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Nitrogen dynamics in runoff from two small heathland catchments representing opposite extremes with respect to climate and N deposition in Norway

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Abstract

Effects of contrasting climatic conditions and nitrogen (N) deposition levels on streamwater N dynamics are assessed at two small heathland catchments; Dalelva in northern Norway (69°N) and Øygard in southwestern Norway (58°N). The study comprises 11 years of data on climate, hydrology and N inputs/outputs from Dalelva and 8 years of corresponding data from Øygard. Both sites are comparable in catchment size, geology and land cover characteristics, but have large differences in climate and N deposition. Dalelva is characterised by a cold, arctic climate and low N deposition (2–3 kg N ha⁻¹ yr⁻¹), whereas the Øygard site has a more mild, humid climate with much larger N deposition (13–19 kg N ha⁻¹ yr⁻¹). Streamwater nitrate (NO₃⁻) concentrations at Dalelva generally were negligible during the growing season, but showed a steady increase during the dormant season until a maximum of 40–100 µg N L⁻¹ was reached just before snowmelt. At onset of the snowmelt flood, NO₃⁻ concentrations decreased momentarily to very low levels, suggesting that N eluted from the seasonal snowpack to a great extent was infiltrated and immobilised in the soils. At Øygard, flood peaks occurred frequently during all seasons, and usually there was no distinct spring flood. A lack of clear dilution effects from floods on streamwater NO₃⁻ concentrations may indicate a relatively high NO₃⁻ leaching potential in this catchment. On average, the annual NO₃⁻ export was negligible at Dalelva (<0.1 kg N ha⁻¹ yr⁻¹), while at Øygard it amounted to 3.0±0.3 (±1 s.d.) kg ha⁻¹ yr⁻¹, or nearly 20% of the annual N deposition. In addition to this relatively high annual N loss, elevated NO₃⁻ concentrations during the growing season further indicate that the N supply at Øygard is in excess of the combined plant and microbial demand.

Key words: catchments, surface water, nitrogen deposition, nitrate leaching, climate, hydrology, snowmelt, Dalelva brook, Øygard brook

Introduction

Nitrogen (N) levels and dynamics in streamwater are regulated by interactions between several abiotic and biotic factors operating in the terrestrial catchment and in the aquatic environment (Vitousek *et al.*, 1997). Among the abiotic factors, both climate (air temperature, precipitation, snow accumulation) and N deposition level have a great impact on N pools and transformation processes that subsequently determine concentrations and fluxes of N in surface waters. Several studies have demonstrated a close relationship between N deposition level, soil N status and NO₃⁻ leaching (Stoddard, 1994; Dise and Wright, 1995; Tietema *et al.*, 1998). In addition, increased air- and soil

temperatures caused by a possible climate change in the future may further accelerate NO₃⁻ leaching by increasing the mineralisation of organic matter (Wright, 1998).

Norway spans a large north-to-south gradient with respect to both climate and N deposition. The two small, calibrated heathland catchments Dalelva (69°N) and Øygard (58°N) represent opposite extremes, with cold climate/low N deposition at the northern site and warmer, more humid climate and higher N deposition at the southern site. Both catchments are monitored within the framework of the Norwegian monitoring programme on long-range transported air pollutants (LRTAP), Dalelva since 1988 and Øygard since 1992 (Johannessen, 1995). By having very

similar catchment characteristics, these two sites appear to be quite suitable for assessing the effects of different climate and N deposition on such ecosystems.

To date, the monitoring and research activities at Dalelva have mainly focused on acidification problems caused by sulphur (S) deposition from Russian industry located near the Norwegian border. In this context, Wright and Traaen (1992) have reconstructed the acidification history at site and made predictions of future acidification applying the MAGIC model together with soil monitoring data aggregated by Reuss (1990). During 1992–1995, Øygard was one of several calibrated catchments included in the multidisciplinary research project “Nitrogen from mountains to fjords” (Henriksen and Hessen, 1997). Among the activities connected to the Øygard site during this period were N input/output studies of N (Kaste *et al.*, 1997), assessment of N contribution to acidification (Henriksen *et al.*, 1997) and short-term, experimental P additions to streamwater (Hessen *et al.*, 1997). Since 1999, several sub-catchments within Øygard have been intensively studied as part of the research project “N retention and acidification in mountains and heathlands” (Wright *et al.*, 2001).

Both Dalelva and Øygard (as part of the larger Bjerkreim river system) have recently been included in a joint European project intending to adapt the process-oriented INCA model (Integrated Nitrogen model for CAtchments) (Wade and

Whitehead, 2002) to key ecosystem types throughout 8 countries. By linking hydrology with the microbial processes controlling N behaviour in soils and river reaches, the model simulates the contribution of multiple sources to catchment N pools and river nitrate (NO_3^-) and ammonium (NH_4^+) concentrations.

The main objective here is to assess the effects of contrasting climatic conditions and N deposition on streamwater N dynamics at two sites that are comparable in catchment size, geology and land cover characteristics. Several elements from this assessment will be highly relevant for forthcoming parameterisation, calibration and application of the INCA model, a major theme of this special issue of Hydrology and Earth System Sciences.

Material and methods

STUDY SITES

The study sites are located in diametrically opposite parts of Norway, Dalelva at 69°45'N and Øygard at 58°35'N (Fig. 1). Nevertheless, their catchment characteristics are comparable in many respects (Table 1). Both Dalelva and Øygard have relatively small (3.2 and 2.6 km², respectively) and undisturbed catchments dominated by heathland and mountains. Dalelva is a small mountain to fjord catchment,

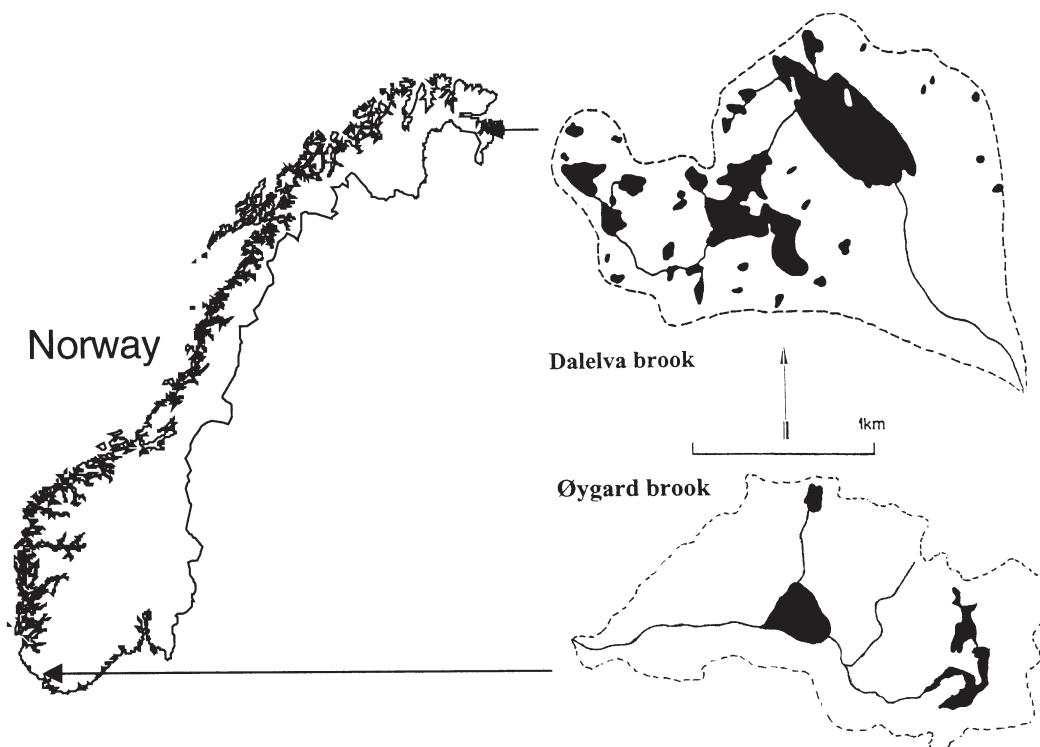


Fig. 1. Map of the catchments, showing the location in Norway. Measuring weirs and measuring stations are located at the outlet of each catchment

Table 1. Site characteristics. Climatic data are mean values (± 1 s.d.) during 1990–2000 at Dalelva and 1993–2000 at Øygard. Streamwater concentrations of total organic carbon (TOC) and hydrogen ions (H^+) are flow-weighted.

Parameter	Unit	Dalelva	Øygard
Catchment area (A_c)	km ²	3.2	2.55
Altitude	m.a.s.l.	0–241	185–544
Mean slope	°	5	8
Land cover:			
- Forest	% of A_c	20	4
- Heath/mountains	“	61	83
- Peat	“	4	6
- Lakes	“	15	7
Climate			
- Air temperature	°C	0.2 \pm 0.8	6.7 \pm 0.6
- Snow season	days yr ⁻¹	214 \pm 13	58 \pm 17
- Snow depth	m	0.50 \pm 0.09	0.11 \pm 0.04
- Precipitation	m yr ⁻¹	0.629 \pm 0.073	2.146 \pm 0.357
Water chemistry:			
- TOC	mg L ⁻¹	3.8 \pm 0.3	1.4 \pm 0.2
- H^+	μ eq L ⁻¹	1.7 \pm 0.4	9.1 \pm 2.0

whereas Øygard is located about 20 km from the estuary of the Bjerkreim river system (685 km²). The bedrock geology in the Dalelva catchment consists of mica schist and micaceous gneiss covered by glacial sediments of similar lithology (Wright and Traaen, 1992). Soils are podsollic, thin and patchy in the upper parts of the catchment. Vegetation comprises birch forest up to about 150 metre elevation with heathlands and moorlands above. About 15% of the Dalelva catchment is covered with lakes. At Øygard, the bedrock geology consists of Precambrian migmatites, banded gneisses and anorthosite covered by glacial sediments (Maijer and Padget, 1987). Soils are thin and patchy, especially in the upper parts of the catchment. The catchment consists largely of heathland with grasses, moors and minor areas of bare rocks or birch forest (Hessen *et al.*, 1997). Approximately 7% of the total catchment area at Øygard is made up of several small lakes (Fig. 1). Even though the fraction of peatland is relatively uniform at the study sites (4% at Dalelva and 6% at Øygard), the average total organic carbon (TOC) concentration in surface water at Dalelva is nearly three times the level at Øygard (Table 1).

Both catchments have atmospheric deposition as the only external N source (besides N fixation). With an average wet plus dry N deposition of 2.6 \pm 0.4 kg N ha⁻¹ yr⁻¹ during 1990–2000, Dalelva characterises an area that is near natural background levels (Table 2). In contrast, Øygard is located

in the part of Norway receiving the highest amount of long-range transported air pollution (Tørseth and Semb, 1997). Average N deposition at this site during 1993–2000 (16.8 \pm 2.6 kg ha⁻¹ yr⁻¹) was approximately six times higher than at Dalelva. The mean ratios of oxidised N to reduced N in atmospheric deposition at Dalelva and Øygard were 0.93 and 1.31, respectively. In general, the contribution of dry N deposition to total N deposition was much higher at Dalelva (54% during 1990–2000) than at Øygard (16% during 1993–2000). This great difference in deposition mechanisms may be explained largely by the low precipitation amounts in the northern catchment (SFT, 2001). Even though N deposition is quite low at Dalelva, the levels of S deposition are quite high (6–9 kg S ha⁻¹ yr⁻¹) due to large emissions from Russian industry located near the Norwegian border (SFT, 2001). Despite this, the annual mean hydrogen ion (H^+) concentration at Dalelva is relatively low compared to the more chronic acidified Øygard brook (Table 1).

In addition to very different N deposition levels, the sites are also characterised by highly different climates in terms of air temperature, precipitation amounts and snow accumulation (Table 1). During the periods of study, the mean air temperatures were close to 0°C in the Dalelva area and nearly 7°C in the Øygard area. Further, the mean precipitation amounts around Dalelva (0.629 m yr⁻¹) were less than one-third the levels around Øygard (2.146 m yr⁻¹). Both the mean air temperatures and the corresponding precipitation amounts were close to or slightly higher than

Table 2. Annual mean concentrations of NO₃⁻ and TN in the streams (flow-weighted).

	Dalelva		Øygard	
	NO ₃ ⁻ μ g N L ⁻¹	TN	NO ₃ ⁻ μ g N L ⁻¹	TN
1990	9	141	-	-
1991	6	144	-	-
1992	13	161	-	-
1993	16	187	168	282
1994	9	125	161	257
1995	11	134	168	252
1996	10	151	189	320
1997	14	135	125	295
1998	12	133	135	228
1999	11	133	159	264
2000	9	154	123	209
Mean	11	145	154	263
s.d.	3	17	23	36

the 1961–1990 normal (DNMI, 2001). Air temperatures above 0°C are rare during winter at Dalelva, and the catchment is usually covered with snow for 6–7 months of the year, ranging from the middle of October to late May. On average, nearly 45% of the annual precipitation at Lanabukt, near Dalelva, accumulated as snow during 1990–2000. The extensive snow accumulation has a heavy impact on the seasonal runoff pattern, which is characterised by very low flow during winter and a distinct snowmelt flood in May–June. At the Øygard catchment, the winters are usually mild and snow accumulation is often a transient phenomenon (Fig. 2). At Helleland, near Øygard, only 9% of the annual precipitation accumulated as snow during 1993–2000. The mild winters, in combination with large precipitation amounts, generate a highly variable runoff pattern with frequent flood episodes that may occur at all seasons. In contrast to Dalelva, there is usually no distinct spring flood following snowmelt at Øygard.

METHODS AND DATA SOURCES

Streamwater chemistry

Water chemistry and stream flow have been monitored at Dalelva and Øygard since 1988 and 1992, respectively. Most data are collected within the framework of the Norwegian monitoring programme on long-range transported air pollutants (LRTAP) (Johannessen, 1995). At Øygard, however, the first four years of data come from the research project “Nitrogen from mountains to fjords” (Henriksen and Hessen, 1997). Since 1990 at Dalelva and 1993 at Øygard, chemical samples have been collected weekly and fortnightly, respectively. The samples have been analysed unfiltered at NIVA for several parameters, of which nitrate (NO_3^-), ammonium (NH_4^+), total nitrogen (TN), pH, and total organic carbon (TOC) are presented in this paper. NO_3^- and NH_4^+ were analysed by automated colorimetry, NO_3^- after reduction with cadmium amalgam, NH_4^+ after reaction with hypochlorite and phenol, and TN after oxidation with

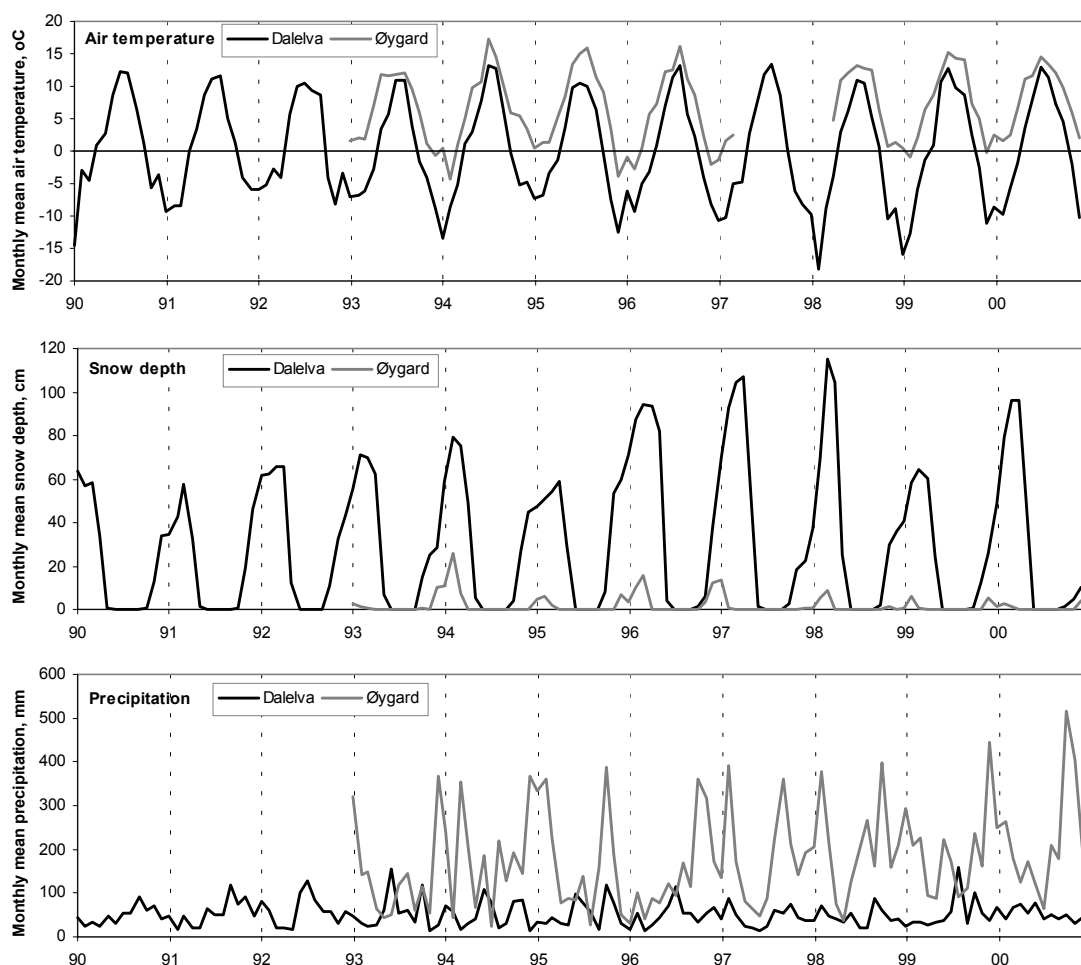


Fig. 2. Monthly mean air temperatures, snow depths and precipitation volumes

peroxodisulphate; TOC was analysed by spectrophotometry after UV oxidation (Wright, 1998). NO_3^- was actually analysed as the sum of NO_3^- and NO_2^- , but the latter species is quickly oxidised to NO_3^- and is usually taken as insignificant in natural surface waters. Detection limits for the NO_3^- , NH_4^+ , TN and TOC analyses are 1, 5, 10 $\mu\text{g N L}^{-1}$ and 0.2 mg C L^{-1} , respectively. Among the N species, NH_4^+ was only analysed in 1988, 2000 and during the first half of 2001. Altogether, NH_4^+ concentrations were determined in a total of 35 and 22 samples at Dalelva and Øygard, respectively. Due to the limited NH_4^+ data, total organic N (TON) was estimated from TN minus NO_3^- and the mean of observed NH_4^+ concentrations. This probably introduces a relative small error to the annual mean TON estimates, in that NH_4^+ concentrations generally were low and showed little variation throughout the year (95 percentile $\leq 13 \mu\text{g N L}^{-1}$ at both sites). Fluxes of NO_3^- and TN were calculated for each day using observed or estimated water discharges at the stream outlets and the corresponding chemical data interpolated from weekly or fortnightly samples.

Atmospheric N deposition

The atmospheric N deposition data are obtained from the nearest monitoring stations included in the LRTAP programme (Johannessen, 1995). The data comprise wet and dry deposition of total inorganic N (TIN), which is the sum of NO_3^- and NH_4^+ . Wet deposition is determined by the chemical composition and amount of precipitation, whereas dry deposition is estimated from ambient air concentrations and dry deposition velocities for individual components (Tørseth and Semb, 1998).

At Dalelva, wet deposition estimates are from the stations Karpdalen (1991–1997) and Karpbukt (1999–2000), which are located 3.5 and 2 km south of the Dalelva outlet, respectively. Missing data in 1990 and 1998 are substituted with corresponding data from the monitoring station Svanvik, which is located 28 km south of Dalelva. Parallel monitoring during 1991–1997 demonstrated that mean TIN deposition amounts at Svanvik deviated less than 10% from those at Karpdalen. In the absence of dry deposition estimates at Karpdalen/Karpbukt, corresponding data from Svanvik are employed for the whole period.

At Øygard, the wet deposition data are obtained from the monitoring station Ualand (15 km southeast of the catchment), whereas dry deposition estimates were obtained from Skreådalen (50 km to the northeast). Due to the relatively long distance between the study site and the monitoring stations, the N deposition estimates at Øygard probably are more uncertain than at Dalelva. Through application of an interpolation routine, Tørseth and Semb

(1997) estimated a precipitation amount and TIN deposition field for the Bjerkreim river catchment. The fields were based on runoff data from 44 sub-catchments (NVE, 1994), precipitation amounts from eight local monitoring sites, and precipitation chemistry from Ualand, Skreådalen, and two temporary monitoring sites located within the main catchment. According to their estimate, annual precipitation at Øygard during 1993–95 was 8–24% (mean 18%) higher than at Ualand, and the annual TIN deposition was 12–23% (mean 19%) higher than at Ualand (including dry deposition equivalent to Skreådalen). The examination demonstrated that the spatial variation in chemical composition was much less than the spatial variation in precipitation amounts. Hence, to obtain more realistic TIN deposition estimates for Øygard during the whole period 1993–2000, wet deposition data from Ualand are scaled up 18%, according to the relative difference in precipitation amounts during 1993–95.

Meteorological and hydrological data

Data on daily air temperatures, precipitation and snow cover are obtained from the nearest monitoring stations operated by the Norwegian Meteorological Institute (DNMI, 2001). At Dalelva, air temperatures are from the station Kirkenes (18 km west of the catchment), whereas precipitation and snow depth data are obtained from the station Lanabukt (7 km to the northeast). At the Øygard site, air temperature data are collected from the monitoring stations Ualand (1993–1997; 15 km southeast of catchment) and Eik-Hove (1998–2000; 24 km to the southeast). Precipitation and snow depths are obtained from the monitoring station Helleland, which is located 10 km south of the Øygard catchment. According to estimates made by Tørseth and Semb (1997), the annual precipitation at the Øygard catchment during 1993–95 was 9% higher than at the monitoring station Helleland.

Both Dalelva and Øygard have weirs at the outlet, where water discharge is monitored continuously by pressure transducers and loggers operated by NIVA. For quality assurance, the runoff data are compared to corresponding data from two adjacent gauging stations operated by the Norwegian Water Resources and Energy Directorate (NVE, 2001). These are Karpelva (124 km^2), located 3 km south of the Dalelva outlet, and Hetland (69.7 km^2), which has its headwaters about 5 km west of the Øygard outlet. In cases where data are missing or the runoff values seem erroneous, scaled data from the NVE stations are inserted. The scaling was accomplished according to catchment size and the 30-year normal runoff (m yr^{-1}) at the respective gauging stations (NVE, 1987). According to the water balances, annual

differences between precipitation and runoff have been in the range 0.129–0.336 m yr⁻¹ (mean 0.226) at Dalelva and 0.216–0.484 m yr⁻¹ (mean 0.361) at Bjerkreim. Despite the uncertainties both in the precipitation and runoff estimates, the numbers seem reasonable when compared with regional evapotranspiration estimates previously reported for Norway (Otnes and Ræstad, 1978).

Results

N SPECIES COMPOSITION

The study sites differed widely with respect to streamwater TN and NO₃⁻ concentrations (Table 3). On an annual basis, the mean flow-weighted NO₃⁻ concentrations at Dalelva were extremely low, 11±3 µg N L⁻¹, whereas TN concentrations were more than ten times higher. Compared to Dalelva, the mean streamwater NO₃⁻ concentrations at Øygard were considerable (154±23 µg N L⁻¹), and on average NO₃⁻ at this site constituted nearly 60% of TN. In general, there were relatively small inter-annual variations in mean concentrations of both NO₃⁻ and TN at the study sites.

Like in most other pristine sites in Norway, the mean (not flow-weighted) NH₄⁺ concentrations at Dalelva and Øygard were very low, 6±3 (n=35) and 6±9 (n=22) µg N L⁻¹, respectively. The relatively high standard deviation value at Øygard was strongly affected by one single value (44 µg N L⁻¹). Apart from this, variation in NH₄⁺ concentrations

has been small at both sites during the study period and it seems reasonable to assume that the mean concentrations above are quite representative for the two catchments. As such, the annual mean TON concentrations at Dalelva and Øygard probably have been close to 130 and 105 µg L⁻¹, respectively. When considering that TOC concentrations at Dalelva are nearly three times higher than at Øygard, the corresponding mean TOC:TON molar C:N ratio was much higher at Dalelva (~35) compared to Øygard (~15).

INPUT/OUTPUT OF N

Yearly outputs of NO₃⁻ in runoff were negligible at Dalelva (mean <0.1 kg N ha⁻¹ yr⁻¹), corresponding to only 2% of the average atmospheric TIN deposition (Table 3). Relative to this, the streamwater TN export was considerably higher, 0.6 kg ha⁻¹ yr⁻¹. At Øygard, which receives much higher TIN deposition, NO₃⁻ leaching was considerably higher than at Dalelva. On average 3.0±0.3 kg N ha⁻¹ yr⁻¹ was lost from the catchment as NO₃⁻ during 1993–2000. This corresponded to nearly 20% of the TIN deposition from atmospheric sources during the same period. When including TON and NH₄⁺, the mean TN export at Øygard amounted to 5.1±0.6 kg ha⁻¹ yr⁻¹. Hence, the TN export at Øygard was nearly ten times larger than at Dalelva, whereas the corresponding NH₄⁺ + TON export was nearly four times higher.

A complete input/output budget for N would imply inputs from atmospheric deposition (both TIN and TON) and N fixation *v.* loss via streamwater runoff and N₂O emissions from denitrification and nitrification. Deposition of TON is

Table 3. Annual output fluxes of NO₃⁻ and TN relative to TIN inputs from atmospheric sources

	TIN	DALELVA			TIN	ØYGARD		
		NO ₃ ⁻ kg N ha ⁻¹ yr ⁻¹	TN	NO ₃ ⁻ % of TIN		NO ₃ ⁻ kg N ha ⁻¹ yr ⁻¹	TN	NO ₃ ⁻ % of TIN
1990	1.8	0.03	0.4	2	-	-	-	-
1991	2.2	0.02	0.4	1	-	-	-	-
1992	3.3	0.06	0.8	2	-	-	-	-
1993	2.6	0.06	0.7	2	12.6	2.5	4.2	20
1994	2.8	0.03	0.4	1	20.2	3.1	4.9	15
1995	2.7	0.04	0.6	2	16.5	3.1	4.7	19
1996	2.7	0.05	0.7	2	13.5	2.8	4.7	20
1997	2.3	0.05	0.5	2	17.7	2.5	5.9	14
1998	2.9	0.05	0.5	2	18.4	3.1	5.3	17
1999	3.2	0.04	0.5	1	16.9	3.5	5.7	20
2000	2.6	0.05	0.9	2	18.7	3.1	5.2	16
Mean	2.6	0.04	0.6	2	16.8	3.0	5.1	18
s.d.	0.4	0.01	0.2	0	2.6	0.3	0.6	2

not measured within the LRTAP programme, but data from monitoring sites in southern Norway indicate that this fraction constitutes an insignificant fraction of TN deposition (Oredalen and Aas, 2000). In other words, TIN deposition data from this area may to a great extent be equivalent to the TN deposition. Both N fixation and N_2O emissions are difficult to estimate on a catchment scale due to large spatial and temporal variations. According to the manual for mapping critical loads of acidifying deposition (UBA, 1996), however, both fluxes are usually low ($0\text{--}3\text{ kg ha}^{-1}\text{ yr}^{-1}$) and roughly balance in boreal ecosystems.

SEASONAL VARIATION IN N OUTPUTS

Both sites show distinct seasonal variations in NO_3^- concentrations, with a late-summer minimum and increasing levels during the dormant season (Fig. 3). At Dalelva, NO_3^- concentrations usually decreased below the detection limits during the growing season, but sometimes also small NO_3^- peaks were recorded during and after especially dry periods. However, these peaks did not contribute significantly to NO_3^- export from the catchment due to very low streamwater flow. At the end of the growing season, when snow usually starts to accumulate, NO_3^- concentrations gradually built up and culminated at levels around $40\text{--}100\text{ }\mu\text{g N L}^{-1}$ just before the onset of the spring flood. Also, during this period, the water discharges were quite low, such that the N fluxes

from the catchment remained small. The large difference between summer and winter NO_3^- concentrations under low flow conditions probably reflects differences in soil water NO_3^- content in the growing season (when plant demand for NO_3^- is high) v. the dormant season. Each year at the onset of the spring flood, there was an immediate decrease in the NO_3^- concentrations, suggesting an extensive dilution effect from the high water discharges following snowmelt.

Also at Øygaard, there was a steady increase in NO_3^- concentrations from late autumn (usually October–November), until maximum levels ($200\text{--}550\text{ }\mu\text{g N L}^{-1}$) were reached during winter (usually March–April). The culmination of the winter NO_3^- peak at Øygaard did not seem related to a particular hydrological event (e.g. snowmelt), but rather to be associated with initial stages of the growing season. The annual NO_3^- minimum ($40\text{--}100\text{ }\mu\text{g N L}^{-1}$) usually was reached during late summer or early autumn. It is worth noting that concentrations rarely decreased below $50\text{ }\mu\text{g N L}^{-1}$ during any time of the growing season. In most years there were relatively short periods with low flow conditions during summer. Beyond that, summer NO_3^- export could reach considerable levels ($30\text{--}40\text{ g N ha}^{-1}\text{ day}^{-1}$) during floods following heavy rain events. In contrast to Dalelva there was no significant decrease in NO_3^- concentrations during floods at any time of the year.

There were also large differences among the sites with respect to NO_3^- v. TN export (Fig. 4). Despite decreasing

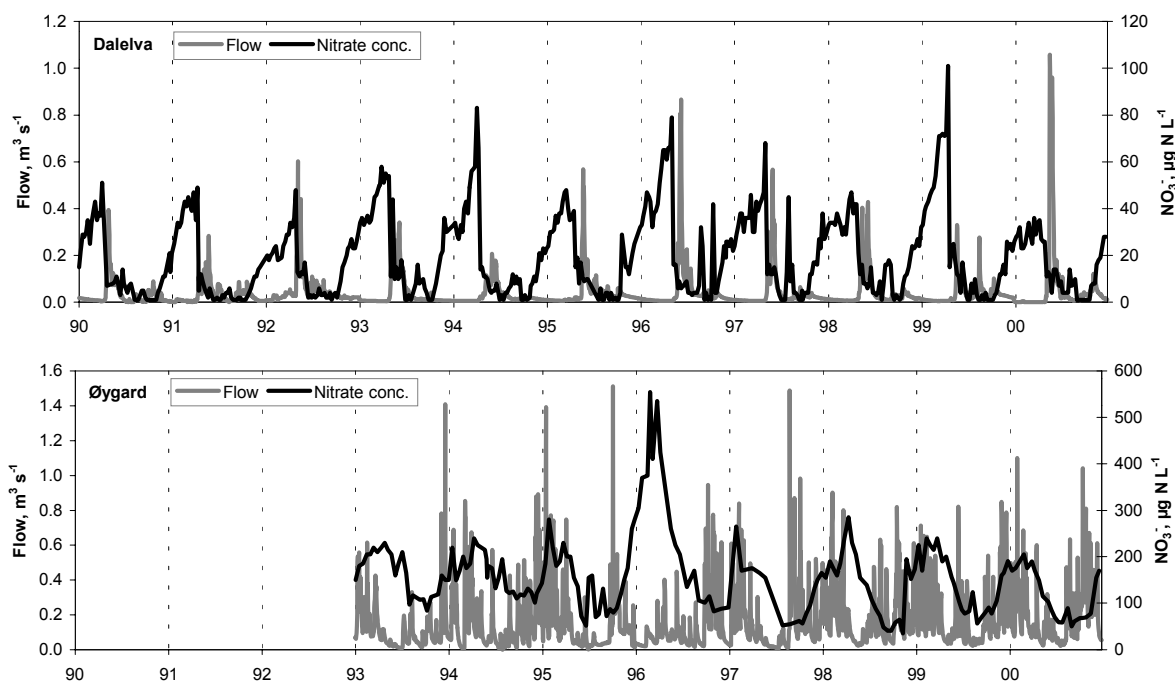


Fig. 3. Time series of NO_3^- concentrations and observed flow. Note differences in scale.

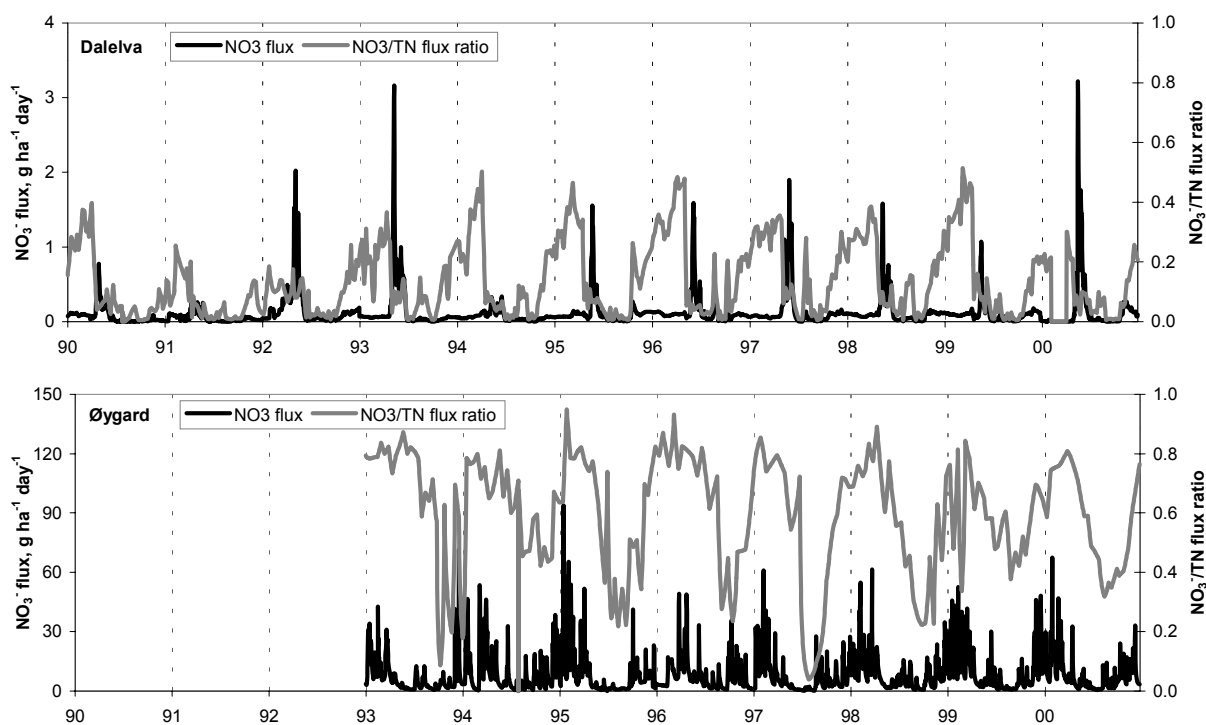


Fig. 4. Fluxes of NO_3^- and the corresponding NO_3^-/TN flux ratio. Note differences in scale.

concentrations, the NO_3^- export at Dalelva usually increased during the snowmelt flood due to the high water discharges. During 1990–2000, approximately 55% of the annual runoff and 51% of the annual NO_3^- export at Dalelva occurred during May–June, when the spring flood regularly occurred. Even at this time of year the maximum NO_3^- export, however, was much lower than at Øygard. The relative contribution of NO_3^- to the TN export at Dalelva usually increased during winter and culminated at levels around 30–50% just before the onset of the spring flood (Fig. 4). During summer and early autumn, however, N export from Dalelva was almost completely dominated by TON. At Øygard, approximately 44% of the annual runoff and 55% of the annual NO_3^- export during 1993–2000 took place during the winter months December–April. At this time of year, NO_3^- was the dominant N species, often consisting of >80% of the TN export. Dominance of TON export (NO_3^-/TN ratio <0.5) was only seen in short periods during late summer and early autumn. In general, the long-term data revealed no distinct seasonal pattern in TON concentrations or fluxes from the two catchments (data not shown).

Discussion

Streamwater NO_3^- concentrations at Dalelva and Øygard represent the lower v. the higher end of the range characterising Norwegian monitored catchments and lakes.

According to the 1995 Norwegian lake survey including 1006 statistically selected lakes covering the whole country (Henriksen *et al.*, 1998), 35% and 90% of the lakes had lower NO_3^- output concentrations than Dalelva ($11 \mu\text{g N L}^{-1}$) and Øygard ($154 \mu\text{g N L}^{-1}$), respectively. On a broader European scale, the Øygard catchment represents a more intermediate level with respect to output NO_3^- concentrations. In a database of 204 European and North American catchments and lakes included in the ICP Waters programme (Skjelkvåle *et al.*, 2001), about 8% and 45% of the sites had NO_3^- concentrations lower than 10 and $150 \mu\text{g N L}^{-1}$, respectively.

RELATIONS BETWEEN CLIMATIC CONDITIONS AND SEASONAL VARIATIONS IN STREAMWATER NO_3^-

In addition to representing opposite ends of the climatic gradient in Norway, the sites are also affected by a considerable climatic variability ranging from short-time episodes, seasonal and between-year variations, to long-term trends and cyclic changes that are beyond the time-scale of this study. By affecting terrestrial and aquatic N transformation processes, climatic conditions and their variation indirectly also have a large influence on concentrations and fluxes of N in streamwater. Due to lake percentages of 15 and 7% at Dalelva and Øygard,

respectively, water residence time in the lakes may to some extent have a damping effect on the seasonal pattern of streamwater NO_3^- concentrations. According to studies of 12 Norwegian lake and stream sites with lake percentages in the range 0–26%, however, the damping effects on seasonal NO_3^- concentrations were not clear until the lake percentage exceeded 15% (Ø. Kaste, *unpubl. data*).

During the long and cold winters at Dalelva, when in practice all precipitation is accumulated as snow, the stream flow is very low and dominated by baseflow. The steady increase in NO_3^- concentration observed every winter at this site probably is a result of minimal dilution effects from percolating water and little or no N uptake by plants after establishment of a permanent snow cover. Additionally, net N mineralisation and subsequent nitrification below the snowpack may have contributed to the observed NO_3^- pattern (Williams *et al.*, 1996). Low streamwater NH_4^+ concentrations during winter indicate that mineralised N is rapidly immobilised or transformed to NO_3^- within the catchment (Dillon and Molot, 1990) or in the stream (Peterson *et al.*, 2001). Due to the low flow conditions prevailing during winter, the elevated NO_3^- concentrations do not contribute significantly to the annual streamwater NO_3^- export.

Immediately at onset of the snowmelt flood, streamwater NO_3^- concentrations decrease rapidly to very low levels. This rapid decrease obviously was caused by dilution from snowmelt and maybe also assimilation by biota. Analogous patterns in NO_3^- concentrations during snowmelt are observed in other alpine and sub-alpine catchments (e.g. Williams *et al.*, 1996). Tracer studies have demonstrated that NO_3^- in snowmelt runoff mainly comes from pre-melt soil water rather than the melting snow itself (Kendall *et al.*, 1995; Williams *et al.*, 1996). In other words, NO_3^- and NH_4^+ eluted from the seasonal snowpack appear to be infiltrated and immobilised in the soils. If flushed directly to the stream during snowmelt, this amount of N would have caused a considerable export of NO_3^- and maybe also NH_4^+ . Despite the rapid decrease observed in streamwater NO_3^- concentrations during snowmelt, the NO_3^- export during the spring flood still constituted a considerable fraction of the annual NO_3^- output from the catchment (Fig. 4).

After snowmelt, the Dalelva catchment experiences a short and intensive growing season that lasts approximately four months. During summer, air temperatures can be as high as at Øygard (Fig. 2), and due to relatively low precipitation amounts, streamwater flow usually is very low at that time. In general, NO_3^- concentrations are near the detection limits during summer due to a high plant and microbial demand for N (Fig. 3). However, small NO_3^- peaks are sometimes observed after warm and dry periods e.g. during mid-

summer in 1996 and 1997. In these cases, drought and re-wetting of soils might have caused elevated mineralisation and nitrification with subsequent increases of soil water NO_3^- concentrations (Reynolds *et al.*, 1992; Murdoch *et al.*, 1998).

In contrast to Dalelva, the Øygard site is characterised by mild winters that often are accompanied by high stream flow. Due to the lack of a permanent snow cover and often absence of soil frost, the N transformation processes at Øygard may proceed throughout large parts of the winter. With air temperatures largely fluctuating around 0°C during winter, the upper soil horizons will undergo frequent freezing/thawing that may affect plant roots and soil organic matter and increase the substratum available for mineralisation (Monteith *et al.*, 2000). In turn, this may lead to increased nitrification, reduced plant uptake and occasionally also increased denitrification (Ryan *et al.*, 2000). The net effect, however, is often increased N leaching to surface waters (Mitchell *et al.*, 1996; Monteith *et al.*, 2000).

Also at Øygard the NO_3^- concentrations increased steadily throughout the dormant season, even though the stream flow was highly variable. Especially high NO_3^- concentrations were observed during the winter of 1995–96 (Fig. 3). This winter was atypical in terms of climate, with a relatively long period characterised by low temperatures and unusual low stream flow (Figs. 2 and 3). Apart from the concentration levels, the NO_3^- pattern this winter had many similarities with an average winter situation at Dalelva. The elevated NO_3^- concentrations at Øygard this winter probably reflect strongly reduced dilution of soil water due to small amounts of percolating water. Due to the unusual low stream flow, the streamwater NO_3^- export remained low during this period (Fig. 4).

The massive NO_3^- concentration peak observed at Øygard during winter in 1995/1996 appeared simultaneously in lakes and streams in the UK (Monteith *et al.*, 2000). Hence, streamwater NO_3^- dynamics at that time could be controlled by a large-scale climatic event. In this context, Monteith *et al.* (2000) found that variations in NO_3^- concentrations at the UK sites showed a strong negative correlation with winter values for the North Atlantic Oscillation (NAO) index and mean winter temperature. High winter values for the NAO index are associated with relatively warm and wet winter conditions in northwestern Europe, whereas low values coincide with colder and drier conditions (Hurrell, 1995).

Except for the extraordinary conditions prevailing during the winter of 1995–96, there was usually no clear relationship between stream flow and the corresponding NO_3^- concentration at Øygard (Fig. 5). Thus, there were no obvious dilution effects on NO_3^- concentrations even during large flood events. This stands in sharp contrast to Dalelva,

where there was a strong inverse relationship between stream flow and the corresponding NO_3^- concentration. Hence, the lack of clear dilution effects at Øygard may indicate that the NO_3^- leaching potential from soils in this catchment is relatively high. One possible mechanism contributing to such a pattern might be that during heavy flood episodes, some fraction of NO_3^- in event water is flushed directly to the stream instead of percolating through the soil profile: parts of the catchment are dominated by bare rocks and patchy soils with low infiltration capacity. Such mechanisms are, however, difficult to identify without continuous sampling (Moldan and Wright, 1998) or by applying isotope techniques (e.g. Durka *et al.*, 1994; Kendall, 1998).

DO THE SITES REPRESENT STAGES IN A PROBABLE TIME COURSE OF CATCHMENT RESPONSE TO CHRONIC N ADDITIONS?

Due to similar catchment characteristics (size, geology, land

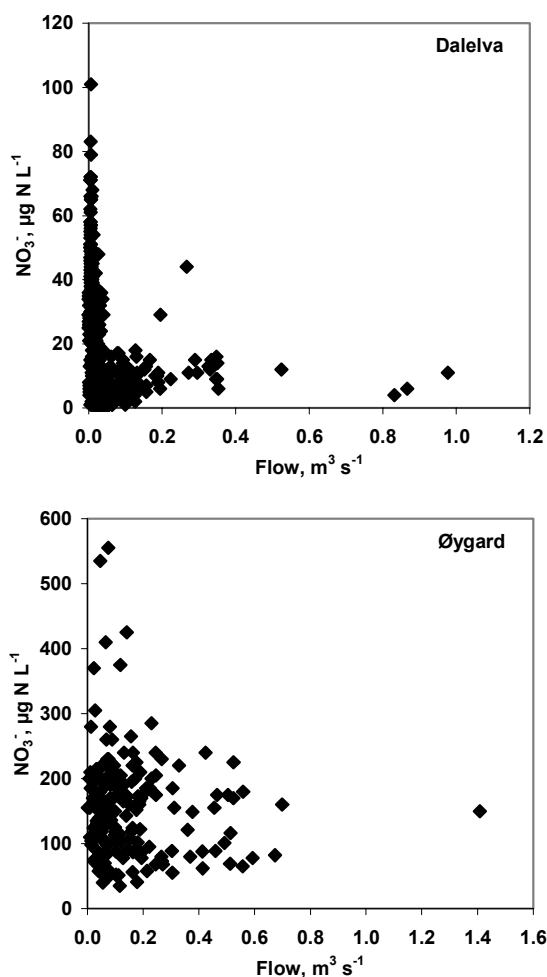


Fig. 5. NO_3^- concentrations vs. observed flow. Note differences in scale.

cover), water chemistry at Dalelva and Øygard probably was quite uniform prior to the anthropogenic influences that altered N deposition at Øygard and S deposition at both sites. On this basis, it might be reasonable to regard Dalelva as a sort of “reference” to Øygard with respect to trends in catchment N status and N leaching. Thus, the differences between the Dalelva and the Øygard catchments today may be viewed as a response to long-term N deposition.

After decades with elevated N deposition at Øygard, the catchment today appears quite enriched with N compared to Dalelva. In addition to the differences in streamwater NO_3^- concentrations, this also seems to be reflected in C:N ratios of the soil organic layers. At Dalelva, average C:N ratios in four plots covered with birch trees were in the range 29–39 (SFT, 1989; A.O. Stuanes, unpubl. data), whereas corresponding C:N values in four wetland- and heather plots at Øygard were in the range 22–29 (Wright *et al.*, 1999). In this context, it must be noted that slower and less effective decomposition of organic matter in the cold climate characterising the northern site may to some extent maintain a high C:N ratio. The net effect of this is difficult to quantify. It is, however, documented from several forested sites in Europe that elevated atmospheric N inputs have led to decreased soil C:N ratios, and empirical data demonstrate a negative correlation between NO_3^- leaching and C:N ratios in the forest floor (Gundersen *et al.*, 1998). A similar, but less significant negative relationship is also found between NO_3^- concentration and the DOC:DON ratio in runoff water (Harriman *et al.*, 1998). In this context, C:N ratios in dissolved organic matter runoff were much higher at Dalelva (~35) than at Øygard (~15) during the study period.

Both the damped seasonal variations in NO_3^- concentrations and the present NO_3^- leaching illustrate that the N supply at Øygard is in excess of the combined plant and microbial demand. According to the N saturation stage classification developed by Stoddard (1994) and Stoddard and Traaen (1995), the site falls in-between stage 1 and 2, which denotes N saturated systems with moderate to high annual NO_3^- losses. In comparison, the Dalelva catchment falls into stage 0, which characterises a non-saturated system. At Øygard, NO_3^- leaching takes place despite the fact that the growing season, and hence the plant demand for N, is much more extensive than at Dalelva. The absence of dilution effects on NO_3^- concentrations during floods at Øygard may further underline that there is a large N pool in the soils available for leaching in this catchment. This stands in sharp contrast to the Dalelva catchment, where the N pool obviously is much smaller and flood events effectively dilute streamwater NO_3^- concentrations. Even in this pristine catchment receiving small amounts of N deposition, however, it seems inevitable that some of the NO_3^-

accumulated in the soils during winter is lost via runoff before onset of the growing season.

If Dalelva and Øygard represent a probable time course of catchment response to chronic N additions, there has been a considerable increase in NO_3^- leaching at Øygard during the last century. This is demonstrated spatially in several European forest sites, where very little NO_3^- leaching occurs at N deposition below 9–10 kg N ha^{-1} yr^{-1} , leaching can occur at intermediate deposition between 9 and 25 kg N ha^{-1} yr^{-1} , and significant leaching occurs at almost all sites receiving deposition greater than 25 kg N ha^{-1} yr^{-1} (Dise and Wright, 1995). One crucial question is whether the present NO_3^- leaching at Øygard represents steady-state conditions with respect to the ambient N deposition level, or if further increased NO_3^- leaching may be expected in the future as a result of decreasing C:N ratios (Gundersen *et al.*, 1998) and possibly also climate change (Murdoch *et al.*, 1998; Wright, 1998). In contrast to forest ecosystems, the knowledge of N retention capacity in heathland and mountainous areas is relatively scarce. Thus, a better understanding of N transformation processes in these areas is essential for assessing the possible consequences of a maintained (or slightly reduced) N deposition level in the future.

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