


Fig. 1. Location of the study lakes, showing the Precambrian Shield boundary

portions of the lakes' catchments are dominated by sugar maple (*Acer saccharum* Marsh.) and red maple (*A. rubrum* L.) with some American beech (*Fagus grandifolia* Ehrh.), yellow birch (*Betula alleghaniensis* Britt.), red oak (*Quercus rubra* L.), eastern hemlock (*Tsuga canadensis* L.) and white pine (*Pinus strobus* L.). Nutrient cycling in the catchments of several of the lakes (Plastic, Harp, Chub, Red Chalk) has been studied extensively (Watmough and Dillon, 2001, 2003a,b,c). Over the past 20 years, the mean precipitation in the study area was 976 mm (30% as snow), with about 50–55% generating stream runoff, and the mean temperature was 5.3 degrees C.

Seven of the eight lakes are headwater lakes, while the eighth (Red Chalk) has one lake upstream; as a result, catchment to lake area ratios are relatively low for this region, and water replenishment rates are long, ranging from 1.6 to 5.7 years (Table 1). Because Red Chalk Lake has two distinct basins (East and Main) that differ chemically, and because the basin that is downstream (Main) receives only about 10% of its inflowing water from the upstream (East) basin, the basins have been treated as individual,

independent sites in the following analyses. The lakes are all dilute (conductivities of 22–35 $\mu\text{S cm}^{-1}$), with low levels of algal nutrients (TP of 4–14 $\mu\text{g L}^{-1}$), and with moderate concentrations of dissolved organic carbon (DOC 2–5 mg CL^{-1}) that depend on the relative importance of wetlands in each lakes' catchment. Both the sensitivity of the lakes to acid deposition and the degree of acidification of the study lakes varies substantially, from lakes such as Plastic and Heney that have very low alkalinities and documented biological damage (Dillon *et al.*, 1987), to others such as Harp and Blue Chalk that, because of their thicker surficial deposits, have higher alkalinities, base cations and pHs (Table 2).

Methods

LAKE SAMPLING

The lakes were sampled from 1976; however, as there were significant changes in the analytical methods used for SO_4^{2-} in 1979, data reported here for lakes begin in 1980. These

Table 1. Area (A_0), mean depth (z), drainage area (A_d), water replenishment time (T_w), conductivity, pH and DOC; chemical data are whole-lake, ice-free period, long-term averages for 1980–2001. Red Chalk morphometry data are from the whole lake, while chemistry data are from the larger basin, Red Chalk Main.

Lake	A_0 (ha)	z (m)	A_d (ha)	T_w (yr)	Cond ($\mu\text{S cm}^{-1}$)	pH	DOC (mg/L)
Blue Chalk (BC)	52.4	8.5	105.9	5.71	28.4	6.63	1.8
Chub (CB)	34.4	8.9	271.8	2.15	27.1	5.62	4.9
Crosson (CN)	56.7	9.2	521.8	1.67	25.2	5.57	4.2
Dickie (DE)	93.6	5.0	406.4	1.90	30.2	5.87	5.2
Harp (HP)	71.4	13.3	470.7	3.40	35.5	6.26	3.9
Heney (HY)	21.4	3.3	71.7	1.55	23.6	5.88	3.0
Plastic (PC)	32.1	7.9	95.5	4.04	21.4	5.75	2.2
Red Chalk Main (RC)	57.1	14.2	480.0	2.58	28.7	6.29	2.5

Table 2. Average SO_4^{2-} , total base cations (TBC) and alkalinity (measured by Gran titration) in 1980 and 1981, and in 1997 and 1998. All data in $\mu\text{equiv L}^{-1}$.

Lake	Alkalinity		Total Base Cations		Sulphate	
	1980/81	2000/01	1980/81	2000/01	1980/81	1997/98
Blue Chalk	77.8	87.9	234	215	148	112
Chub	13.6	18.8	217	179	161	117
Crosson	10.6	11.7	196	162	150	108
Dickie	18.2	34.5	210	316	143	105
Harp	62.7	69.4	281	311	174	126
Heney	5.4	22.3	183	167	155	92
Plastic	12.7	3.3	170	129	141	111
Red Chalk Main	67.3	67.7	240	206	148	116

changes in methods did not affect precipitation samples, or any other samples with low dissolved organic matter. Lakes were sampled, on average 6–8 times per year during the ice-free season (typically late April to the end of November). All samples were collected at the deepest location in the lake either from the upper 5 m of the water column during spring and fall overturn, or from the entire water column during the period of thermal stratification (June to October). In the latter case, samples collected from each 2 m interval were volume-weighted according to the relative contribution of that 2 m stratum to each thermal layer (epi-, meta-, hypolimnion) of the lake and combined for chemical analysis. This approach yielded volume-weighted samples representative of each layer; the analytical results for the three layers were then weighted according to the proportion of the whole lake volume that each thermal layer represented to give single, whole-lake, volume-weighted chemistry measures for each sampling date.

DEPOSITION SAMPLING AND METEOROLOGICAL PARAMETERS

Samples of bulk deposition for chemical analysis were collected in continuously open, 0.25 m², Teflon-coated collectors positioned in the study lakes' catchments and at the field laboratory. Between four and eight collectors located throughout the study area were used over the duration of this study (1976 to the present). Deposition was estimated daily at each collector site and, because all deposition collectors (and lakes) are within a 50 km radius, all collectors with data on any day were averaged to yield a daily deposition. These were summed to give monthly and annual depositions for the region. Wet-only deposition was also measured for several of the years of study and was typically 15–20% lower than bulk deposition. This difference was consistent with estimates of dry deposition made using measured SO₂ and particulate SO₄²⁻ data with estimated deposition velocities (Dillon *et al.*, 1988). Only bulk deposition data are reported here.

Indices describing the El Niño Southern Oscillation, the North Atlantic Oscillation and the Arctic Oscillation were used in the statistical analyses to determine whether these global-scale phenomena were related to observed changes in the chemistry of the study lakes. The Southern Oscillation Index is a standardised measure of the concurrent differences in atmospheric pressure across the South Pacific Ocean. The occurrence of El Niño is indicated by a negative SOI value for several consecutive months (NOAA, <http://www.cpc.noaa.gov/data/indices/soi>). The El Niño phase of the Southern Oscillation generally results in colder and wetter than average conditions on the west coast of North

America but warmer, drier conditions in central and eastern North America. The North Atlantic Oscillation Index is related to conditions over the Atlantic Ocean, and is calculated using sea level atmospheric pressure differences between Iceland and Gibraltar or the Azores. High positive values of the NAOI are strongly related to climatic conditions in Europe, with strong westerly winds, frontal precipitation and relatively high temperatures. Less is known about relationships with North American climate. Similarly, the Arctic Oscillation Index (AOI) relates to Arctic air masses and pressure differentials across that region. In addition to the annual values for these indices (the sum of the reported monthly values), annual data lagged by one and two years were used because earlier evidence indicated that droughts induced by the El Niño followed the onset of negative SOIs by a half year or more. In addition, seasonal values of the indices were used (each three-month period, with the summer and fall seasons being those from the previous year), and a seven-month period from September to March, as this window best captured the majority of El Niño events. In this way, each climate phenomenon created 8 summary indices. The SO₄²⁻ deposition data were also combined in an analogous fashion to yield eight measures corresponding to the same time steps as the climate indices.

ANALYTICAL METHODS

From 1976–79, SO₄²⁻ in deposition samples was measured colorimetrically (methyl thymol blue) and, from 1980 to 1999, sulphate was measured in both lake and deposition samples by ion chromatography following standard methods (Ontario Min. Envir., 1983). The two methods yielded virtually identical results for samples low in dissolved organic matter, such as deposition samples.

STATISTICAL ANALYSIS

Magnuson *et al.* (1990) evaluated temporal coherence among a set of Wisconsin lakes using Pearson product-moment correlations between their time-series data for all-possible pairs of lakes. Herein, temporal coherence has been estimated in SO₄²⁻ concentrations between lakes using the intraclass correlation from a repeated-measures (or randomised block) analysis of variance (ANOVA; see Donner, 1986; Somers *et al.*, 1996). The ANOVA was a two-way model without replication, with year as one factor incorporating the repeated-measures nature of the data and the pair of lakes as the second factor. The two-way interaction term in the ANOVA represents variation between lakes over years (i.e. interaction plus error given no replication). Here, lake is a blocking factor (i.e. this variation

is factored out) such that differences in the overall means for each lake are partitioned from the year-to-year variation and ignored. Since this blocking does not control differences in variances that are often correlated with the mean value, the variances for each time series were standardised to unity to ensure that the intraclass and Pearson correlations provide the same values, being analogous to using z-scored time series in either approach, but with slightly different degrees of freedom (Cronbach and Gleser, 1953). Both the Pearson and intraclass correlations measure the parallel or synchronous nature of the two time series but, relative to the Pearson correlation, the intraclass correlation is a more-general type of correlation that is not restricted to pair-wise comparisons. The intraclass correlation is better suited when the variables are responses (i.e. measured with error), and the intraclass correlation can accommodate missing data.

As in Magnuson *et al.* (1990), individual correlations summarise temporal coherence among pairs of lakes. The time series for 9 lake basins resulted in 36 correlations among all-possible pairs of lakes. Evaluating each of these correlations for significance presents a variety of statistical problems; alternatively an overall test of the entire matrix of between-lake correlations is available (Brien *et al.*, 1984) to determine if: (1) the matrix of between-lake correlations is homogeneous (i.e. all pair-wise correlations are equal and hence, all of the time series are synchronous); and (2) the average between-lake correlation is significantly different from zero. Because the full matrix of correlations failed the test of homogeneity (i.e. all of the time series were not synchronous), Rusak *et al.* (1999) were followed and Brien's test was used as an exploratory tool to identify homogeneous subsets of lakes by examining groups of lakes sequentially much as in cluster analysis. This analysis revealed two subsets of lakes with synchronous SO_4^{2-} patterns. One set included 6 lakes, and the second set 4 lakes; one lake was intermediate and could be included in either subset of lakes. The z-score SO_4^{2-} concentrations for each synchronous subset of lakes were averaged for each year to produce two summary time series, one for the subset of 6 lakes, and one for the subset of 4 lakes. Subsequently, multiple regression was used to determine which summary climate indices best predicted these two SO_4^{2-} time series.

For each climate index, simple correlation and multiple regression were used to identify the seasonal and annual summaries that best explained variance in the SO_4^{2-} patterns ($P < 0.10$). Because the various seasonal summaries are inter-correlated to varying degrees (i.e. redundant or multi-collinear; see Rencher and Pun, 1987), the resultant r , F and P values (and associated partial correlation statistics) were used to identify the best subsets of predictors for each of the SO_4^{2-} patterns. In the final assessment, the best predictors

for each of the climate indices were included in forward- and backwards- selection stepwise multiple regressions to determine which indices accounted for the greatest amount of the between-year variation in multi-lake SO_4^{2-} concentrations. Again, among-predictor redundancies were minimised to reduce problems with multi-collinearity in the multiple regressions.

Results and discussion

DEPOSITION TRENDS

Changes in SO_4^{2-} , NO_3^- and NH_4^+ deposition in the study area for parts of the present investigation period have been discussed in Dillon *et al.* (1988), Dillon *et al.* (1997), Dillon and Evans (2001), Eimers and Dillon (2002) and are summarised briefly here. Deposition from 1976 to 2000 is shown in Fig. 2(A). Sulphate deposition has declined by 53% from that measured in 1976–78, and 46% from that of 1980; most of the decrease occurred prior to 1986. Although there were several years in the 1990s with very low SO_4^{2-} deposition, these resulted from unusually low precipitation depths rather than reduced concentrations (Fig. 2B). N deposition is also reported here because it now exceeds S deposition in the study area, although NO_3^- and NH_4^+ deposition (and concentration) have changed little in 25 years, with no significant downward trends. Total inorganic nitrogen ($\text{NO}_3^- + \text{NH}_4^+$) deposition in the study area now exceeds SO_4^{2-} deposition by a substantial margin. At present, most of the inorganic nitrogen is retained in lakes (Kaste and Dillon, 2003) and catchments (Watmough and Dillon, 2003b,c), although in a few catchments NO_3^- leakage is relatively high (Schiff *et al.*, 2003).

These decreases are consistent qualitatively with the change in S emissions in eastern North America. Emissions of SO_2 in the U.S.A. dropped from an estimated 17.3 million tons in 1980, the baseline year for the Clean Air Act amendments of 1990, to 15.7 million in 1990 followed by 10.6 million in 2001, a total drop over 21 years of 39% (Stoddard *et al.*, 2003). A further drop of 3.3 million tons is scheduled for 2010, bringing the total decrease to 10 million tons or almost 60%. In Canada, SO_2 has declined from 3.81 million tonnes in 1980 to 2.52 million in 1990, 1.74 million in 1996 and 1.25 million in 2000, a drop of 68% over 20 years. Thus, the decrease in measured deposition of SO_4^{2-} is consistent with the changes in SO_2 emissions in eastern North America as a whole.

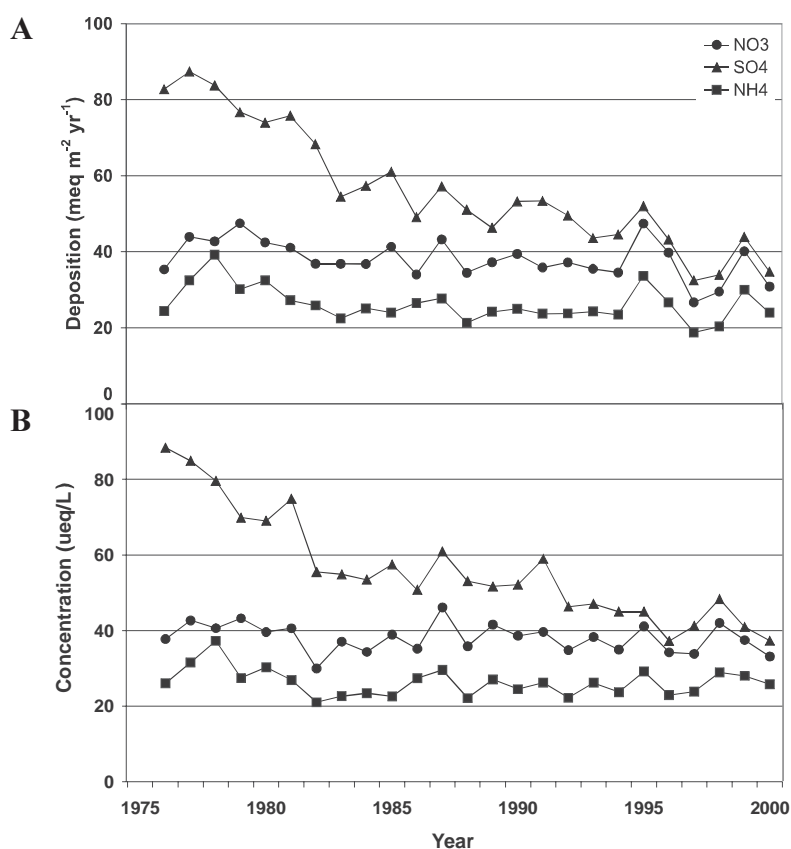


Fig. 2. Annual bulk deposition (top panel) and volume-weighted concentration in deposition (bottom panel) of SO_4^{2-} , NO_3 and NH_4 in the study area.

CHANGES IN LAKE SO_4^{2-} CONCENTRATION

Although the 8 lakes are within a 50 km radius and SO_4^{2-} deposition was uniform in this region, the SO_4^{2-} concentrations differed between lakes by up to $33 \mu\text{eq L}^{-1}$ (range 141 to $174 \mu\text{eq L}^{-1}$) at the start of the study and by $34 \mu\text{eq L}^{-1}$ (range 92 to $126 \mu\text{eq L}^{-1}$) at the end of the study (Fig. 3A). The decreases in concentration over the 19-year duration of this study ranged between 21 and 41% (Table 2). The overall decline from 1980 to 1998 averaged 27%, which was substantially less than the decline in deposition over the same period. The biggest decline in SO_4^{2-} concentration (Heney) was similar in magnitude to the decrease in SO_4^{2-} deposition, while the decline in SO_4^{2-} in all the other lakes was substantially less than that in deposition.

Changes in lake SO_4^{2-} concentrations that result from a change in deposition and subsequent input rates or mass balances including catchment sources are unlikely to be instantaneous; rather, they will reflect the water replenishment rate (a measure of the hydrological turnover time) of each lake. If SO_4^{2-} behaviour in the lakes is conservative (the mean mass transfer coefficient for SO_4^{2-}

in these lakes based on mass balance measurements over 18 years was not significantly different from 0.0 m yr^{-1} ; P. J. Dillon, unpub. studies), then, based on a simple mass balance model, the half-life for change in lake concentration is $(\ln 2/\text{replenishment time})$. Therefore, in three half-lives, the concentration would have changed *c.* 90% of the way to the new steady-state level, i.e. in 4.7 to 17 years for these lakes. Thus, lakes with a short water replenishment time such as Heney should respond more quickly to decreasing inputs than those with long replenishment times, e.g. Blue Chalk (Table 1). However, because almost all of the change in SO_4^{2-} deposition took place pre-1986, the current lake SO_4^{2-} concentrations should almost fully reflect the expected changes in all of the lakes if there are no lags in response.

A second observation is that the trends in SO_4^{2-} concentration were not monotonic, nor did they match the temporal changes in SO_4^{2-} deposition. During certain time periods such as 1983–84 and 1988–92, SO_4^{2-} in all of the lakes, in fact, increased rather than declined. When the lake SO_4^{2-} concentrations were standardised (i.e., z-score transformed — Fig. 3B), it was apparent that the pattern in all lakes was consistent. Sulphate concentration declined

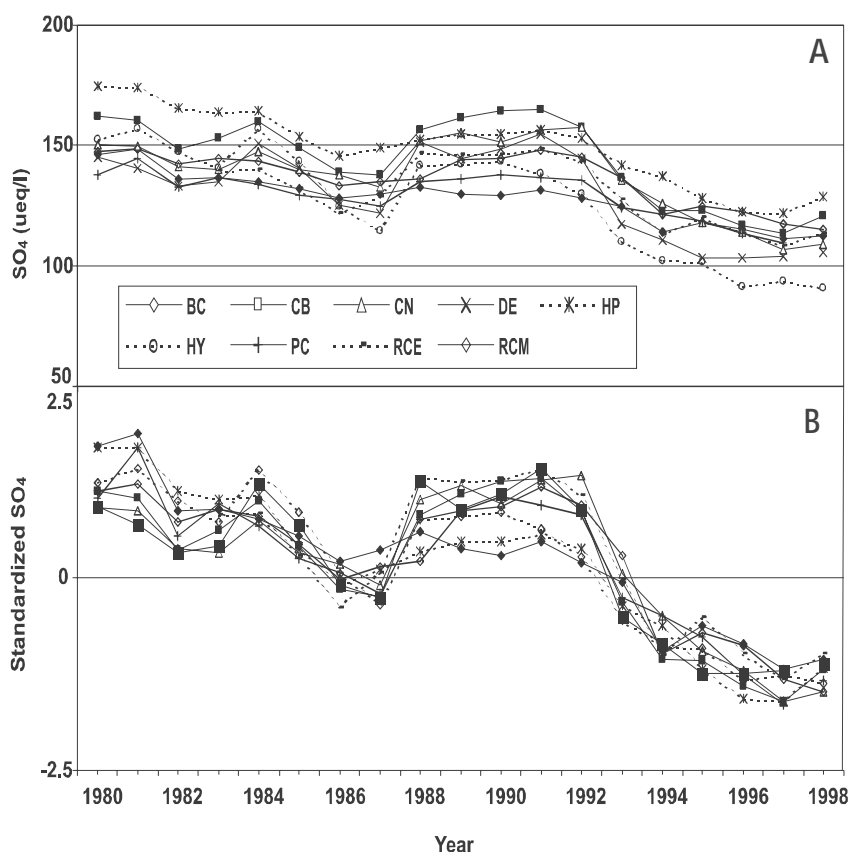


Fig. 3. (A) Average annual SO_4^{2-} concentration during the ice-free period (May–November) in the study lakes, and (B) z-score transformed average annual ice-free SO_4^{2-} concentration.

and increased in all lakes at the same times over the 19 years, and the patterns were independent of the water replenishment time of the lakes. This synchrony in patterns was independent of the physical or chemical characteristics of the lakes or their catchments; although differences in these factors may well have caused the differences in initial concentrations of SO_4^{2-} in these lakes.

Using coherence analysis to pick the subsets of lakes with patterns that were not significantly different, it was determined that Red Chalk East, Chub, Dickie, Red Chalk Main, Crosson and Plastic Lakes formed a set with a common temporal SO_4^{2-} pattern, while Harp, Blue Chalk, Heney and Plastic Lakes formed a second set. Since the lake selection was done with replacement, it was possible for a lake to belong to more than 1 subset; however, Plastic Lake was the only instance where this occurred. Three of the four lakes in the second group have the longest water replenishment times, suggesting that their response to changes in deposition or other factors may become evident at a slower rate, i.e. that the response of the second group may be similar to, but less than that of the first group.

The mean of the z-scored concentrations of the lakes in each grouping were used in the subsequent development of predictive models. Although this analysis generated two groups of lakes, the two SO_4^{2-} concentration patterns were very similar (Fig. 4). As noted above, the similarity in the patterns may simply reflect a delay in the timing of the response between the two subsets of lakes. In addition, the fact that Plastic Lake was intermediate between the two subsets suggests that a gradient in response actually exists. Alternatively, different climatic factors may be involved.

FACTORS AFFECTING THE SO_4^{2-} CONCENTRATION PATTERNS

The SO_4^{2-} pattern of each of the subset of lakes demonstrated increases in specific periods, e.g. 1983–84, 1988–92, etc. Dillon and Evans (2001) reported increases in lake SO_4^{2-} concentration after each El Niño, i.e. following periods where the SOI was negative for 6 or more months (Fig. 5). Previously, this has been explained, qualitatively, in the cases of Plastic and Heney Lakes (Dillon *et al.*, 1997; Dillon and

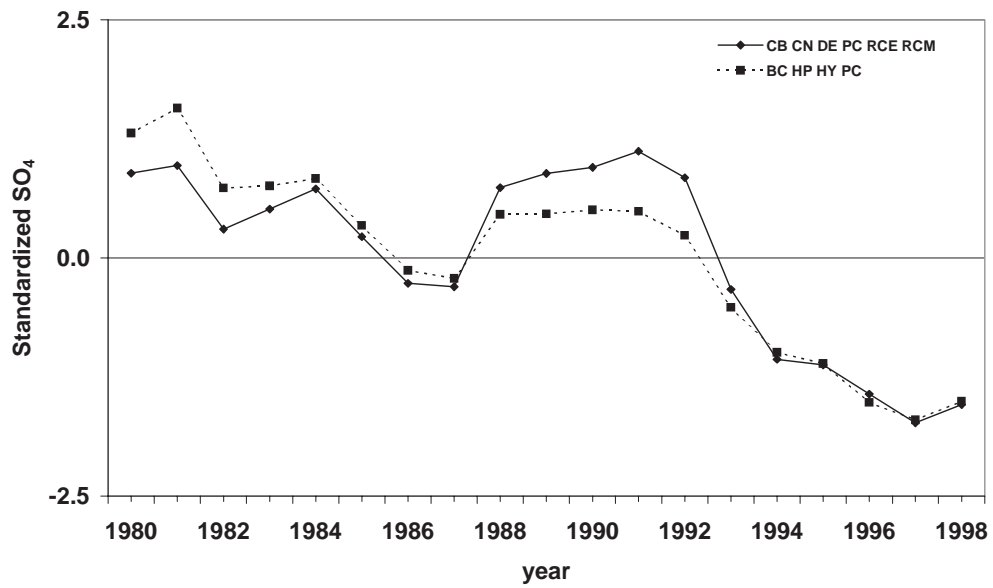


Fig. 4. Trend in z-scored SO_4^{2-} concentration for the 2 lake groups. Group 1 includes Chub, Crossson, Dickie, Plastic, Red Chalk East and Red Chalk Main, while group 2 includes Harp, Blue Chalk, Heney and Plastic Lakes.

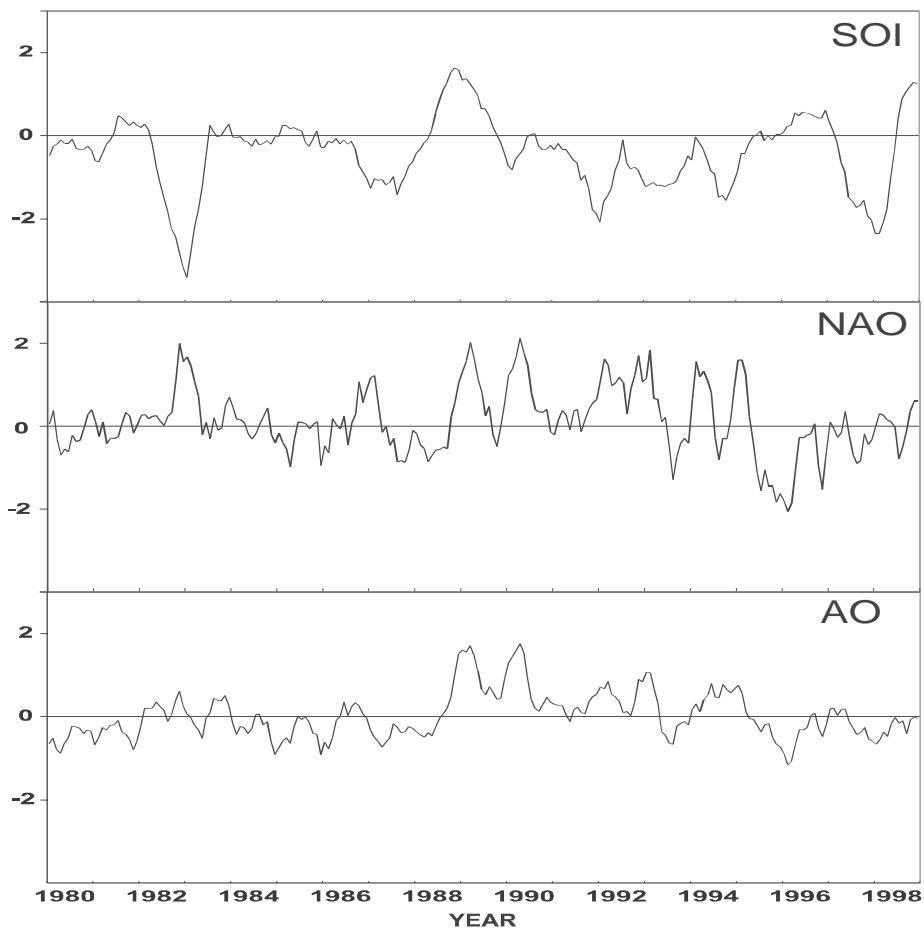


Fig. 5. Southern Oscillation Index (SOI), North Atlantic Oscillation Index (NAOI) and Arctic Oscillation Index (AOI) (standardised values). Lines represent 6-month moving averages of monthly data.

Evans, 2001; Eimers and Dillon, 2002) on the basis of drought-induced re-oxidation of previously stored S and/or enhanced mineralisation of S-containing material that was subsequently mobilised from the catchments into the lakes as SO_4^{2-} . The results reported here are consistent with this interpretation. Although it might be expected that the North Atlantic and the Arctic Oscillations are linked to the ENSO, correlations of the NAOI and AOI with the SOI were weak in almost all cases. The strongest correlation coefficient between any SOI and any of the NAOI or AOI was only 0.47 (the SOI lagged by 1 year with the spring AOI), and virtually all correlations were not significant. On the other hand, almost half (24 of 64) of the correlations between NAOI and AOI were significant.

Because there were only 19 years of lake chemistry data, the predictor variables in each suite were examined separately to avoid over-parameterisation of the models. In both cases, i.e. for each lake group, good models using the SOI, the NAOI or the SO_4^{2-} deposition resulted (Table 3). Using only the SOI, significant contributors to the model included the annual SOI, with both one-year and two-year lags, as well as the spring and summer SOI. The same four SOI indices were chosen for both lake groups, suggesting a common response to the El Niño phenomenon. Although there is, obviously, some degree of collinearity between these four SOI indices, the modelling procedure takes this into consideration and includes additional variables only after the most strongly correlated variables have been selected, and only if they explain significantly more of the variance. It is noteworthy that the indices chosen that are related to the ENSO all represented atmospheric conditions prior to the period when the lake measurements were made, with data up to two years earlier being important (annual SOI lagged two years). These results are consistent with the hypothesis that the response of the lakes is affected by the ENSO.

When the suite of NAO indices was used, significant models based on the unlagged annual index plus the fall index (lake set 1) and the winter, fall and seven-month indices (lake set 2) were produced. These were weaker (lake set 1) or comparable (lake set 2) to the models based on the SOI suite (Table 3). The suite of AO indices yielded no significant models despite the many significant correlations between individual NAO and AO indices.

The suite of SO_4^{2-} deposition indices was also tested separately. Since both lake SO_4^{2-} concentration and SO_4^{2-} deposition declined over the duration of the study, and since other studies had reported that declining lake levels were related to declining deposition, it was thought that changes in deposition would have been at least partially responsible for the measured changes in lake chemistry. Significant

Table 3. SO_4^{2-} predictions (using backwards elimination stepwise multiple regression) for each of the two groupings of lakes based on suites of parameters

Predictor suite	r	F	p	Important predictors
6-lake set				
SO indices	0.71	3.63	0.03	annual (lag1), annual (lag2), spring, summer (lag1)
NAO indices	0.57	3.92	0.04	annual (lag0), fall (lag1)
AO indices				none
SO_4^{2-} deposition	0.72	18.1	0.0005	fall (lag1)
combined	0.83	7.61	0.002	spring SOI, annual NAOI (lag 1), fall NAOI, fall SO_4 deposition
4-lake set				
SO indices	0.63	2.34	0.1	annual (lag1), annual (lag2), spring, summer (lag1)
NAO indices	0.64	3.5	0.04	7month, winter, fall (lag1)
AO indices				none
SO_4^{2-} deposition	0.78	25.7	0.0001	fall (lag1)
combined	0.91	7.85	0.002	spring SOI, annual SOI (lag2), winter NAOI, fall NAOI, 7-month NAOI, winter AOI, fall SO_4^{2-} deposition

fall = average for Sept-Nov; winter = average for Dec-Feb; spring = average for Mar-May; summer = average for Jun-Aug; ann (lag0) = annual average for same year; ann (lag1) = average value for preceeding year; ann (lag2) = annual average 2 years prior

models for both lake sets resulted, in each case with the SO_4^{2-} deposition in the previous fall period being the only predictor. For lake set 1, this model was comparable in predictive ability to the SOI-based model, while for lake set 2, it was somewhat better (Table 3). Models that were almost as good as these models based on the fall deposition could be built using the SO_4^{2-} deposition measurements for any of the time periods but, because of the very strong inter-correlation between the depositions measured over each of the time intervals, only one parameter was needed.

The subset of parameters from each of the four suites used to generate the suite-specific models were combined to build a model based on both SO_4^{2-} deposition and climate-related parameters. This allowed a smaller, more reasonable total

number of parameters to be used (7 identified in Table 3 for lake set 1, 8 for lake set 2). To each model set was added the AOI parameter (winter AOI) that had the strongest correlation with the lake SO_4^{2-} because backwards stepwise multiple regression procedures may include pairs of variables that are not significant individually but, jointly, account for significant amounts of the total variation. The backwards stepwise regression yielded better models (Table 3) than those based on single predictor suites, with 69 and 83% of the variance in lake SO_4^{2-} being explained for lake sets 1 and 2, respectively. In both cases, SOI, NAOI and SO_4^{2-} deposition measures were included in the overall model.

These results indicate clearly that the response of lakes to declining SO_4^{2-} deposition is controlled by changes in both deposition and climate. These climate factors are regional if not global in scale, rather than dependent simply on local variations in meteorological variables. They may act by influencing local climate or weather but almost certainly have broad regional impacts, as phenomena such as the ENSO result in perturbations on scales greater than continental.

There is evidence that the ENSO results in changes in hydrology through altered precipitation and temperature regimes in eastern North America (Shabbar *et al.*, 1997). This is consistent with drought-induced changes to S cycling, including redox and mineralisation processes. Reduced precipitation and higher temperature both favour lower water tables in the wetland portions of lakes' catchments; these, in turn, can result in re-oxidation of previously-stored reduced S species (Schindler *et al.*, 1996). Since re-oxidation is an acid-producing process and since the product (SO_4^{2-}) is much more soluble than the initial reduced species, this mechanism can effectively buffer downstream lake SO_4^{2-} concentrations. Similarly, mineralisation of organic matter in the lakes' catchments, with release of soluble SO_4^{2-} (Löfgren *et al.*, 2001) is favoured by higher temperatures. The third SO_4^{2-} buffering mechanism, desorption of previously adsorbed SO_4^{2-} from the soil (Alewell, 2001), is likely to be more dependent on changes in S deposition than on climate parameters. These mechanisms that relate the climate indices to the measurements of lake response are currently being explored.

Summary

Temporal coherence can be defined as the correlation or synchrony between time series. When temporal coherence in SO_4^{2-} concentrations between nine lake basins was estimated using the intraclass correlation from a repeated-measures analysis of variance, two subsets of lakes with

synchronous patterns were found. The difference between the subsets' patterns is most likely related to the response time of the lakes in each subset; one subset (4 lakes) included those with the longest water replenishment time, which are expected to respond less to decreases in SO_4^{2-} deposition. The response of the lakes to the decrease in SO_4^{2-} deposition over two decades was independent of the degree of acidification or sensitivity to acidification of the lakes. The factors which best predicted the temporal pattern of SO_4^{2-} in the lakes were regional/global-scale climate indices, specifically the Southern Oscillation Index describing the ENSO and the North Atlantic Oscillation Index, and SO_4^{2-} deposition indices; each produced good individual predictive models. When the predictor variables were combined, the best models describing the long-term changes in lake SO_4^{2-} concentration included the SOI, NAOI and SO_4^{2-} deposition. Thus, large-scale climate factors play a major role in determining the response of aquatic systems to changes in SO_4^{2-} deposition and should be considered when predicting and interpreting the response of lakes to changes in SO_4^{2-} deposition.

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