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**Scaling ozone  
photochemistry II**

B. Ainslie and D. G. Steyn

# A scaling analysis of ozone photochemistry: II Investigation of the similarity relationship

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## Abstract

The similarity relationship for maximum ozone concentration as a function of initial precursor concentrations developed in the first article of this 2-part series exhibits a scale break, most easily seen after a Weibull transformation, which identifies a characteristic scale for ozone photochemistry. In this paper, we investigate the similarity relationship using simple semi-quantitative models and model output. From this investigation, we develop a set of properties characterizing ozone-precursor relationships. We find the scaling break is associated with a change in the temporal variability of ozone production. Finally, we examine a series of smog chamber experiments for evidence of a scaling break. We find the data support a slight break after suitable transformation. It is difficult to tell if the lack of distinct break is due to smog chamber experimental limitations, a bias in the test conditions or if the processes leading to the scaling break are overly enhanced by chemical mechanisms.

## 1. Introduction

Ozone isopleth diagrams are a useful tool for summarizing the dependence of ozone formation on initial precursor concentrations. The most striking feature of these diagrams is the non-linear relationship between ozone and its precursors – creating a response surface with distinct VOC- and NO<sub>x</sub>-sensitive regions. These regions can be characterized in several (related) ways: in terms of response of maximum ozone concentration to changing initial NO<sub>x</sub> concentration at fixed VOC concentration (National Research Council, 1991); in terms of the relative importance of peroxides to nitric acid production (Sillman, 1995); in terms of the fraction of radicals lost in reactions involving NO<sub>x</sub> (Kleinman, 2005); by the fraction of HO<sub>2</sub><sup>•</sup> radicals that react with NO (Tonnesen and Dennis, 2000); by the extent to which NO<sub>x</sub> has reacted to produce the maximum possible ozone concentration (Blanchard et al., 1999) or by the ratio of the lifetime of OH<sup>•</sup> with respect to losses by reactions with VOC and NO<sub>x</sub> (Kirchner et al., 2001).

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The dimensional analysis of Ainslie and Steyn (2005) suggests these regions can also be characterized by the change in behaviour of the ratio of dimensionless maximum ozone concentration to dimensionless initial  $\text{NO}_x$  concentration:

$$\frac{[\text{O}_3]_{\max}}{j_{\text{av}}/k_{\text{NO}}} / \left( \frac{[\text{NO}_x]_0}{j_{\text{av}}/k_{\text{NO}}} \right)^a \quad (1)$$

5 across a specific value ( $\beta$ ) of the ratio of initial VOC to  $\text{NO}_x$  concentrations ( $R$ ). In this paper, we use this behaviour to guide the development of semi-quantitative models for ozone formation. From these models, we infer a set of properties which characterize simple ozone-precursor relationships. In many ways, our line of inquiry parallels the development of the Integrated Empirical Rate (IER) model (Johnson, 1984) except  
10 this model was based on smog chamber data whereas we use model output. As such, our results express the behaviour of a set of mathematical relations describing ozone photochemistry – an imperfect representation of the actual atmosphere – which may differ from the (also imperfect) representation of the atmosphere given by smog chambers. As a result, we also examine how well the properties developed in this  
15 paper can be used to characterize ozone formation in a smog chamber.

## 2. A step-function model for ozone production

Ozone concentrations are the product of the integrated rate of chemical production in excess of chemical destruction over the history of a precursor mixture. As such, the net chemical production,  $P(\text{O}_3)$ , provides a fundamental framework for studying ozone-precursor relationships. Our analysis begins with a simple model for ozone production  
20 which is a function of initial precursor concentrations and the integrated  $\text{NO}_2$  photolysis rate constant ( $j_{\text{NO}_2}$ ):

$$j(t) = \int_{t_0}^t j_{\text{NO}_2} ds. \quad (2)$$

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We follow Johnson (1984) by using  $j(t)$  as a proxy for cumulative actinic flux and consider it a dimensionless process time. We define  $P_j(\text{O}_3)$  as the sensitivity of ozone concentration to the integrated  $\text{NO}_2$  photolysis rate constant:

$$P_j(\text{O}_3) \equiv \frac{d[\text{O}_3]}{dj(t)} \quad (3)$$

We now develop a semi-quantitative model for  $P_j(\text{O}_3)$  in order to provide a qualitative description of ozone formation. This model is not intended to capture specifics of photochemical process, rather it is a toy model used to illustrate general principles. Specifically, we relate individual properties of the semi-quantitative model with basic features of the similarity function such as the scaling break and the “s-shaped” profile.

We start with the simplest possible semi-quantitative model but add more realism in Sects. 3 and 4 as the analysis reveals model limitations.

We begin with a step-function model for  $P_j(\text{O}_3)$  where production is initially constant before switching to a zero-production regime at some time  $j^*$ . We give  $P_j(\text{O}_3)$  a VOC dependence by assuming the constant production increases with increasing initial VOC concentration (for now we ignore any  $\text{NO}_x$  dependence but include these in Sect. 4). We define the model with the following four rules which we illustrate in Fig. 1:

**Rule I:**  $P_j(\text{O}_3)$  is initially constant until a critical  $j$ -value ( $j^*$ ) at which production drops to zero. We motivate this rule by noting that most precursor mixtures are *initially* radical limited, regardless of VOC to  $\text{NO}_x$  ratio, with production limited by the rate of  $\text{OH}^\bullet$  attack on VOC, however, as the system reacts and  $\text{NO}_x$  is converted to  $\text{NO}_z$ , the system stops reacting and production ceases.

**Rule II:** The larger the initial VOC concentration, the larger initial  $P_j(\text{O}_3)$ . Again, we assume production is limited by rate of  $\text{OH}^\bullet$  attack on VOC and so scales with  $[\text{VOC}]_0$ .

**Rule III:** The larger the initial VOC concentration, the earlier production ceases. With increased production, we expect ozone formation to occur sooner accompanied by more rapid  $\text{NO}_z$  production and the earlier onset of  $\text{NO}_x$ -limited conditions.

**Rule IV:** The switchover time ( $j(t)^*$ ) is such that the area under the  $P_j(\text{O}_3)$ - $j$  curve is independent of initial VOC concentration. We associate the shift in production with

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a switch to  $\text{NO}_x$ -limited conditions where we have found maximum ozone concentrations (corresponding to the total area under the  $P_j(\text{O}_3)-j(t)$  curve) to be independent of  $[\text{VOC}]_0$  (Ainslie and Steyn, 2005).

In Fig. 2 we consider the relationship between ozone production, ozone concentration and maximum ozone concentration, using the step-function model, for a set of precursor mixtures having fixed  $[\text{NO}_x]_0$ , a range of  $[\text{VOC}]_0$  and total actinic flux set at  $j(t_{\text{end}})=300$ . The production curves (Fig. 2A) give rise to ozone concentration curves which increase linearly (Fig. 2C) and if initial production is great enough and switchover occurs, then ozone concentrations reach a maximum and remain constant. For example, while the highest production curve ( $[\text{VOC}]_0=4.0$  ppm) results in ozone concentrations rising quickest, the  $[\text{VOC}]_0=3.0$  and  $[\text{VOC}]_0=2.0$  ppm curves also achieve the same final ozone concentration. The low VOC curve ( $[\text{VOC}]_0=1.0$ ) rises the slowest and does not reach the maximum before simulation end.

In Fig. 2D, we plot  $[\text{O}_3]_{\text{max}}$  concentration as a function of  $[\text{VOC}]_0$ , where two linear line segments have been fitted to the results. The arrow running between Figs. 2C and 2D shows how the *final* ozone concentration along any curve in Fig. 2C becomes the *maximum* concentration for that  $[\text{VOC}]_0$  and is therefore used as the ozone concentration in the  $[\text{O}_3]_{\text{max}}$  versus  $[\text{VOC}]_0$  plot (to avoid clutter, only four curves are shown in Figs. 2A and C, whereas Fig. 2D has been plotted using an additional eight  $[\text{VOC}]_0$ ). Figure 2D shows the step function model produces a maximum-ozone/initial-VOC curve which is piecewise linear and attains its maximum value – unlike the “s-shaped” curves found in non-diluted model output (Fig. 1 of Ainslie and Steyn, 2005)<sup>1</sup>. We check for the pres-

<sup>1</sup>While the similarity relationship in Ainslie and Steyn (2005) results from a plot of  $[\text{O}_3]_{\text{max}}/[\text{NO}_x]_0^a$  (or more precisely dimensionless  $[\text{O}_3]_{\text{max}}$  and dimensionless  $[\text{NO}_x]_0$ ) versus  $R$ , it can be interpreted as a series of cross-sectional cuts from an isopleth diagram each along a different line of constant  $[\text{NO}_x]_0$ . Scaling  $[\text{O}_3]_{\text{max}}$  by the underlying  $[\text{NO}_x]_0$  vertically compresses each cross-section to a common maximum. Furthermore, plotting against  $R$  (and not  $[\text{VOC}]_0$ ) shifts each cross-section horizontally until they overlap. As a result, a single  $[\text{O}_3]_{\text{max}}$  versus  $[\text{VOC}]_0$  (at a fixed  $[\text{NO}_x]_0$ ) plot produces a curve analogous to the similarity relationship;

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ence of a scaling break by Weibull transforming the maximum ozone concentrations and plotting against  $\ln[\text{VOC}]_0$  (Fig. 2B). To avoid singularities when Weibull transforming a value of unity, we choose the normalizing constant to be 1% larger than the maximum ozone concentration; a choice which does not affect our conclusions. The Weibull transformed data cluster onto two curves with the one corresponding to maximum ozone concentrations under  $\text{NO}_x$ -limited conditions being a straight line with zero slope. This zero slope and the linear dependence of  $[\text{O}_3]_{\text{max}}$  on  $[\text{VOC}]_0$  highlight two discrepancies between the step-function model and the similarity functions given in Ainslie and Steyn (2005). Both are related to the simplistic way the dependence of  $[\text{VOC}]_0$  and  $j(t)$  on  $P_j(\text{O}_3)$  have been parameterized in the step-function model. Despite these discrepancies, it should be noted that the step function model forms the basis of the IER model (Johnson, 1984), a simple set of formulae used to quantitatively describe ozone formation, developed from the analysis of an extensive set of outdoor smog chamber experiments. As a result, Fig. 2C duplicates the IER model (but with  $\text{O}_3$  plotted against  $j(t)$  instead of primary smog products (*PSP*)). While this model has been criticized for its linear dependence of maximum ozone concentration on initial  $\text{NO}_x$  (Blanchard, 2000), the basis for the model is supported by environmental chamber data and current understanding of atmospheric chemistry (Blanchard et al., 1994). As such, we find the step function model provides a useful framework for our analysis and modify it in Sect. 3 to include a more realistic temporal variability for  $P_j(\text{O}_3)$ .

### 3. A more complex model for ozone production

The step-function's linear  $[\text{O}_3]_{\text{max}}$ - $[\text{VOC}]_0$  relationship gives rise to “L” shaped isopleths (not shown) which suggests the parameterization does not capture the strong inhibition of ozone formation above the ridgeline exhibited by most chemical mechanisms (Tonnesen and Jeffries, 1994). This “ $\text{NO}_x$ -inhibition” involves a complex set of interactions involving a competition between  $\text{NO}_2$  and VOCs for available radicals with ozone independent of any underlying parameterization.

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photolysis representing a major inorganic source for these radicals. To parameterize this behaviour, we let  $P_j(\text{O}_3)$  mirror the temporal evolution for the ozone photolysis rate: production starts at zero, increases, reaches a maximum value and then decreases. In addition, to ensure the corresponding ozone concentrations asymptotically approach their maximum value, we make the decrease exponential. Finally, we make the maximum value of  $P_j(\text{O}_3)$  increase with  $[\text{VOC}]_0$  and the  $j$ -value at this maximum decrease with  $[\text{VOC}]_0$ . We illustrate this ozone production model in Fig. 3.

This model is easily parameterized using a Weibull distribution with scale factor inversely proportional to  $[\text{VOC}]_0$ . For such a parameterization, both the corresponding  $[\text{O}_3]$  versus  $j(t)$  curves and the  $[\text{O}_3]_{\text{max}}$  versus  $[\text{VOC}]_0$  curve, would be parameterized by Weibull functions and produce “s-shaped” curves. However, it is evident that such a parameterization model would not produce a scaling break upon Weibull transformation (the single Weibull function parameterization produces a single straight line upon transformation)<sup>2</sup>. This suggests that in order to produce a realistic parameterization for ozone production which includes a scaling break, it is not sufficient for the parameterization to include: a local maximum in production, a change in curvature or even exponentially decaying production. For the step-function model, the break arises from the abrupt change in the functional form of  $P_j(\text{O}_3)$ , reflecting a structural change in behaviour of the chemical mechanism, occurring between high and low  $\text{NO}_x$  regimes. As Kleinman (1994) suggests, this change represents more than continuous differences in production as  $\text{NO}_x$  is varied. We conclude that in order to produce a scaling break, our semi-quantitative model should have an abrupt change in functional form for  $P_j(\text{O}_3)$ . To further characterize this change in functional form, we consider the influence of  $\text{NO}_x$  on  $P_j(\text{O}_3)$ . In so doing, we generalize the four rules used to develop the toy model in Sect. 2 and define a set of properties characterizing ozone-precursor relationships.

<sup>2</sup>The same conclusions hold even if  $P_j(\text{O}_3)$  is parameterized by a lognormal or gamma function. In these cases, the Weibull transformed data is slightly curved but, again, does not produce a distinct break.



#### 4. General properties for ozone-precursor relationships

In order to characterize ozone production with realistic  $\text{NO}_x$  dependence, we examine the temporal variability of ozone concentration from a set of nine simulations using the OZIPR trajectory model with an urban speciation based on Stockwell et al. (1997).

Each simulation begins with an initially well mixed charge of VOC and  $\text{NO}_x$ , and in order to study the chemical processes in isolation from competing physical processes, we allow no subsequent emissions, dilution or deposition. A fixed temperature of  $25^\circ\text{C}$  is used throughout the 11-h simulation with an actinic flux appropriate to Vancouver B.C. (latitude  $49.1^\circ\text{N}$ ) on 21 June. To make presentation of results more concise and guided by our analysis of the dependence of dimensionless maximum ozone concentration on dimensionless actinic flux (Ainslie and Steyn, 2005), we plot  $[\text{O}_3](t)$  versus  $R \times j(t)^c$  (with  $c=1.2$ ) in Fig. 4. This figure shows three plots, each giving results from three simulations having different  $[\text{VOC}]_0$  but common  $[\text{NO}_x]_0$ , with corresponding  $R$ -values of: 12, 24 and 30 (all greater than  $\beta (=9.4)$ , the  $[\text{VOC}]_0/[\text{NO}_x]_0$  ratio at the scaling break). In Fig. 4A each simulation starts with an initial  $\text{NO}_x$  concentration of 15 ppb, while in Fig. 4B it is 45 ppb and 75 ppb in Fig. 4C. From these plots, we make the following observations:

- The rate of increase in  $[\text{O}_3]$  slows down at  $R \times j^c \approx 8600$ ; independent of both initial  $[\text{NO}_x]_0$  and  $[\text{VOC}]_0$ . We associate the slowing with a switch in the behaviour of ozone production and observe the timing of this switch depends on  $R$  – occurring sooner for mixtures having higher  $R$ -values.
- Ozone concentrations at the time of the switch are roughly the same for simulations with identical  $[\text{NO}_x]_0$ ; independent of  $[\text{VOC}]_0$ . In terms of production, this implies that for a given initial  $[\text{NO}_x]_0$ , switchover occurs for a fixed area under the production curve, independent of  $[\text{VOC}]_0$ . For fixed  $[\text{NO}_x]_0$ , with the roughly constant  $[\text{O}_3]$  at switchover and with the switchover occurring sooner for higher  $R$ -values, peak  $P_j(\text{O}_3)$  increases with  $R$ . Finally, ozone concentrations at the switch

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increase with increasing  $[\text{NO}_x]_o$  for simulations with constant  $R$ -values.

- For simulations with identical  $R$ -values, increasing  $[\text{NO}_x]_o$  causes  $[\text{O}_3]$  to initially increase more rapidly (implying peak  $P_j(\text{O}_3)$  increases with  $[\text{NO}_x]_o$ ) and ultimately reach a higher final value.

5 In an effort to understand the scaling break, we developed a toy model for  $P_j(\text{O}_3)$  based on four rules. While the Weibull transformed  $[\text{O}_3]_{\text{max}}$  vs  $[\text{VOC}]_o$  plot (Fig. 2C) based on this toy model produced a scaling break, the line segment after the break had zero slope – a characteristic not observed in similar plots using model output (Ainslie and Steyn, 2005). In an effort to remedy this and other discrepancies, a  
10 more complex model was considered which produced an “s-shaped” profile for the similarity relationship but at the expense of eliminating the scaling break. From this we concluded a semi-quantitative model producing a scaling break required a change in function form for the parameterization of  $P_j(\text{O}_3)$ . To generalize our results and include a  $\text{NO}_x$ -dependence into our analysis, we have examined model output. From this  
15 output, it appears that any parameterization of ozone formation should include the following properties:

**Property I:** To ensure the temporal variability of  $[\text{O}_3]$  is “s-shaped”,  $P_j(\text{O}_3)$  should initially increase, reach a peak value and then asymptotically decrease.

**Property II:**  $P_j(\text{O}_3)$  should increase with  $R$  and with  $[\text{NO}_x]_o$ .

20 **Property III:** To ensure the corresponding Weibull transformed similarity relationship has a scaling break, the parameterization of  $P_j(\text{O}_3)$  should have a change in function form.

**Property IV:** Larger  $R$ -values should cause the transition time ( $j^*$ ) of  $P_j(\text{O}_3)$  to occur sooner.

25 **Property V:** The total area under the  $P_j(\text{O}_3)$ - $j(t)$  curve, as well as the area up to the transition time should be functions of  $[\text{NO}_x]_o$  only.

**Property VI:** The corresponding parameterization for  $[\text{O}_3]$  should scale with  $R \times j(t)^c$ .

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It is easy to show properties II–VI generalize the IER model (see Appendix) and based on the scaling analysis of Ainslie and Steyn (2005), and work by Blanchard (2000) and Chang and Rudy (1993), we suggest one additional property:

**Property VII:** The corresponding relationship between  $[O_3]_{\max}$  and  $[NO_x]_o$ , for large VOC to  $NO_x$  ratios, should show  $[O_3]_{\max} \propto [NO_x]_o^a$  with  $a < 1$ .

In Ainslie and Steyn (2005), dimensional analysis was used to represent the integrated behaviour of a photochemical mechanism. In this paper, the results of the dimensional analysis have guided the development of a semi-quantitative model for ozone production and help produce a set of properties which characterize ozone-precursor relationships. We find a parameterization of  $P_j(O_3)$  which includes a scaling break, must have a change in functional form which is more than a local maximum or change in curvature and is related to change in the temporal dynamics controlling ozone production. However, the similarity function developed by Ainslie and Steyn (2005) shows the scaling break arising from a plot of (Weibull transformed and dimensionless)  $[O_3]_{\max}/[NO_x]_o$  versus  $R$ . To understand the link between the two, we note the present analysis shows the relative abundance of VOC to  $NO_x$  ( $R$ -value) determines whether the switch in chemical regimes occurs within the given process time (c.f. Property IV). Thus precursor mixtures with  $R$ -values greater than  $\beta$  will switch to the low production regime during the process time; noticeably limiting maximum ozone concentrations (the plotting of maximum ozone concentration divided by initial  $NO_x$  concentration serves to scale the model output onto a single curve and does not contribute to the break in scaling). Thus the behaviour of the  $[O_3]_{\max}/[NO_x]_o$  versus  $R$  curve will change for  $R$ -values greater than  $\beta$ .

Property VI suggests it should be possible to extend the parameterization of the similarity relationship (Eq. 10 of Ainslie and Steyn, 2005) to include a temporal dependence

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through the replacement of  $R$  with  $Rj^c$ :

$$\frac{[\text{O}_3]}{j_{av}/k_{\text{NO}}}(j(t), R, [\text{NO}_x]_o) = \gamma \left( \frac{[\text{NO}_x]_o}{j_{av}/k_{\text{NO}}} \right)^a \left( 1 - \exp \left[ -\lambda \left( \frac{R}{\beta} \left( \frac{j}{J} \right)^c \right)^\alpha \right] \right) \quad (4)$$

where  $j(t_{\text{end}}) = J$ . To see how well this parameterization holds, it has been compared with model output from the RADM2 mechanism using a urban VOC (Stockwell et al., 1997) in Fig. 5. Parameter values for Eq. (4) were not found by regression against the temporal ozone profiles but against 100 maximum ozone and initial VOC and  $\text{NO}_x$  triplets as described in Ainslie and Steyn (2005). Except for the low  $R$ /low  $\text{NO}_x$  plots ( $R=6$  and  $[\text{NO}_x]_o=0.015, 0.045$  ppm), Eq. (4) shows good agreement with the model output. In many cases, Eq. (4) does not match the transition time exactly but does manage to be within  $\pm 10\%$ . In the next section, we use scaling behaviour properties, developed from model output, to analyze smog chamber data.

### 5. Scaling properties of smog chamber data

Smog chamber data from a series of experiments performed at the CSIRO outdoor smog chamber facility in the North Ryde suburb of Sydney (latitude  $33^\circ 52' \text{ S}$ , longitude  $151^\circ 13' \text{ E}$ ) was used in the analysis. The data consist of ten sets of experiments performed in two separate cubical smog chambers using a mixture of VOCs typical of Sydney's air (Hess et al., 1992). All experiments were performed without dilution and with the VOC and  $\text{NO}_x$  initially well mixed. Typically, the experiments started at sunrise and continued until late afternoon. Table 1 lists the test conditions for the 20 experiments. As Table 1 shows, the CSIRO data set covers only a limited range of  $R$ -values ( $[1.6, 10.2]$ ), so temporal ozone concentrations (as a function of  $Rj^c$ ) were examined (using Eq. 4) to allow an analysis over a wider range. To complicate matters, the experiments span a wide range of temperatures which must be accounted for.

We start by expressing the data as dimensionless groups:  $\pi(=[\text{O}_3]/(k_{\text{NO}}/j_{av}))$  for the ozone concentrations,  $\pi_1(=[\text{NO}_x]_o/(k_{\text{NO}}/j_{av}))$  for the initial  $\text{NO}_x$  concen-

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trations,  $\pi_2(=[\text{VOC}]_o/[\text{NO}_x]_o)$  for the relative abundance of VOC to  $\text{NO}_x$  and  $\pi_4 = \exp[-504/(T_{av} + 273) - 1/298]$  for the dimensionless temperature dependence<sup>3</sup>. In terms of Pi-groups the scaling analysis suggests:

$$\frac{\pi}{\pi_1^a \pi_4^b} = \gamma f(\pi_2 j^c) \quad (5)$$

and so the objective is to plot  $\pi/(\pi_1^a \pi_4^b)$  versus  $\pi_2 j^c$  to determine if the data collapse onto a common curve *and* to see if this curve (i.e. the similarity relationship) can be parameterized by a composite of two Weibull functions. To test these two conjectures, the following steps were taken:

1. For each experiment, the total  $\text{NO}_2$  photolysis rate and total test time were used to calculate the average photolysis rate ( $j_{av}$ ).
2. The average temperature was used to determine  $k_{\text{NO}}$  and  $\pi_4$ .
3. The initial  $\text{NO}_x$  concentration was used to determine  $\pi_1$ ; the initial VOC and  $\text{NO}_x$  concentrations for  $\pi_2$ ; and ozone concentrations for  $\pi$
4. The cumulative  $\text{NO}_2$  photolysis rate was used to determine  $j(t)$  based on calculated local solar elevation.

The resulting dimensionless data are plotted as a function of  $\pi_2 j^c$  in Fig. 6A. Also shown on this plot is the similarity function  $f(\pi_2 j(t)^c)$  that best fits the data. Values for the three exponents ( $a$ ,  $b$ ,  $c$ ) as well as the similarity parameters are given in Table 2. The level of agreement between the similarity function and the data can be seen in the scatter plot of observed versus calculated ozone concentrations (Fig. 6B). The plot shows our model fitting the data quite well (the scatter plot statistics are given in Table 3) with little bias.

<sup>3</sup>We have slightly modified our dimensionless temperature with the inclusion of the  $\exp[504/298]^b$  term to be consistent with IER model (Johnson, 1984).

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The dimensionless data, after normalization and Weibull transformation are plotted in Fig. 6C. It is difficult to see the scaling break in Fig. 6C because the “kink” in the data (at  $\ln(Rj^c)=9.7$  ( $Rj^c=17\,000$ )) is slight. While the data support the break, it is not as distinct as those seen in the model output for the reactive RADM2 classes (Ainslie and Steyn, 2005). For a sharp break to occur, a large change in ozone production is needed across  $R=\beta$ . We suspect this change can be understood in terms of the strength of  $\text{NO}_x$ -inhibition occurring within a VOC- $\text{NO}_x$  mixture. At low VOC to  $\text{NO}_x$  ratios, ozone formation is constrained by radical availability with  $\text{NO}_2$  and VOCs in competition from the mainly ozone-produced  $\text{OH}^\bullet$  radicals. Under these conditions, increasing  $\text{NO}_x$  concentrations increases competition for  $\text{OH}^\bullet$  and ozone formation slows, further limiting radical availability. We find VOC classes which readily produce their own photolytic source of radicals (e.g. HCHO), and hence are less affected by  $\text{NO}_x$ -inhibition, have less pronounced scaling breaks (Ainslie, 2004). Since smog chambers are often plagued by unaccounted radical sources (Carter et al., 1981), it is possible that such sources reduce the  $\text{NO}_x$ -inhibition of the smog chamber mixtures; in fact Tonnesen and Jeffries (1994) show the GRS mechanism, developed from CSIRO smog chamber data, shows noticeably less  $\text{NO}_x$ -inhibition than the CB4 chemical mechanism.

## 6. Conclusions

In this paper we have explored the similarity relationship using semi-quantitative models and model output. Our analysis shows the scaling break arises from a change in the temporal evolution of ozone production. Our analysis suggests the initial relative abundance of VOC to  $\text{NO}_x$  ( $R$ -value) determines whether the switch in processes will occur within a fixed process time. The  $R$ -value at the break ( $\beta$ ) can be used to create a process-based ridgeline on an isopleth diagram that lies above a sensitivity based ( $d[\text{O}_3]_{\text{max}}/d[\text{NO}_x]_0=0$ ) ridgeline. As Sillman (1999) points out, the actual split between VOC and  $\text{NO}_x$ -sensitive is better viewed as a broad transition region rather than a sharp line. For this reason, we propose the most natural division of an ozone response

surface is into three regions: VOC sensitive (or NO<sub>x</sub>-saturated) region, ridge or transition region and a region where maximum ozone concentrations scale with initial NO<sub>x</sub> concentrations only (NO<sub>x</sub>-only scaling region). We suggest  $\beta$  marks the boundaries between the VOC-sensitive/ridge regions.

Our analysis has allowed us to developed a set of properties that characterize ozone-precursor relationships. These findings, all based on model output, generalizes the IER model, developed from smog chamber data. These properties have also enabled us to include a temporal dependence into the similarity parameterization of Ainslie and Steyn (2005) (Eq. 9) which can be used to quantitatively describe ozone formation.

The scaling methods were applied to smog chamber data allowing observations from a variety of experiments, having different initial precursor concentrations, temperatures and actinic fluxes, to be presented on a common graph. Only a slight break in the corresponding Weibull transformed observations was seen with the chamber data. The present analysis cannot tell if the lack of sharp break is due to experimental uncertainties stemming from the presence of the chamber membrane (especially if these act as additional radical sources); a bias in the test conditions with more experiments performed at  $R$ -values below the scaling break; or if chemical mechanisms overly enhance the process leading to the scaling break. More research is needed in this area, perhaps making use of ambient ozone concentrations drawn from a variety of environments.

## Appendix

The IER model consists of a simple set of formulae which quantitatively describe ozone formation developed from the analysis of an extensive set of outdoor smog chamber experiments (Johnson, 1984). The model predicts that initially, smog potential ( $PSP$ ) ( $[O_3](t) - [O_3](0) - [NO](0) + [NO](t)$ ) increases linearly with  $j(t)$  and initial VOC concentra-

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tion until reaching a constant value:

$$PSP = \begin{cases} kf(T)[VOC]_o j(t), & kf(T)[VOC]_o j(t) < \beta[NO_x]_o \\ \beta[NO_x]_o, & kf(T)[VOC]_o j(t) \geq \beta[NO_x]_o \end{cases} \quad (A1)$$

where  $k$  is a constant and  $f(T)$  is a temperature dependent measure of VOC reactivity. The corresponding parameterization for  $PSP$  production ( $P(PSP)$ ) has:

$$P_j(PSP) = \begin{cases} kf(T)[VOC]_o, & kf(T)[VOC]_o j(t) < \beta[NO_x]_o \\ 0, & kf(T)[VOC]_o j(t) \geq \beta[NO_x]_o \end{cases} \quad (A2)$$

From Eqs. (A1) and (A2) we find:

- $P_j(PSP)$  increases with  $[VOC]_o = R [NO_x]_o$  (Property II)
- the parameterization for  $P(PSP_j)$  shows a discontinuity when  $k f(T)[VOC]_o = \beta[NO_x]_o$  (Property III).
- The transition time ( $j^*$ ) occurs when:

$$j(t) = \frac{\beta [NO_x]_o}{k f(T) [VOC]_o} \propto \frac{1}{R} \quad (A3)$$

i.e. transition occurs sooner for larger  $R$ -values (Property IV).

$$- \int_0^{j^*} P_j(PSP) dj = \int_0^{j^*} P_j(PSP) dj = \beta[NO_x]_o \quad (A4)$$

showing the area under the production curve at the transition, as well as the total area are functions of initial  $NO_x$  concentration (Property V).

- The parameterization for  $[PSP]$  is a function of  $[VOC]_o j(t) = [NO_x]_o R j(t)$  which scales with  $R \times j^c(t)$  with  $c=1$  (Property VI).

Thus we can think of the IER model as a parameterization of  $PSP$  production, using a step function, which satisfies Properties II–V.



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**Table 1.** Test conditions for the 10 CSIRO smog chamber experiments. The VOC is a mixture of organics typically found in the Sydney air (Hess et al., 1992). The total  $\text{NO}_2$  photolysis rate is given by  $J$  while  $j_{av}$  is its average value over the duration of the experiment.

Test name	Month	Date	$[\text{VOC}]_o$ (ppb)	$[\text{NO}_x]_o$ (ppb)	R	$T_{av}$ (C)	J	$j_{av}$ ( $\text{s}^{-1}$ )
340P	September	21	200.0	19.6	10.2	22.5	228.3	0.00634
340L	September	21	200.0	19.7	10.2	22.2	227.6	0.00632
341P	September	28	200.0	53.4	3.7	18.0	231.2	0.00642
341L	September	28	200.0	24.8	8.1	17.4	226.9	0.00664
342P	October	5	100.0	54.0	1.9	22.2	248.4	0.00657
342L	October	5	200.0	55.6	3.6	21.2	250.5	0.00633
343P	October	10	400.0	53.6	7.5	21.5	250.8	0.00664
343L	October	10	100.0	54.5	1.8	21.5	247.3	0.00654
344P	October	11	200.0	44.6	4.5	26.9	251.3	0.00665
344L	October	11	400.0	46.2	8.7	26.0	253.9	0.00672
345P	October	19	200.0	100.4	2.0	28.5	250.8	0.00697
345L	October	19	400.0	97.2	4.1	28.8	245.0	0.00717
359P	March	15	100.0	30.9	3.2	25.2	174.6	0.00809
359L	March	15	50.0	28.2	1.8	25.0	187.1	0.00800
360P	March	19	50.0	31.4	1.6	22.8	215.2	0.00664
360L	March	19	100.0	30.3	3.3	22.7	221.6	0.00648
361P	March	21	100.0	17.4	5.7	22.4	225.0	0.00622
361L	March	21	100.0	31.1	3.2	22.5	230.9	0.00639
362P	March	26	50.0	19.3	2.6	25.2	241.9	0.00640
362L	March	26	100.0	18.9	5.3	24.7	237.9	0.00630

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**Table 2.** Similarity parameters and power law exponents used to fit the CSIRO data.

a	b	c	$\gamma$	$\alpha_1$	$\alpha_2$	$\beta'$	$\lambda$
0.25	7.6	1.5	13.1	1.4	0.88	17 000	0.96

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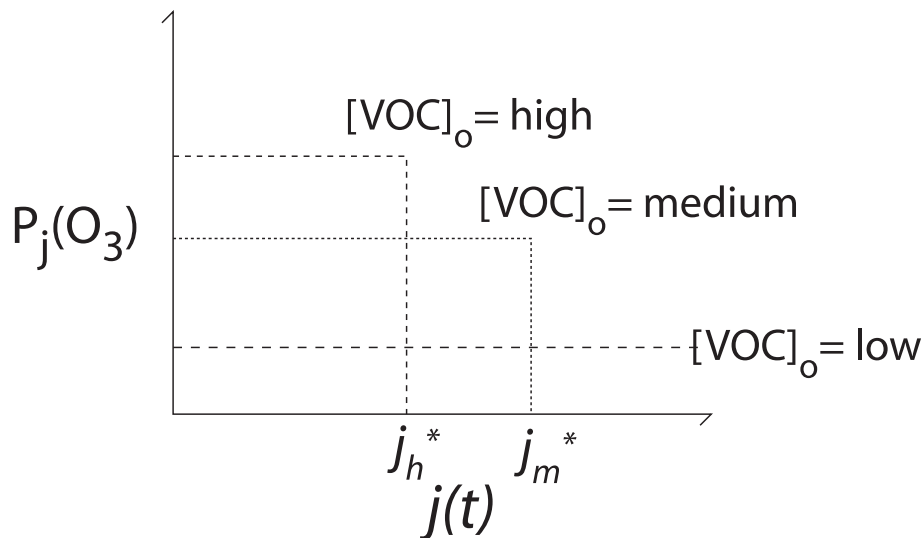
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**Table 3.** Summary of statistics for the similarity scatter plots.

Slope	Intercept (ppb)	$R^2$	RMSE (ppb)	MBE (ppb)
0.94	4.7	0.96	12.5	0.4

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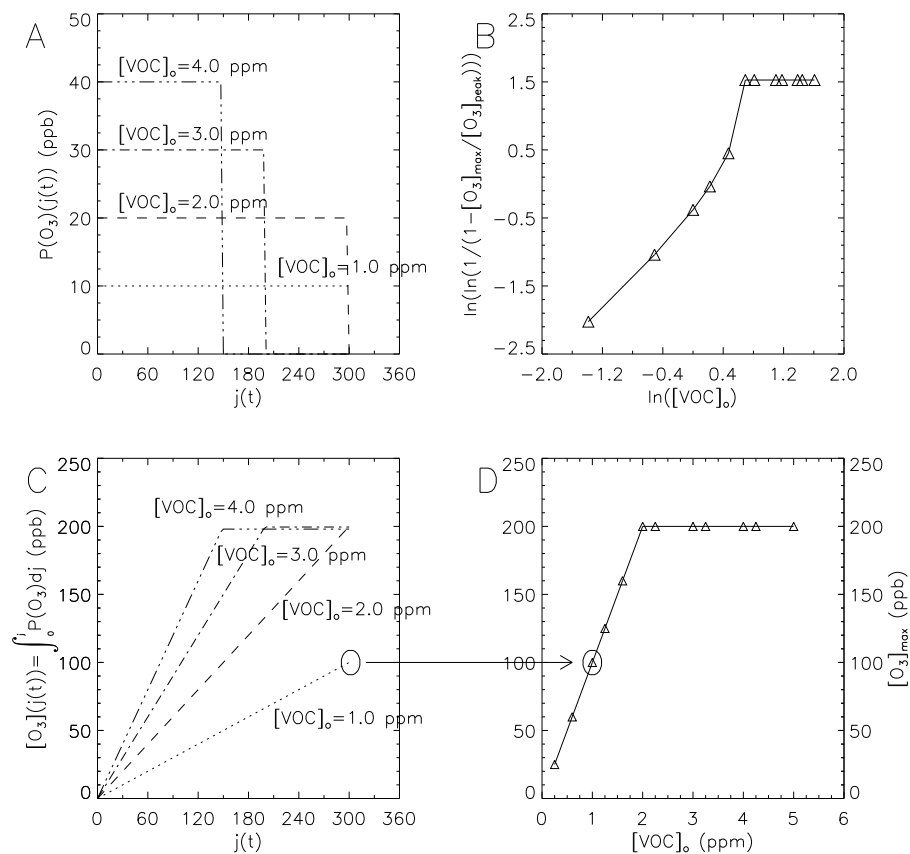
**Fig. 1.** Three members of a family of a step-function model for ozone production. The high  $[VOC]_o$  switches to zero production at  $j_h^*$ , the medium at  $j_m^*$  while the low VOC never switches in the examined process time.

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**Fig. 2.** Ozone production **(A)**, Weibull transformed maximum ozone concentration **(B)**, concentration **(C)** and maximum concentration **(D)** for a step function model. The arrow between Figs. (C) and (D) highlights the relationship between final ozone concentration and maximum ozone concentration.

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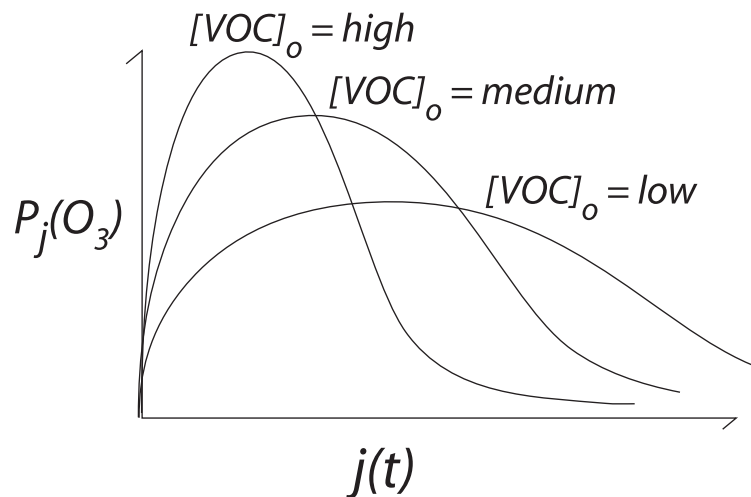
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**Fig. 3.** Three members of a family of a model for ozone production which increases, reaches a peak value and decays.

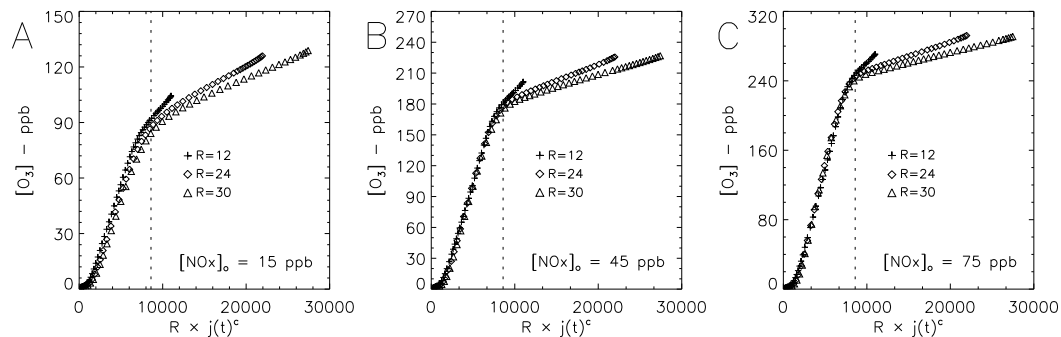
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**Fig. 4.** Ozone concentration as a function of  $R \times j(t)^c$  ( $c=1.2$ ) for nine different initial  $\text{NO}_x$  and VOC conditions. In Fig. (A), all simulations started with 15 ppb  $\text{NO}_x$  and with 45 ppb in (B) and 75 ppb in (C). In each figure initial VOC concentrations are chosen so that the resulting  $R$ -values are: 12, 24, 30. Also shown in each plot is a vertical line at  $R \times j(t)^c = 8600$ .

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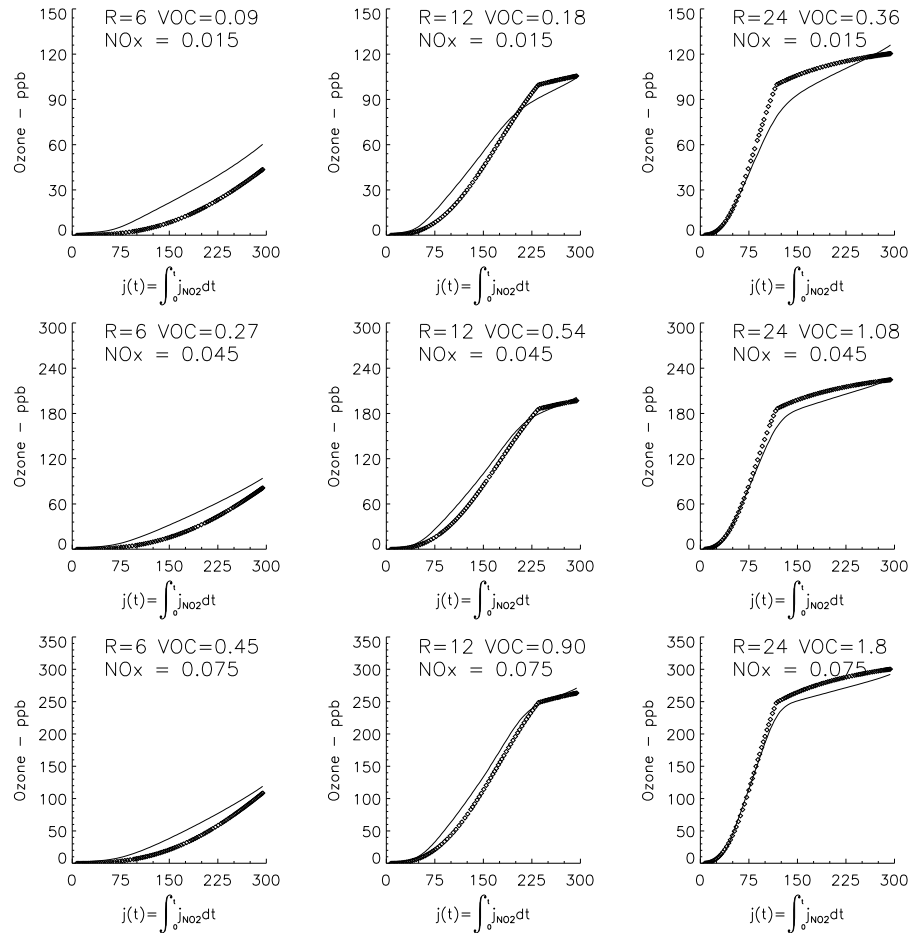
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**Fig. 5.** Ozone concentration versus time for a series of nine box model simulations. Solid lines represent model output using the RADM2 mechanism (Stockwell et al., 1990) and circles represent results based on Eq. (4).

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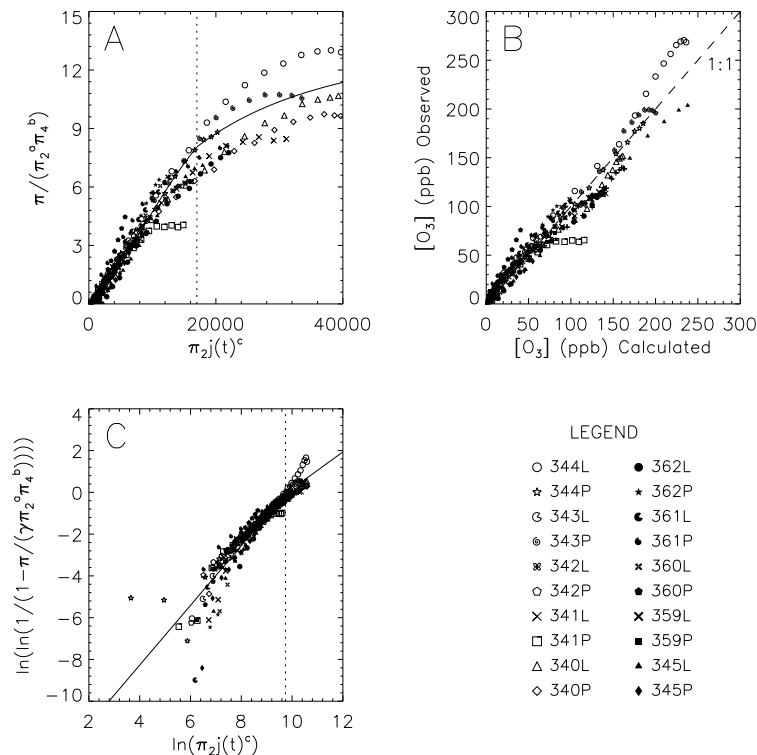
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**Fig. 6.** Comparison of the similarity function and the CSIRO smog chamber data. Figure (A) shows the non-dimensional data as a function of  $R_j(t)^c$ . Also shown (solid line) is the similarity function which best fits the data and the location of the scaling break (dotted line). The model has been used to calculate ozone concentrations which are compared with the actual concentrations in Fig. (B). Figure (C) shows the data in the Weibull plane along with the similarity function (solid line) and the scaling break (dotted line).

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