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Peroxy radical chemistry and the control of ozone photochemistry at Mace Head, Ireland during the summer of 2002

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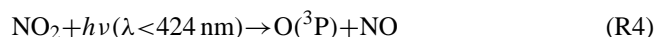
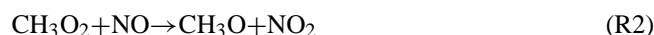
Abstract. Peroxy radical ($\text{HO}_2 + \Sigma\text{RO}_2$) measurements, using the Peroxy Radical Chemical Amplification (PERCA) technique at the North Atlantic Marine Boundary Layer Experiment (NAMBLEX) at Mace Head in summer 2002, are presented and put into the context of marine, boundary-layer chemistry. A suite of other chemical parameters (NO , NO_2 , NO_3 , CO , CH_4 , O_3 , VOCs, peroxides), photolysis frequencies and meteorological measurements, are used to present a detailed analysis of the role of peroxy radicals in tropospheric oxidation cycles and ozone formation. Under the range of conditions encountered the peroxy radical daily maxima varied from 10 to 40 pptv. The diurnal cycles showed an asymmetric shape typically shifted to the afternoon. Using a box model based on the master chemical mechanism the average model/measurement agreement was 2.5 across the campaign. The addition of halogen oxides to the model increases the level of model/measurement agreement, apparently by respeciation of HO_x . A good correlation exists between $j(\text{HCHO}) \cdot [\text{HCHO}]$ and the peroxy radicals indicative of the importance of HCHO in the remote atmosphere as a HO_x source, particularly in the afternoon. The peroxy radicals showed a strong dependence on $[\text{NO}_x]$ with a break point at 0.1 ppbv, where the radicals increased concomitantly with the reactive VOC loading, this is a lower value than seen at representative urban campaigns. The $\text{HO}_2/(\text{HO}_2 + \Sigma\text{RO}_2)$ ratios are dependent on $[\text{NO}_x]$ ranging between 0.2 and 0.6, with the ratio increasing linearly with NO_x . Significant night-time levels of peroxy radicals were measured up to 25 pptv. The contribution of ozone-alkenes and NO_3 -alkene chemistry to night-time peroxy radical production was shown to be on average 59 and 41%. The campaign mean net ozone production rate was 0.11 ± 0.3 ppbv

h^{-1} . The ozone production rate was strongly dependent on $[\text{NO}]$ having linear sensitivity ($d\ln(P(\text{O}_3))/d\ln(\text{NO})=1.0$). The results imply that the $\text{N}(\text{O}_3)$ (the in-situ net photochemical rate of ozone production/destruction) will be strongly sensitive in the marine boundary layer to small changes in $[\text{NO}]$ which has ramifications for changing NO_x loadings in the European continental boundary layer.

1 Introduction

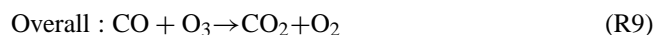
Peroxy radicals (HO_2 and RO_2 , predominantly CH_3O_2 in semi-polluted atmospheres) can be thought of as the intermediates between the hydroxyl (OH) radical and ozone formation or destruction (Monks, 2005). Peroxy radicals also control the removal of primary pollutants such as NO_x ($\text{NO} + \text{NO}_2$) and Volatile Organic Compounds (VOCs). Understanding the radical chemistry that controls ozone formation will improve our basic understanding of tropospheric photochemistry and the effect of natural and man-made emissions on ozone formation.

The relative contribution of ozone production and loss processes in the troposphere is highly sensitive to competition between the reaction of peroxy radicals with NO and their self- and cross-reactions to form peroxides. In the presence of NO_x , the reaction of peroxy radicals with NO leads to the formation of NO_2 , which, upon photolysis, forms ozone:



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In conditions of low $[\text{NO}_x]$, a catalytic cycle leads to net ozone destruction, shown here for the reaction with CO:



OH oxidation and other VOCs forms peroxy radicals:



Both urban and rural environments are affected by air pollution of photochemical origin, and the modelling of photochemical ozone formation in the British Isles from European emissions are important for the policy-makers in order to develop emission-reduction targets for ozone precursors (Metcalf et al., 2002, and Derwent et al., 2003). Ozone is one of the major components of photochemical smog, together with contributions from compounds such as carbonyls, peroxy acetyl nitrates (PANs) and various nitrogen oxides. It has been seen in past studies in the relatively clean rural/marine conditions of Mace Head (Salisbury et al., 2001, 2002) during the EASE 96 and 97 (Eastern Atlantic Summer/Spring Experiment) and in the very clean air of Cape Grim in Tasmania (Monks et al., 1998, 2005b)¹ at the SOAPEX 2 (Southern Ocean Atmospheric Photochemistry EXperiment) campaign in 1998 that ozone formation is part of a natural cycle that can be easily perturbed by pollution events.

Tropical maritime air which is depleted in ozone can be advected to Mace Head over a distance of several thousand kilometres without significant net ozone formation occurring (Derwent et al., 1998). Measurements at Mace Head found that the site experienced more photochemical ozone production than destruction during the EASE 96 and 97 campaigns (Salisbury et al., 2002) and at ATAPEX-95 (Atlantic Atmospheric Photochemistry Experiment) (Carpenter et al., 1997). Cape Grim experienced far more days with net ozone destruction as in SOAPEX 1 in 1995 (Monks et al., 1998, 2000, and Carpenter et al., 1997) and in 1998 at SOAPEX 2 (Monks et al., 2005). Andres-Hernandez et al. (2001) also found that during the Atlantic and Southern Indian Ocean cruise of AEROSOL 99 net ozone destruction predominated.

Penkett et al. (1997) showed that the relationship between the sum of peroxy radicals and the ozone photolysis rate coefficient (to the singlet oxygen atom), $j(\text{O}^1\text{D})$ can serve as a diagnostic for photochemical ozone production and destruction. In high NO_x conditions $\text{HO}_2 + \Sigma\text{RO}_2$ is generally proportional to $j(\text{O}^1\text{D})$ and in clean conditions, to $\sqrt{j(\text{O}^1\text{D})}$

¹Monks, P. S., Salisbury, G., and Fleming, Z. L., et al.: The role of peroxy radicals in photochemical destruction of ozone at mid-latitudes in the Southern Hemisphere, to be submitted, 2005b.

(Penkett et al., 1997; Monks et al., 1998; Zanis et al., 1999; Creasey et al., 2003). The shape of $j(\text{O}^1\text{D})$ throughout the day produces a typical summer peroxy radical diurnal cycle with maximum values towards solar noon, and minimum values during the night. Photolysis of other photo-labile compounds (e.g. HCHO and HONO) becomes noticeable in the early morning or evening, when the light is of longer wavelengths than those at which ozone photolysis occurs, and can lead to a broader peroxy radical diurnal cycle than that expected from ozone photolysis alone.

In the absence of photochemistry, there is a series of nighttime peroxy radical-producing channels. NO_3 (nitrate) radical (Allen et al., 1999; Penkett et al., 1999; Salisbury et al., 2001; Geyer et al., 2003) and ozone reactions with alkenes (Hu and Stedman, 1995; Rickard et al., 1999; Salisbury et al., 2001) were found to be two dominant channels in the marine influenced atmosphere.

In this paper, by use of peroxy radical measurements coupled to a suite of supporting trace species measurements, the photochemical environment of Mace Head is explored. In particular, the shape, concentration and form of the diurnal cycle are explored in relation to both primary production and the potential contribution of the photolysis of secondary compounds such as carbonyls (e.g. HCHO). The dependence of peroxy radical concentration with changing NO_x and VOC is described. The role of both NO_3 and O_3 -alkene reactions as night time source of peroxy radicals is investigated. Finally, role of the peroxy radicals in the in situ photochemical formation and destruction of ozone is quantified.

2 Experimental

2.1 Site

NAMBLEX took place from 27 July to 2 September 2002 at the Mace Head Atmospheric Research Station ($53^\circ 20' \text{N}$, $9^\circ 54' \text{W}$). Mace Head is located on the west coast of Ireland, 88 km west of Galway city, and is in the path of the mid-latitude cyclones which frequently traverse the North Atlantic. Heard et al. (2005) describe the location in more detail, together with the local meteorology of the site, and Norton et al. (2005) provide a detailed analysis of the specific boundary layer conditions encountered during NAMBLEX. The prevailing wind direction is from a westerly marine sector but significant pollution events also reach the site from European continental air-masses, from easterly directions.

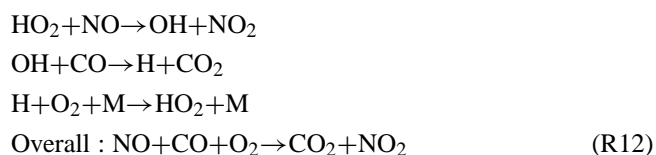
2.2 Peroxy radical measurements (PERCA)

Measurements of peroxy radicals ($\text{HO}_2 + \Sigma\text{RO}_2$) were carried out using the jointly operated University of Leicester – University of East Anglia (UEA) PEROxy Radical Chemical

Amplifier (PERCA IV) instrument, reported for the first time in Green et al. (2005) and Fleming et al. (2006).²

The technique was pioneered by Cantrell et al. (1984) and described by Clemitshaw et al. (1997), Monks et al. (1998) and Green et al. (2003) and the current apparatus uses a dual channel inlet and detection system (as in Cantrell et al., 1996).

Briefly, the method relies upon the HO₂ and OH radical-catalysed conversion of NO and CO into CO₂ and NO₂ respectively, through addition of NO and CO into the inlet region viz.,



Organic peroxy radicals are readily converted into HO₂ in the presence of NO with varying efficiencies (Ashbourn et al., 1998). The yield of both CO₂ and NO₂ is equal to CL*($\sum \text{RO}_2 + [\text{HO}_2] + [\text{OH}]$), where CL is the chain length, i.e. the number of HO₂/OH inter-conversion cycles that occur before radical termination. The ratio of [HO₂]/[OH] ranges from ~50–200 in the atmosphere, therefore the PERCA technique effectively measures the sum of inorganic and organic peroxy radicals. The yield of NO₂ is measured using commercial LMA-3 detectors (calibrated daily using NO₂ permeation sources) and this is converted into [HO₂ + $\sum \text{RO}_2$] using $\Delta[\text{NO}_2]/\text{CL}$. The chain length was calculated on a weekly basis, using a calibration source based upon the photolysis of CH₃I at 253.7 nm to yield CH₃O₂ at varying concentrations (Clemitshaw et al., 1997).

Background [NO₂] signals (caused by the reaction of ambient ozone with NO in the inlet) were measured by changing the addition points of the reagent gases, so that the amplification reactions are not initiated. The overall radical levels are calculated by subtracting the termination signal from the amplification signal.

The dual-inlet system comprises two distinct sampling systems, inlet reaction sites and detectors. The advantage of this dual-inlet system is that the two systems are run out of phase in two modes, with one in amplification and the other in termination mode. Switching the two inlets between phases on a minute time scale leads to a continuous amplification and termination signal.

Mihele et al. (1998, 1999) have shown that the chain length of a chemical amplifier is reduced in the presence of atmo-

spheric water vapour (see also Reichert et al., 2003). Salisbury et al. (2002) demonstrated that the chain length of the Leicester-UEA PERCA instrument falls approximately linearly with increasing specific humidity. From a series of laboratory experiments, a humidity correction factor equation (using ambient humidity and inlet temperatures) was derived and applied to all PERCA data. In order to minimise the applied correction factor, the inlet temperature was kept above ambient temperatures at the constant value of 30°C. The humidity correction factor varied between 1.5 and 2.5 during NAMBLEX. (see also Fleming et al., 2006).

The accuracy and precision of the dual-inlet PERCA have been assessed in detail by Fleming (2006). The overall uncertainty for any given peroxy radical measurement is 38% (at 1 σ) from a combination of uncertainties associated with the radical calibration, NO₂ quantification and humidity correction. The precision on a 1 pptv measurement averaged over a minute assessed from the reproducibility of the radical calibration was 15%. The detection limit of the instrument was of the order of 0.5 pptv at a S/N of 1 on a 1 min average.

The PERCA instrument inlet box was securely attached 6 m above ground level to a tower on a temporary laboratory building (sea container) close to the main shoreline site and an umbilical line carried the reagent gases and sample lines down to the main rack in the laboratory (see Heard et al., 2005, for a site plan). The PERCA instrument took measurements continuously during the campaign in minute averages and analysis was carried out with ten minute- or hourly-averaged data.

2.3 Other measurements

Meteorological data were obtained from the site's fixed meteorological station, which recorded ambient air temperature, relative humidity, wind speed and wind direction. The other instruments were housed in the permanent cottages of the site or in similar self-contained temporary buildings (sea-containers). Details of the instruments, their detection limits and accuracy are given in Heard et al. (2005).

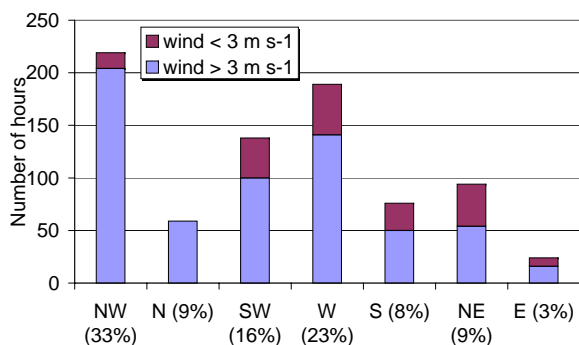
2.4 Modelling studies

The Master Chemical Mechanism (currently MCMv3.1, available online at <http://mcm.leeds.ac.uk/MCM/>) Developed by Jenkin et al. (1997) and subsequently refined and updated by Saunders et al. (2003); Jenkin et al. (2003) and Bloss et al. (2005a, b) MCMv3.1 contains the oxidation mechanisms of 135 primary emitted VOCs. This mechanism was added to a campaign optimised box model used to simulate HO₂, RO₂ and OH radical concentrations constrained with H₂, CO, CH₄, O₃, NO, NO₂, HCHO, measured VOCs, chloroform, temperature and various photolysis rates measurements. In general, the most complete model shown in this analysis is termed "full-oxy" and is detailed extensively in Sommariva et al. (2006a). In addition, the model was then

²Fleming, Z. L., Monks, P. S., Rickard, A. R., Bandy, B. J., Brough, N., Green, T. J., Reeves, C. E., and Penkett, S. A.: Seasonal dependence of peroxy radical concentrations at a Northern hemisphere marine boundary layer site during summer and winter: Evidence for photochemical activity in winter, *Atmos. Chem. Phys. Discuss.*, submitted, 2006.

Table 1. Campaign air-mass sector-averaged chemical and physical parameters (1σ standard deviation given in brackets.)

	NW	N	SW	W	S	NE	E
$\text{HO}_2 + \sum \text{RO}_2$ /pptv	7.6 (6)	7.9 (4)	7.1 (6)	10.2 (8)	13.7 (10)	13.4 (10)	13.5 (7)
NO_x /pptv	151 (39)	73 (68)	111 (136)	63 (33)	230 (418)	275 (212)	352 (72)
NO /pptv	19 (47)	10 (14)	13 (12)	10 (7)	24 (55)	31 (21)	34 (26)
CH_4 /ppbv	1813 (20)	1816 (32)	1785 (30)	1800 (13)	1821 (30)	1863 (50)	1925 (29)
CO /ppbv	90 (12)	81 (3)	77 (14)	83 (6)	82 (5)	112 (18)	149 (10)
O_3 /ppbv	33 (4)	29 (4)	25 (7)	24 (12)	23 (12)	32 (4)	39 (2)
H_2O_2 /pptv	0.19 (0.19)	0.21 (0.14)	0.09 (0.07)	0.20 (0.10)	0.14 (0.05)	0.18 (0.11)	0.37 (0.10)
HCHO /ppbv	1.38 (0.18)	1.20 (0.08)	1.22 (0.17)	1.28 (0.26)	1.34 (0.12)	1.62 (0.22)	2.09 (0.17)
DMS /pptv	120 (98)	233 (66)	388 (264)	244 (227)	100 (107)	131 (88)	23 (24)
Isoprene /pptv	13 (24)	4 (1)	2 (1)	5 (6)	30 (61)	72 (119)	15 (26)
Benzene /pptv	29 (14)	24 (17)	20 (4)	27 (17)	18 (7)	62 (22)	114 (11)
Methanol /pptv	1068 (365)	852 (233)	1536 (384)	1086 (370)	1204 (291)	1747 (630)	1559 (596)

**Fig. 1.** Air-mass sector divisions for winds less than and greater than 3 m s^{-1} .

also constrained with 23 hydrocarbons, 3 oxygenates and 2 peroxides and to temperature and photolysis measurements. OH and HO_2 model/measurement comparisons are reported in Smith et al. (2006), Sommariva et al. (2006a) and nighttime HO_2 and RO_2 in Sommariva et al. (2006b). Model results at NAMBLEX for HO_2 were in much better agreement with the measurements when the model was additionally constrained to measured halogen oxides (Sommariva et al., 2006a; Bloss et al., 2005c).

3 Results

3.1 Meteorological conditions and chemical climatology

Local wind speed and direction measurements were recorded on an hourly basis during NAMBLEX. During the EASE campaigns Salisbury et al. (2002), and Rickard et al. (2002) divided all the data into five sectors, according to local wind direction.

More detailed boundary layer structure measurements in combination with back trajectory analysis during the NAMBLEX campaign is described by Norton et al. (2005). However, caution must be used when assigning air-mass sectors only according to local in-situ wind direction as measurements during the NAMBLEX data did not necessarily correspond with the origin of the air-mass back-trajectory analysis. For example, there were a number of land-sea breeze events (particularly at night). From 1 to 5 August, these breeze events brought local easterly winds to the site during westerly trajectories.

The British Atmospheric Data Centre's (BADC) trajectory service (<http://www.badc.nerc.ac.uk>) was used to plot five-day air-mass back-trajectories at six-hourly intervals. According to the origin of the air-masses of these trajectories, a new division into seven areas of origin was developed as shown in Tables 1, 2 and 3. The most common air-mass sector was the north-westerly (NW), followed by westerly (W) and then south-westerly (SW) as shown in Fig. 1 (see also Heard et al., 2005). The effects of local sea and coastal breezes were removed from the analysis by only selecting data where the local wind speed was greater than 3 m s^{-1} (Salisbury et al., 2002).

Table 1 shows the average $\text{HO}_2 + \sum \text{RO}_2$, NO_x , CO, CH_4 , O_3 , CH_3OOH , H_2O_2 , HCHO, DMS, isoprene, benzene and methanol mixing ratios and $j(\text{O}^1\text{D})$ for the corresponding air-mass sectors. The data in Table 2 corresponds to the daylight (06:00–19:00) averages. Table 3 shows the night-time concentrations of the same species, as well as $[\text{NO}_3]$ and total alkenes.

Heard et al. (2005) provides a comprehensive overview of all the other species and supporting measurements made during NAMBLEX and their respective time series.

Table 2. Daylight-only (06:00 to 19:00 UT) air-mass sector-averaged chemical and physical parameters (1σ standard deviation given in brackets.)

	NW	N	SW	W	S	NE	E
$\text{HO}_2 + \sum \text{RO}_2$ /pptv	10.6 (6)	9.6 (5)	8.5 (7)	15.1 (9)	16.4 (11)	21.0 (9)	17.4 (9)
NO_x /pptv	133 (256)	88 (77)	80 (39)	65(30)	190 (497)	216 (112)	342 (93)
NO /pptv	31 (61)	16 (17)	15 (11)	13 (7)	34 (66)	43 (20)	57 (21)
CH_4 /ppbv	1814 (22)	1818 (36)	1783 (30)	1799(13)	1817 (29)	1852 (44)	1904 (29)
CO /ppbv	90 (12)	80 (3)	76 (15)	83(5)	82 (5)	111 (18)	143 (8)
O_3 /ppbv	33 (4)	28 (4)	25 (7)	26(12)	20 (14)	33 (4)	39 (3)
$j(\text{O}^1\text{D}) \times 10^{-6}/\text{s}^{-1}$	7.8 (7)	6.9 (6)	5.9 (6)	7.2 (5)	8.5 (6)	8.1 (6)	7.3 (8)
$\text{N}(\text{O}_3)$ /ppbv h^{-1}	0.023 (0.2)	0.028 (0.1)	-0.019 (0.08)	0.025 (0.2)	0.087 (0.3)	0.51 (0.5)	0.47 (0.3)
H_2O_2 /pptv	0.20 (0.2)	0.17 (0.1)	0.08 (0.1)	0.22 (0.1)	0.14 (0.04)	0.17 (0.1)	0.41 (0.08)
HCHO /ppbv	1.5 (0.2)	1.2 (0.1)	1.2 (0.1)	1.4 (0.3)	1.4 (0.2)	1.7 (0.2)	2.1 (0.2)

Table 3. Night-time-only (19:00–06:00) air-mass sector-averaged chemical and physical parameters (1σ standard deviation given in brackets.)

	NW	N	SW	W	S	NE	E
$\text{HO}_2 + \sum \text{RO}_2$ /pptv	4.2 (2)	5.6 (1)	4.4 (3)	4.1 (2)	6.4 (5)	7.7 (2)	10.5 (3)
NO_x /pptv	174 (521)	59 (57)	152 (200)	61 (36)	301 (214)	344 (274)	359 (54)
NO /pptv	4.8 (3)	4.7 (4)	9.6 (13)	6.1 (6)	4.9 (3)	17.0 (12)	15.2 (8)
CH_4 /ppbv	1813 (17)	1814 (28)	1787 (31)	1802 (14)	1827 (33)	1877 (53)	1941 (16)
CO /ppbv	90 (13)	81 (2)	78 (12)	84 (6)	81 (3)	113 (19)	154 (11)
O_3 /ppbv	32 (4)	30 (2)	26 (8)	22 (12)	28 (6)	31 (3)	40 (2)
NO_3 /pptv	4.7 (2)	5.9 (3)	nd	4.5 (2)	2.7 (1)	5.7 (5)	11.7 (6)
HCHO /ppbv	1.28 (0.16)	1.16 (0.05)	1.25 (0.23)	1.14 (0.17)	1.33 (0.07)	1.54 (0.23)	nd
Alkenes /pptv	60 (48)	33 (11)	71 (57)	64 (30)	71 (47)	187 (141)	188 (82)

nd: No Data

3.1.1 Peroxy radical levels and diurnal cycles

The data in Table 1 and the peroxy radical time series in Fig. 2 show that peroxy radical concentrations are highest when the air is of continental origin (S, E and NE air-mass sectors). Also the concentrations of NO_x , isoprene and formaldehyde were significantly higher in these sectors than in the marine W sector.

The lowest peroxy radical concentrations during the campaign were near the detection limit of the instrument (see experimental section) at around 1 pptv but were rarely less than 4 pptv, even at night. The maximum peroxy radical concentration reached 60 pptv for the occasional short-lived spike and 40 pptv for midday maximum values (see Fig. 2). Generally, the average day-time peroxy radical concentrations were between two and three times higher than night-time levels.

The campaign average $[\text{HO}_2 + \sum \text{RO}_2]$, $[\text{NO}_x]$ and $j(\text{O}^1\text{D})$ diurnal cycles and the W and NE air-mass sector-averaged

diurnal cycles are shown in Fig. 3. Peroxy radicals track the $j(\text{O}^1\text{D})$ diurnal cycle fairly closely, with the cycle being shifted towards the end of the day as high midday concentrations persist well into the afternoon. This form of asymmetrical diurnal cycle has been noted before in low- NO_x environments by Monks et al. (1996) at Cape Grim, Carpenter et al. (1997) at Mace Head in 1995, and in high NO_x environments by Holland et al. (2003) and Mihelcic et al. (2003) at the BERLIOZ campaign near Berlin. In each of these environments there are different mechanisms that drive this asymmetry. The peroxy radical diurnal cycle for the W air-mass sector has less relative variability, tracking $j(\text{O}^1\text{D})$ with a slight bias in maxima towards the afternoon. The NE sector peroxy radical diurnal is broader in both the morning and the evening. The S sector diurnal cycle (not shown) follows a similar pattern to the NE average in that it has a broader shape in both the morning and the afternoon.

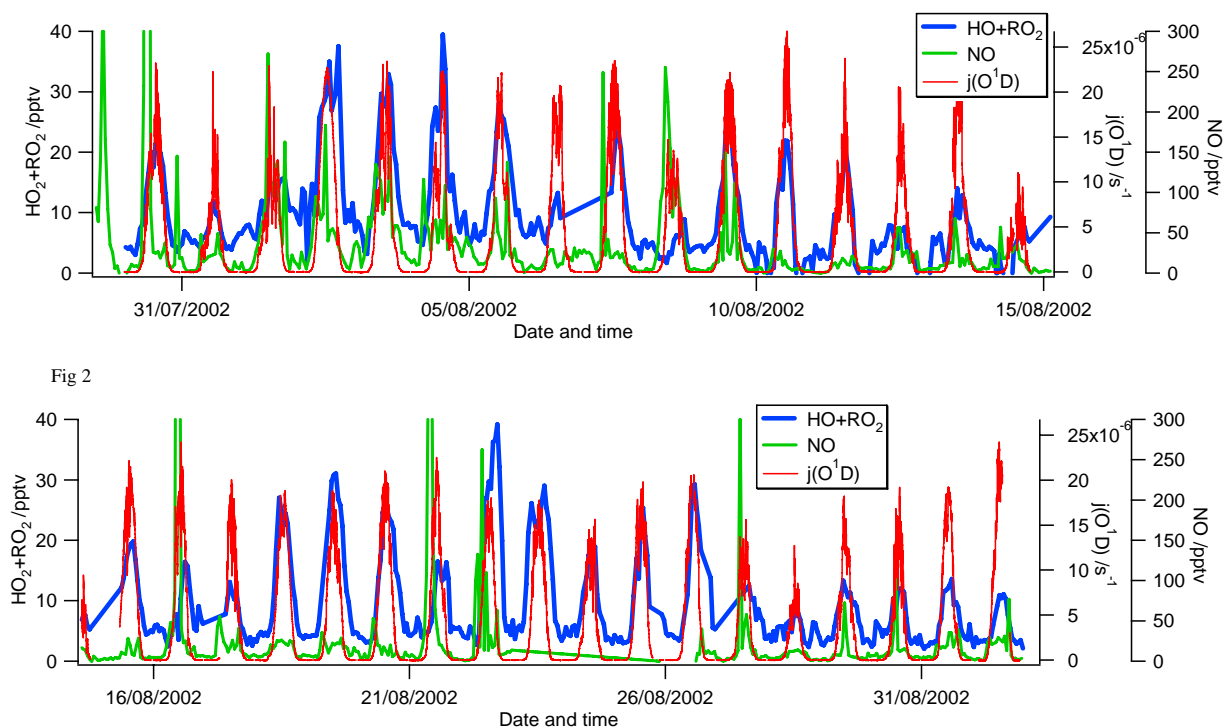


Fig. 2. [HO₂ + ΣRO₂], $j(\text{O}^1\text{D})$ and [NO] campaign time series.

Table 4. Sensitivity of the Model/Measurement HO₂ + ΣRO₂ agreement for varying chemical complexity.

Model Run ^a	Model/Measurement
CO + CH ₄ chemistry (Het with IO chemistry) – constrained to measurements of CO and CH ₄	3.3±1.8
Full Chemistry (Het with IO chemistry) – constrained to measurements of CO, CH ₄ , 23 hydrocarbons and chloroform	2.6±1.3
Full + Oxy – constrained to measurements of CO, CH ₄ , 23 hydrocarbons, chloroform and 3 oxygenates	2.5±1.3
Full + Oxy (Het with IO chemistry) – constrained to measurements of CO, CH ₄ , 23 hydrocarbons, chloroform and 3 oxygenates	2.3 ₄₁ ±1.2
Full + Oxy + Per (Het with IO chemistry) – constrained to measurements of CO, CH ₄ , 23 hydrocarbons, chloroform, 3 oxygenates and 2 peroxides	2.3 ₄₃ ±1.2

a: Full details of model runs in Sommariva et al. (2006); Het with IO chemistry-model additionally constrained to measured IO and $j(\text{HOI})$ using a transition regime expression to calculate the heterogeneous uptake of the gas-phase species.

An extensive modelling study by Sommariva et al. (2006a) has investigated the impact of oxygenate and halogen chemistry on the radical chemistry. Table 4 shows the sensitivity of the average model/measurement agreement for HO₂ + ΣRO₂ with varying model assumptions. Figure 4a shows the results using the full oxygenate chemistry constrained to measured

IO with corrected heterogeneous uptake model and measured peroxy radical concentrations for 15–22 August, and the same model run with the measured HO₂ from the FAGE instrument (Smith et al., 2006). For the model run with full oxygenate chemistry constrained to measured IO with corrected heterogeneous uptake (see Table 4), the peroxy radi-

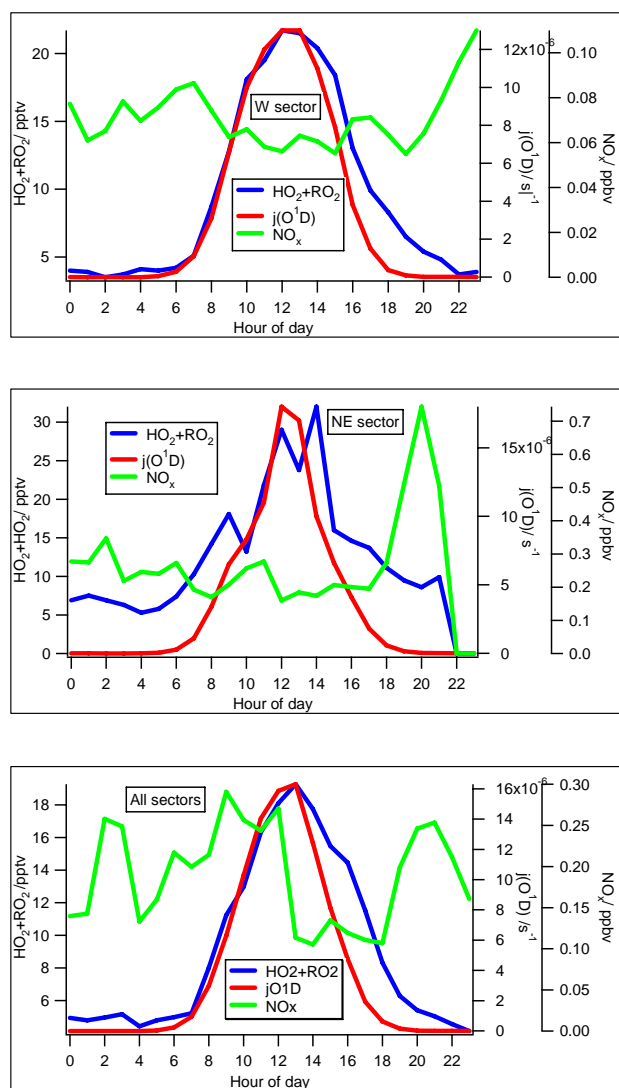


Fig. 3. Hourly-averaged $[\text{HO}_2 + \sum\text{RO}_2]$, $[\text{NO}_x]$ and $j(\text{O}^1\text{D})$ diurnal cycles for all campaign data and W and NE air-mass sector data subsets.

cal measurement-model agreement is within the uncertainty of both the model and measurements. In general, the radical measurements are slightly higher than the model during the day-time. However, the model HO_2 constantly over-predicts the measured HO_2 concentrations, by at least a factor of 2–3 (Sommariva et al., 2006a). Inclusion of the halogen chemistry, in terms of measured IO seems to give some small improvement in the agreement. It is worth noting that there is some evidence for spatial inhomogeneity in the [IO] and the “local” [IO] could be a factor or 10 higher than that measured by long-path DOAS methods (see Sommariva et al., 2006a, and Smith et al., 2006). The effect of halogens on the partitioning of NO and NO_2 and OH and HO_2 is dealt with later. Figure 4b shows a correlation plot of modelled PERCA peroxy radical (full + oxy, heterogeneous with IO chemistry, see

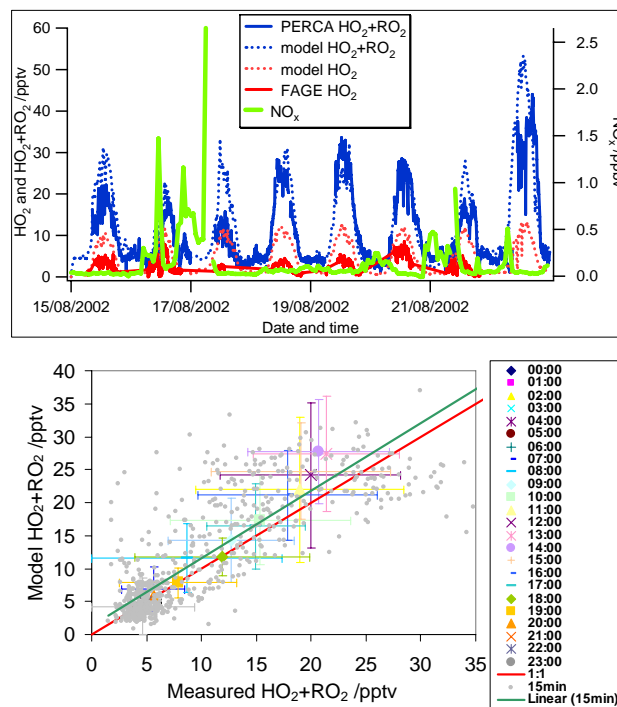


Fig. 4. (a) Measured and modelled PERCA $[\text{HO}_2 + \sum\text{RO}_2]$ and FAGE $[\text{HO}_2]$; (b) MCM modelled vs. measured $[\text{HO}_2 + \sum\text{RO}_2]$ (15 min data) and hourly-averaged data.

Table 4) concentrations against measured values for all fifteen minute-averaged data and also the hourly-averages with their corresponding standard deviation. From Fig. 4b, there is a good correlation with the slope = 1.02 ($R^2 = 0.73$). At low $[\text{HO}_2 + \sum\text{RO}_2]$, the model-measurement comparison lies below the 1:1 line. At higher $[\text{HO}_2 + \sum\text{RO}_2]$, there is a wider spread of data and the model, particularly for the hourly averages around solar noon over-predicts the measured peroxy radical levels.

Correlations of $j(\text{O}^1\text{D})$, $\sqrt{j(\text{O}^1\text{D})}$ vs. $[\text{HO}_2 + \sum\text{RO}_2]$ and diurnal cycles for 23 and 24 August are shown in Fig. 5. The back trajectories shown in Fig. 5 suggest a SW air-mass origin on 23 August coming off the French coast, bringing higher concentrations of CO, CH_4 , acetone, methanol and NO_x . The following day, the air-mass changes to a cleaner NW origin, where the NO/NO_x ratio and [NMHC] are significantly lower. The narrower shape of the diurnal cycle on 24 August could be due to a reduction in species that could be photolysed to form peroxy radicals.

On the NW day (24 August), peroxy radicals track $j(\text{O}^1\text{D})$, whereas on the SW day (23 August), the peroxy radicals have a broader shape than $j(\text{O}^1\text{D})$. The correlation with both $j(\text{O}^1\text{D})$ and $\sqrt{j(\text{O}^1\text{D})}$ ($r^2 = 0.87$ and 0.89 respectively) is good on the NW day but poor ($r^2 = 0.30$ and 0.43) on the SW day. Since the correlation with $j(\text{O}^1\text{D})$ on the SW day is poor, this is suggestive of an increased secondary

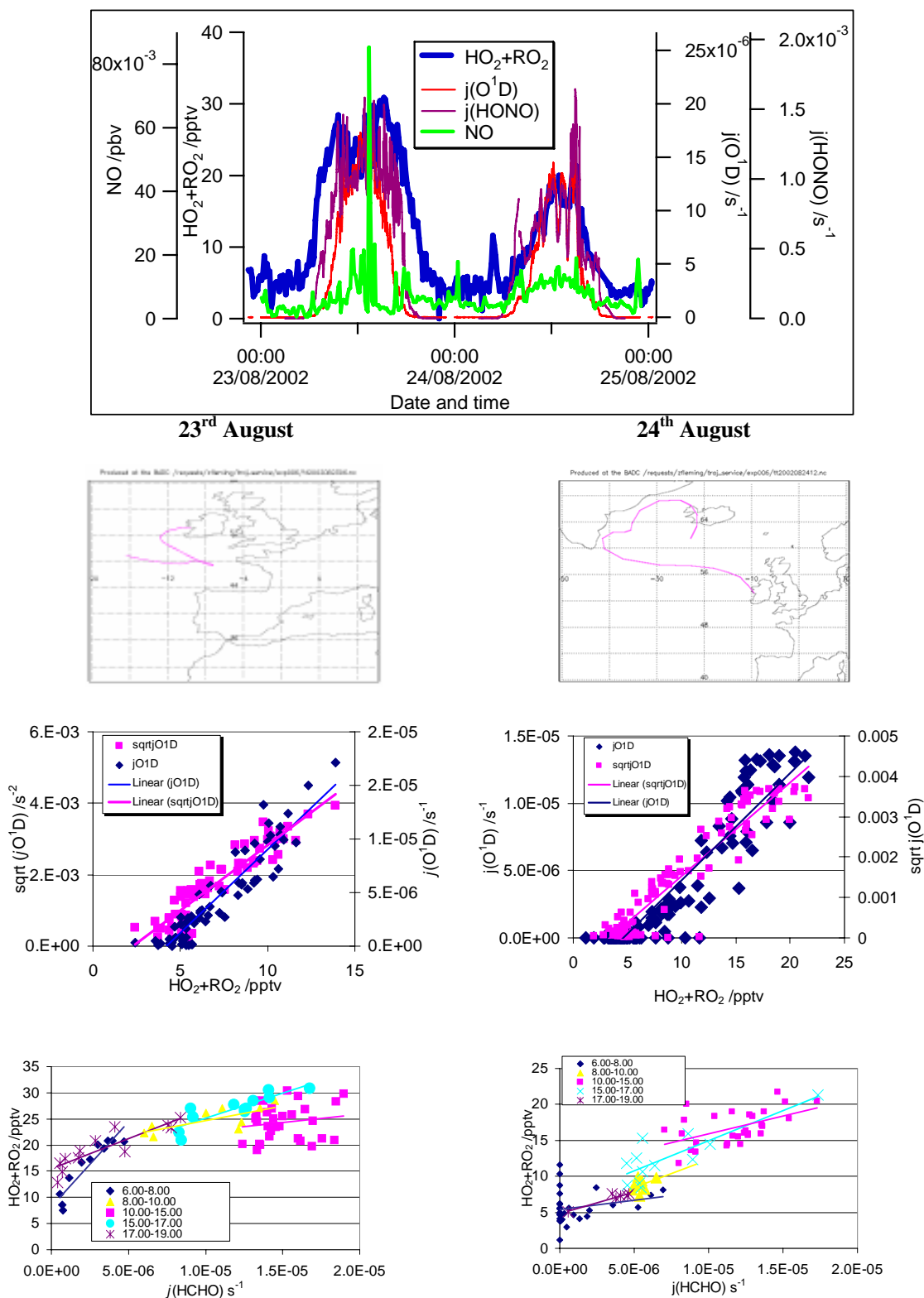


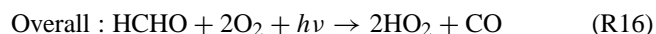
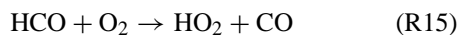
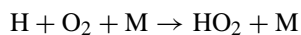
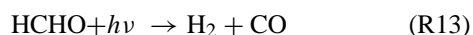
Fig. 5. Data for 23 and 24 August, $\text{HO}_2 + \sum \text{RO}_2$ and $j(\text{O}^1\text{D})$ diurnal cycles and five-day back-trajectories. $j(\text{O}^1\text{D})$ and $\sqrt{j(\text{O}^1\text{D})}$ vs. $[\text{HO}_2 + \sum \text{RO}_2]$ linear regressions and $[\text{HO}_2 + \sum \text{RO}_2]$ vs. $j(\text{HCHO})_{\text{rad}}$ trends at different periods of the day are shown beneath the back-trajectories. The average concentrations/rates on 23 and 24 August where $\text{NO} = 6.3 \pm 8.3, 7.3 \pm 3.9$ pptv, $\text{HO}_2 + \sum \text{RO}_2 = 14.7 \pm 9.4, 8.3 \pm 5.4$ pptv, $j(\text{O}^1\text{D})_{\text{day}} = 7.3 \times 10^{-6} \pm 5.7 \times 10^{-6}, 5.5 \times 10^{-6} \pm 4.8 \times 10^{-6} \text{ s}^{-1}$.

source of HO_x from the photolysis of other compounds. This is particularly apparent in the early morning and the evening.

Photolysis of species other than O_3 (e.g. HCHO , HONO , H_2O_2 , $\text{CH}_3\text{O}_2\text{H}$) could lead to a broadening of the peroxy radical diurnal cycle noted in the afternoon or early morning, when $j(\text{O}^1\text{D})$ (primary production) is reduced. $[\text{HO}_2 + \sum\text{RO}_2]$ vs. $j(\text{HCHO})$ correlations are divided into five different periods of the day in Fig. 5 (23 August, r^2 (6–8) 0.87, (8–10) 0.57, (10–15) 0.03, (15–17) 0.75, (17–19) 0.79; 24th August, r^2 (6–8) 0.05, (8–10) 0.18, (10–15) 0.28, (15–17) 0.70, (17–19) 0.90). On the 23 August, the largest increase in peroxy radicals as $j(\text{HCHO})$ increases was seen at 06:00–08:00 and 15:00–17:00 UT. Despite the large increase in peroxy radicals with increasing $j(\text{HCHO})$ between 06:00 and 08:00, the $j(\text{HCHO})$ values were very low and would not have led to significant peroxy radical formation from this channel alone. Between 17:00–19:00 UT the product ($j(\text{HCHO}) \cdot [\text{HCHO}]$) was significantly large as to produce peroxy radicals at this time when $j(\text{O}^1\text{D})$ was greatly reduced. 24 August showed very poor trends. The same form of analysis with $j(\text{HONO})$ showed no observable trends (N.B. $[\text{HONO}]$ estimated).

Figure 6a shows the campaign averaged $[\text{HCHO}]$ and $[\text{HO}_2 + \sum\text{RO}_2]$ diurnal cycles and the amount of HCHO photolysed each h ($j(\text{HCHO}) \cdot [\text{HCHO}]$). Formaldehyde has a shifted diurnal cycle with values persisting into the evening. Daily maximum $[\text{HCHO}]$ were around 1.6 ppbv (Still et al., 2005), much higher than the 0.2–0.8 ppbv found during a comparable campaign at Mace Head (Cardenas et al., 2000). The relative performance of the different HCHO measurement methods are discussed in Still et al. (2005).

Figure 6b shows correlation plots of $[\text{HO}_2 + \sum\text{RO}_2]$ and $[\text{HCHO}]$ against the amount of formaldehyde photolysed per hour ($j(\text{HCHO}) \cdot [\text{HCHO}]$). Both peroxy radicals and formaldehyde show a good correlation with formaldehyde photolysis. Formaldehyde is both a source of peroxy radicals (through photolysis and OH oxidation) and is produced from the peroxy radical reactions with NO . Figure 6b correlates peroxy radical and formaldehyde concentrations to the HO_2 production rate from formaldehyde photolysis. In the radical channel (14), one HCHO molecule yields two HO_2 molecules upon photolysis



The good correlation is suggestive of persistent peroxy radical levels in the late afternoon having a major contribution from formaldehyde photolysis.

An interesting phenomenon was observed on a number of days (namely 9, 15, 17 and 21 August and 1 September);

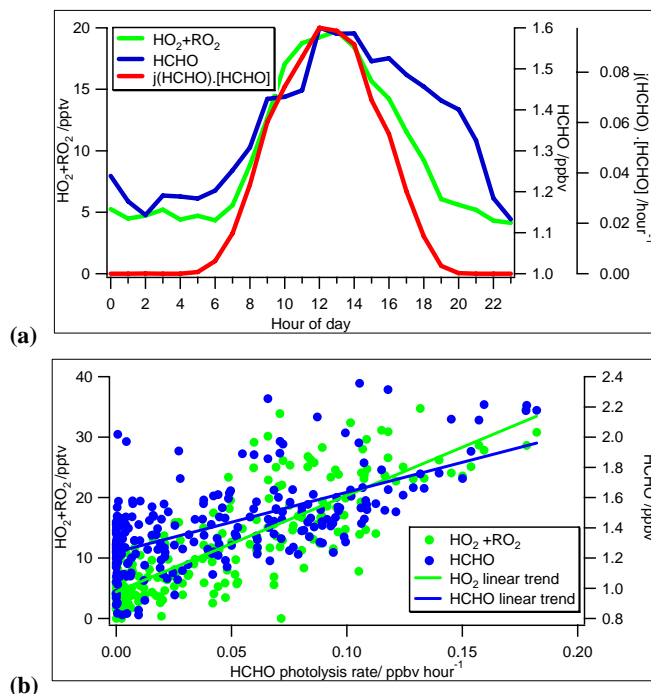


Fig. 6. (a) Hourly-averaged diurnal cycle of $[\text{HCHO}]$, $[\text{HO}_2 + \sum\text{RO}_2]$ and $j(\text{HCHO}) \cdot [\text{HCHO}]$, (b) Trend of ten minute-averaged $[\text{HO}_2 + \sum\text{RO}_2]$ and $[\text{HCHO}]$ vs. $j(\text{HCHO}) \cdot [\text{HCHO}]$ ($r^2_{(\text{HCHO})} = 0.626$, $r^2_{(\text{HO}_2 + \sum\text{RO}_2)} = 0.834$).

when the solar intensity and photolysis rates showed a sudden increase in the early evening and a disproportionately large peroxy radical increase was observed. On 21 August (see case day 21st August in Fig. 7c) a peak in the $j(\text{O}^1\text{D})$ was accompanied by a sudden peroxy radical increase at around 17:00. The same increase in $[\text{OH}]$ was observed by (Smith et al., 2006) on this day, and was reflected in the modelled $[\text{OH}]$ (Sommariva et al., 2006a). The HO_2 measurements did not show a similar increase. One possible explanation for the phenomena is that if there are clouds at a given height as the sun nears the horizon light passes beneath the cloud at high zenith angles, rather than being attenuated by them, giving a short-lived boost to photochemical peroxy radical production (see for example, Monks et al., 2004).

3.2 Peroxy radicals and NO_x

NO_x concentrations of greater than 0.5 ppbv and up to as high as 300 pptv on 8th, 16th and 21st August (see Fig. 2). The trajectories on the 8, 16 and 21 August were classified as W, despite some local polluted SE winds. $[\text{NO}_x]/[\text{NO}_y]$ were higher than average on 13, 14 and 16 August (days where local pollution re-circulated at the site). All the case days (8 August, 16 August and 21 August) in Fig. 7a, b and c had very high $[\text{NO}_x]$ and a peroxy radical diurnal profile that is shifted towards the evening. This may be due to a

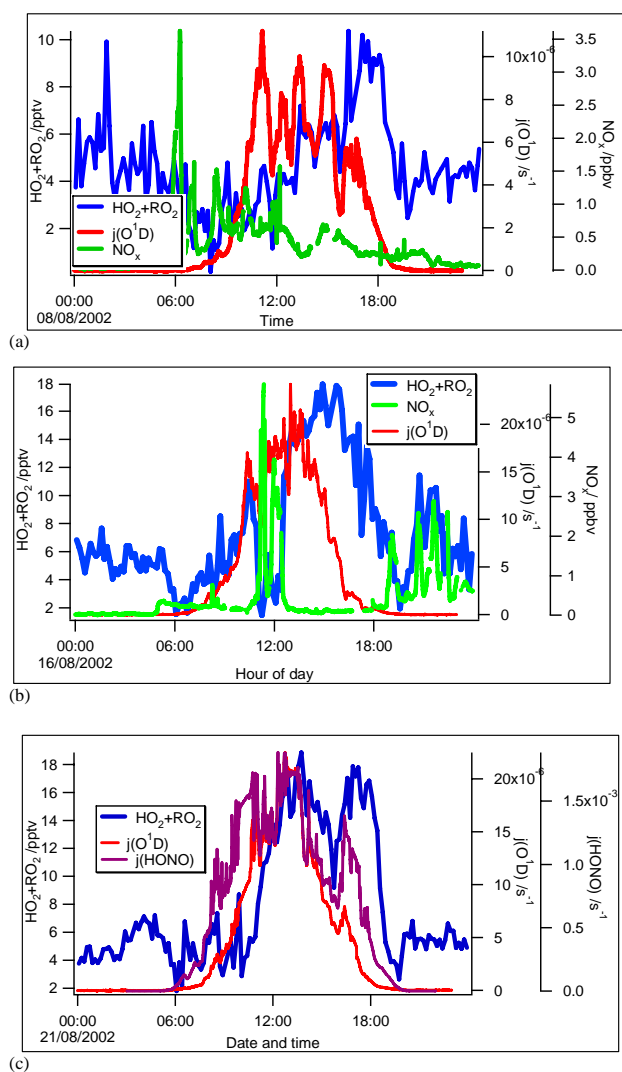


Fig. 7. Case study days: (a) 8 August, (b) 16 August, (c) 21 August $[\text{HO}_2 + \sum\text{RO}_2]$, $j(\text{O}^1\text{D})$ (and $[\text{NO}_x]$ or $j(\text{HONO})$) diurnal cycles.

suppression of peroxy radical formation when NO_x was high, and a sudden period of formation later in the day when NO_x levels dropped.

The campaign average $[\text{NO}_x]$ diurnal cycle is shown in Fig. 3. The values are highest between 09:00 and 12:00. High NO_x levels in the morning can suppress the radical levels via the repartitioning of HO_2 to OH . NO_x suppression may, indeed, contribute to the apparent shift in the maxima of the peroxy radical diurnal cycle. On 16 August in Fig. 7b a sudden NO_x spike in the morning perturbed the peroxy radical concentrations, moving the apparent maximum towards the afternoon.

Mean peroxy radical concentrations vs. binned $[\text{NO}_x]$ (on a logarithmic scale) for all 10-min data are shown in Fig. 8a. The NO_x binning was taken at (equal) log normal intervals. The peroxy radical data were divided into two regimes

according to $j(\text{O}^1\text{D})$ values; $j(\text{O}^1\text{D}) > 7.5 \times 10^{-6} \text{ s}^{-1}$ represents daylight values and $< 3 \times 10^{-7} \text{ s}^{-1}$ that represents night-time conditions, which are discussed later (Fleming et al., 2006).

The data with $j(\text{O}^1\text{D}) > 7.5 \times 10^{-6} \text{ s}^{-1}$ (daylight hours) are used for investigating the effect of NO_x on peroxy radicals during the day. Peroxy radical concentrations decrease with increasing $[\text{NO}_x]$ until values of about 0.1 ppbv $[\text{NO}_x]$. This shift is a result of changes in the $\text{HO}_2:\text{OH}$ ratios towards OH (reactions of HO_2 and RO_2 with NO to form NO_2). Between values of 0.1 and 0.2 ppbv $[\text{NO}_x]$, there is a sudden increase in $[\text{HO}_2 + \sum\text{RO}_2]$, which suggests a switch between NO_x and VOC-limited conditions with respect to ozone production. The corresponding increase in VOCs at $[\text{NO}_x]$ above 0.1 ppbv would lead to a rise in OH oxidation of VOCs, producing more peroxy radicals. Above 0.2 ppbv $[\text{NO}_x]$, increasing $[\text{NO}_x]$ appears to lower $[\text{HO}_2 + \sum\text{RO}_2]$.

Hourly averaged daylight (06:00–19:00) alkane, isoprene, HCHO , CO and CH_4 concentrations, as well as peroxy radical levels, are plotted against binned NO_x in Fig. 8b. The right hand axis is scaled for each hydrocarbon. The sharp increase in all VOCs at $[\text{NO}_x] > 0.1$ ppbv would have a strong link to the rise in peroxy radicals at this time. These high VOC levels change the reactive mixture with respect to peroxy radical speciation. Sudden NO_x increases could reflect changing air-mass composition. Concentrations of the biogenic hydrocarbon, isoprene peak at a lower $[\text{NO}_x]$ than the corresponding anthropogenic hydrocarbons. At higher $[\text{NO}_x]$, it is not clear which hydrocarbons govern the organic peroxy radical concentrations. Until 0.1 ppbv $[\text{NO}_x]$, the peroxy radical trend with increasing $[\text{NO}_x]$ is very similar to the HCHO trend. At $[\text{NO}_x]$ between 0.5 and 1 ppbv the dependence of peroxy radicals on VOCs is very clear, as a drop in all VOCs is reflected in the peroxy radical data.

The rural marine boundary location of Mace Head was seen to be representative of background chemistry but polluted air masses regularly reach the site, bringing higher NO_x levels than experienced in the marine W air-masses as shown in Rickard et al. (2002). The switch to a significant VOC contribution to $[\text{HO}_2 + \sum\text{RO}_2]$ occurs at a lower $[\text{NO}_x]$ level than at more polluted continental urban locations, such as at the BERLIOZ campaign where the maximum $[\text{HO}_2]$ was at 1 ppbv $[\text{NO}_x]$ (Holland et al., 2003).

3.3 Hydrocarbons and $\text{HO}_2/(\text{HO}_2 + \sum\text{RO}_2)$ ratios

The highest mixing ratio of anthropogenic source compounds such as ethene, toluene and benzene was observed between 1 and 5 August. This illustrates the more polluted VOC-laden air masses, originating from the NE sector, passing over Scandinavia, northern Britain and Ireland.

High isoprene concentrations between the 2nd and 4th August were followed immediately by a sharp increase in DMS and this was also seen on 17 and 30 August. DMS concentrations varied from concentrations barely above the detection

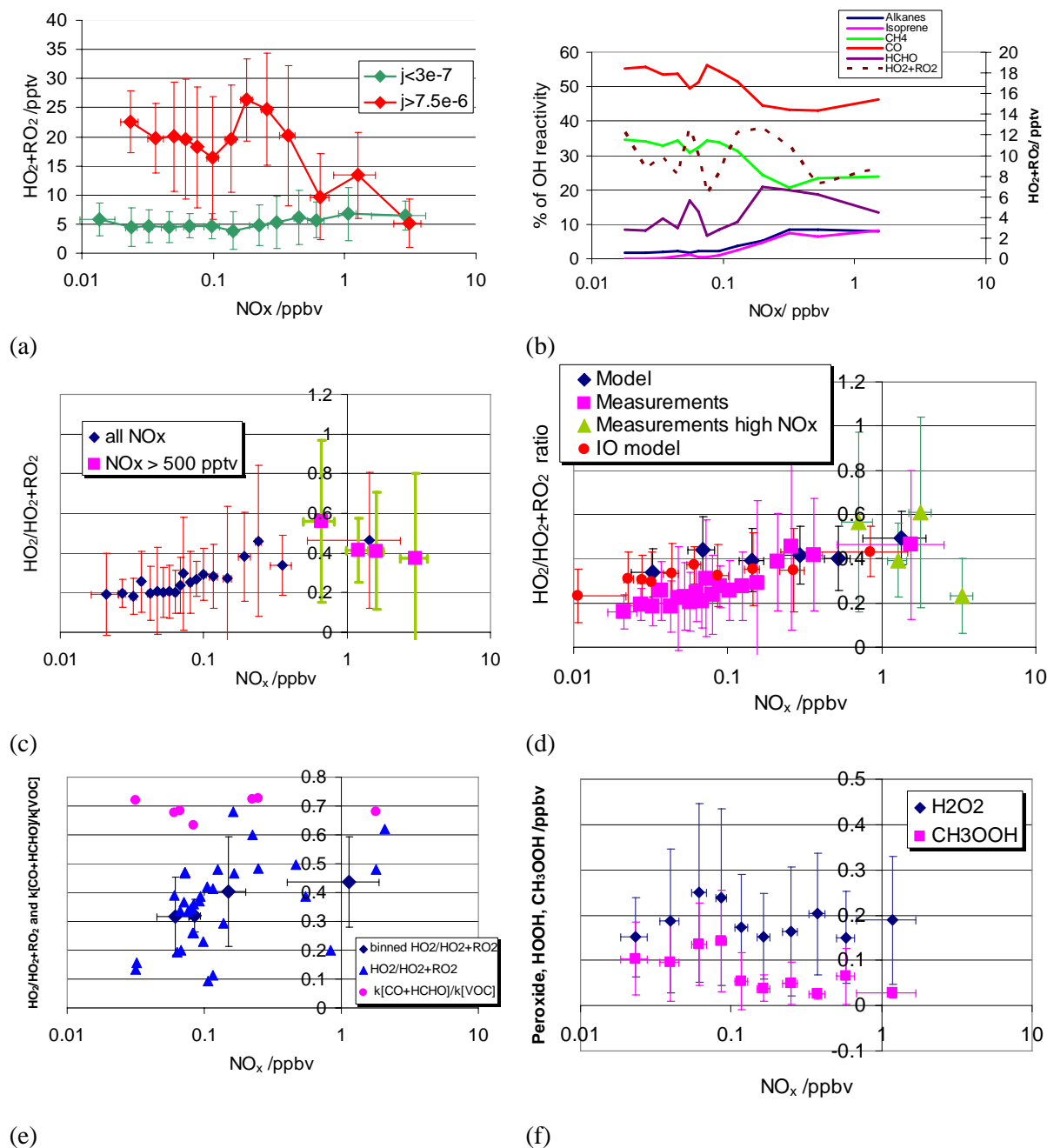


Fig. 8. (a) 10 minute-averaged $[\text{HO}_2 + \sum \text{RO}_2]$ vs. $[\text{NO}_x]$ at two light intensity regimes; (b) Hourly-averaged daylight (06:00–19:00 UT) percentage contribution to OH reactivities for VOCs, CO and $[\text{HO}_2 + \sum \text{RO}_2]$ vs. $[\text{NO}_x]$; (c) $\text{HO}_2 / (\text{HO}_2 + \sum \text{RO}_2)$ vs. $[\text{NO}_x]$: 15 minute-averaged data with values at $[\text{NO}_x] > 500$ pptv in smaller bins; (d) Measured and modelled $\text{HO}_2 / (\text{HO}_2 + \sum \text{RO}_2)$ ratios vs. $[\text{NO}_x]$ for hourly-averages (Model days: 1, 2, 9, 10, 15–22, 31 August and 1 September – model runs both full-oxy with and without halogens (see Table 4)); (e) Hourly-averaged $\text{HO}_2 / (\text{HO}_2 + \sum \text{RO}_2)$ and $\phi(\text{CO} + \text{HCHO})$ vs. $[\text{NO}_x]$; (f) Hourly-averaged [peroxide] vs. $[\text{NO}_x]$; for details of binning procedure, see text.

limit to as high as 900 pptv, with a spike of over 1.5 ppbv on 28 July. DMS levels were highest in the W, N and SW sectors, as shown in Table 1.

OH reacts with hydrocarbons, forming organic radicals, which rapidly react with O_2 to form peroxy radicals (see Reactions 10 and 11). Lewis et al. (2005) calculated the percentage contribution to OH removal by VOCs by com-

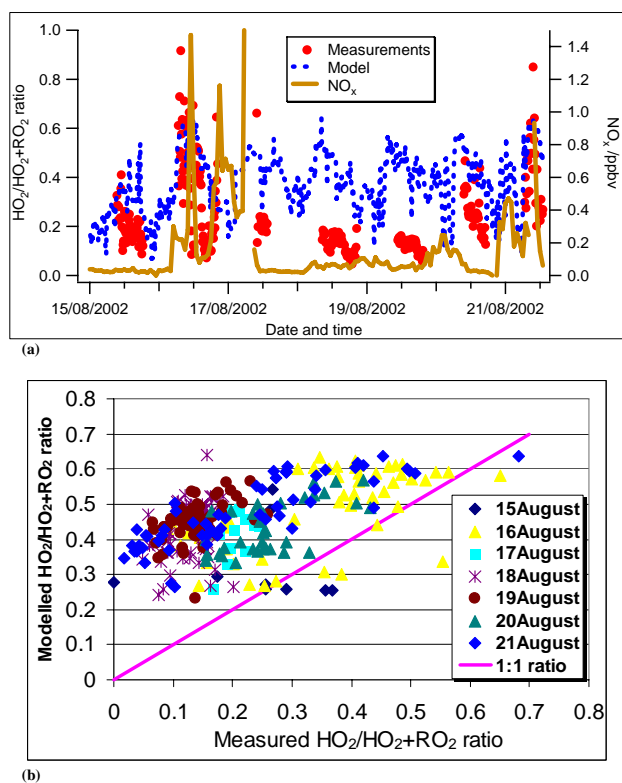
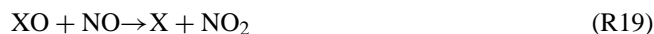


Fig. 9. (a) Hourly-averaged measured and modelled $[\text{HO}_2 + \sum \text{RO}_2]$ diurnal cycles for the 15th–21st August (b) Measured and modelled $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratios 15–21 August.

binning all the VOC-OH reaction rates, $k_{\text{VOC}}[\text{VOC}][\text{OH}]$. Acetaldehyde accounted for up to 20%, CH_4 , and formaldehyde both up to 30% and the other measured non-methane hydrocarbons (NMHCs) between 10 and 15% of OH loss. HO_2 measurements taken by FAGE (Smith et al., 2006) provide a means of comparing HO_2 with $\text{HO}_2 + \sum \text{RO}_2$. Figure 9 shows the measured and modelled (full + oxy, heterogeneous with IO chemistry, see Table 4), hourly-averaged $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratios during the period 15–22 August. The addition of halogens to the system can repartition both OH and HO_2 and NO and NO_2 (e.g. Monks, 2005) via



and



The measured $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratios are lower when $[\text{NO}_x]$ is low as on 18 and 19 August and can reach values over 1 when NO_x -laden air arrives at the site. The $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratio generally decreases from the start of the day towards sunset. The equivalent model ratio shows a similar diurnal profile from day to day, with the distinct

diurnal profile displaying the highest $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratios at midday. Generally, the modelled $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratio is much higher than the measurement equivalent, except when NO_x is high.

The correlation plot of modelled versus measured $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratios is shown in Fig. 9b, with the individual days marked in separate colours. The model generally over-predicts these ratios on all the days. The measured $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratio on 16 August (when local SE winds brought high NO_x levels to the site) showed large variations throughout the day, both for the modelled and measured ratios as seen in both Figs. 9a and b. However, Fig. 9a shows that the measurement ratios displayed greater variability than the modelled ratio during the high NO_x period on this day. The same is observed during the high NO_x period on 21 August, where the model ratio appears not to be influenced by NO_x variations.

To investigate the effect of varying NO_x on the $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratio, a plot of $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratios against binned $[\text{NO}_x]$ is shown in Fig. 8c. The ratio of inorganic to organic peroxy radicals increases as $[\text{NO}_x]$ increases. The highest $[\text{NO}_x]$ bin at 1 ppbv has been divided into smaller bins in order to study the structure at high NO_x . At $[\text{NO}_x] > 0.8$ ppbv, the $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratio appears to decrease with increasing $[\text{NO}_x]$.

Figure 8d shows the measured and modelled $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratios plotted against $[\text{NO}_x]$. Two model runs (with and without IO) are plotted to show the effect of halogens. In general, the $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratio is higher in the model but the model does not show a strong increase with increasing $[\text{NO}_x]$. $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratios at high $[\text{NO}_x]$ for both model and measured values are very similar, but at lower $[\text{NO}_x]$ the model predicts higher $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ ratios. Interestingly, the addition of halogen chemistry improves the agreement between model and measurement, indicating a role for the IO in repartitioning the OH and HO_2 via Reactions (17) and (18). It is clear that at low $[\text{NO}_x]$ the halogens seem to be more important. Sommariva et al. (2006) found that the model mechanism worked better at high NO_x , indicating that peroxy–peroxy reactions at low NO_x are still not fully understood. An earlier Mace Head campaign tailored box model, without OVOC and halogen chemistry, used on Mace Head data, over-predicted $\text{HO}_2/(\text{HO}_2 + \sum \text{RO}_2)$ at low NO_x and under-predicted at high NO_x (Carslaw et al., 1999, 2002). As previously stated, oxidation of CO, CH_4 , HCHO and NMHCs represents a large loss term for OH. The reaction of OH with CO and HCHO leads to the formation of HO_2 . OH reaction with CH_4 forms CH_3O_2 and OH reaction with VOC forms predominantly RO_2 . The fraction of OH removal reactions that form HO_2 can be represented as:

$$\phi(\text{CO} + \text{HCHO}) = \left(\frac{k_{\text{CO}}[\text{CO}] + k_{\text{HCHO}}[\text{HCHO}]}{k_{\text{CO}}[\text{CO}] + k_{\text{HCHO}}[\text{HCHO}] + k_{\text{CH}_4} + k_{\text{VOC}}[\text{VOC}]} \right) \quad (1)$$

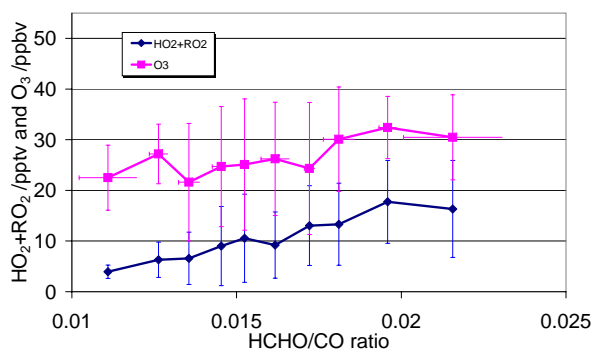


Fig. 10. Hourly-averaged $[\text{HO}_2+\text{RO}_2]$ and $[\text{O}_3]$ vs. $[\text{HCHO}]/[\text{CO}]$.

where k_{CO} , k_{HCHO} , k_{CH_4} and k_{VOC} are the rate coefficients for the reaction of OH with CO, HCHO, CH_4 and VOCs respectively. The rate-coefficients were taken from the National Institute of Standards and Technology (NIST) web site.

The $\phi(\text{CO}+\text{HCHO})$ fraction was calculated for the days that had complete CO, CH_4 and VOC and HCHO concentrations, as in Lewis et al. (2005). Comparing $\phi(\text{CO}+\text{HCHO})$ ratios with $\text{HO}_2/(\text{HO}_2+\sum\text{RO}_2)$ ratios should be indicative of whether $\text{HO}_2/(\text{HO}_2+\sum\text{RO}_2)$ ratio variations were caused primarily by varying HCHO, CO, CH_4 and VOC concentrations.

Figure 8e shows a plot of hourly $\text{HO}_2/(\text{HO}_2+\sum\text{RO}_2)$ and $\phi(\text{CO}+\text{HCHO})$ ratios against binned $[\text{NO}_x]$. The trend for increasing $\text{HO}_2/(\text{HO}_2+\sum\text{RO}_2)$ with increasing $[\text{NO}_x]$ is not replicated for $\phi(\text{CO}+\text{HCHO})$, which does not appear effected by NO_x . $\text{HO}_2/(\text{HO}_2+\sum\text{RO}_2)$ ratios are always lower than $\phi(\text{CO}+\text{HCHO})$ ratios (<0.5 for $\text{HO}_2/(\text{HO}_2+\text{RO}_2)$ and >0.5 for $\phi(\text{CO}+\text{HCHO})$). The ratio of $k[\text{HCHO}]/(k[\text{HCHO}]+k[\text{CO}])$ was found to remain constant at around 0.5, showing that HCHO and CO contribute equally to HO_2 formation. $\phi(\text{CO}+\text{HCHO})$ ratios have a range of between 0.3 and 2.5 in NAMBLEX, much greater than the $\text{HO}_2/(\text{HO}_2+\sum\text{RO}_2)$ ratio range. The $\phi(\text{CO}+\text{HCHO})$ ratios were usually much higher than $\text{HO}_2/(\text{HO}_2+\sum\text{RO}_2)$ ratios, which suggests that modelling the VOC-OH reactivity underestimates the resulting RO_2 concentrations with respect to HO_2 . Also, calculating HO_2 to be directly correlated with CO and HCHO reactivity is not necessarily valid as HCHO is both photolysed and is oxidised by OH to form HO_2 . HCHO is also formed from the reaction of CH_3O_2 with NO (Reactions 2 and 3).

The HCHO:CO ratio can be used as a tracer to distinguish different air masses and differing times since the last major input from pollution. It is of interest because both tracers are primary pollutants, but formaldehyde is also produced in the troposphere by oxidation of CH_4 in the presence of NO_x . Subsequent photolysis of this formaldehyde then produces CO. In polluted high NO_x environments, HCHO production is more important than its pho-

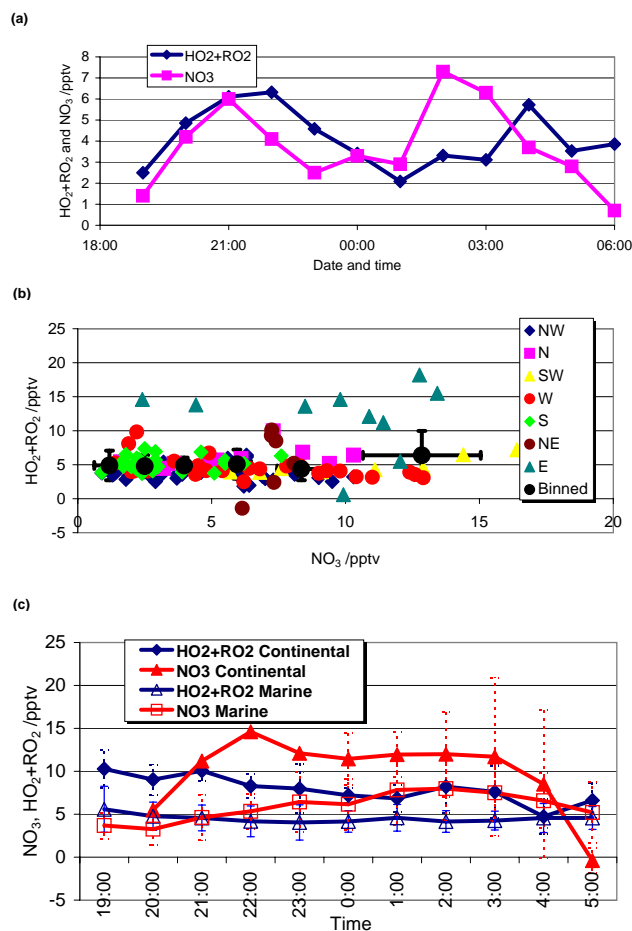


Fig. 11. (a) Typical nighttime profile of $[\text{HO}_2+\sum\text{RO}_2]$ and $[\text{NO}_3]$ for the night of 28/29 August; (b) Hourly-averaged $[\text{HO}_2+\sum\text{RO}_2]$ vs. $[\text{NO}_3]$ for air-mass sectors. Large black markers represent all campaign data divided into six bins according to NO_3 with the corresponding (1σ) standard deviations; (c) $[\text{NO}_3]$ and $[\text{HO}_2+\sum\text{RO}_2]$ at night (19:00 to 06:00) with the (1σ) standard deviation within each hour.

tolysis and the HCHO:CO ratio increases. Figure 10 shows that both $[\text{HO}_2+\sum\text{RO}_2]$ and $[\text{O}_3]$ increase with increasing HCHO:CO. If higher HCHO:CO ratios are a marker for polluted conditions, then this would be likely to lead to higher ozone levels. Higher peroxy radical levels at increased HCHO suggests that HCHO is more effective at producing peroxy radicals than CO. The $\phi(\text{CO}+\text{HCHO})$ ratio presumes that HCHO and CO have equal HO_2 productivity, so any discrepancy between $\phi(\text{CO}+\text{HCHO})$ and $\text{HO}_2/(\text{HO}_2+\sum\text{RO}_2)$ may be due to the inaccuracy of predicting HO_2 from $\phi(\text{CO}+\text{HCHO})$.

3.4 Peroxides

The highest H_2O_2 concentrations of up to 0.5 pptv were between 1 and 3 August when NO_x and hydrocarbon concen-

trations were high. Indeed, in Table 1, high NO_x and high VOC concentrations in the E sector have led to the highest peroxide concentrations. Morgan (2004) found a maximum $[\text{H}_2\text{O}_2]$ of 1.1 ppbv with an average of 0.19 ppbv during NAMBLEX, much lower than the maximum of 7.1 and mean of 1.58 ppbv at Mace Head in June 1999 (Morgan and Jackson, 2002). In the clean marine boundary layer, such as Cape Grim, Tasmania, peroxy radicals are more likely to self-react to form peroxides than they are to react with NO and subsequently produce ozone (Ayers et al., 1997). Thus, high levels of peroxides would signify an ozone-destroying regime and a lower turnover rate of the various species that are part of the ozone-forming cycles.

Figure 8f shows [peroxide] versus $[\text{NO}_x]$ trends that are very similar in shape to peroxy radical – NO_x trends in Fig. 8a, illustrating the strong link between peroxy radicals and H_2O_2 concentrations. H_2O_2 concentrations are highest at around 0.1 ppbv NO_x and decrease slightly at higher NO_x but do not decrease to the same extent as peroxy radical levels at high NO_x as shown in in Fig. 8a. $\text{CH}_3\text{O}_2\text{H}$ is more influenced by NO_x than H_2O_2 , as it is reduced to nearly zero values at high NO_x .

3.5 Night-time chemistry

Table 3 shows night-time averages for the different air mass sectors and the concomitant NO_3 measurements from data on fifteen nights (Saiz-Lopez et al., 2006; see also Bitter et al., 2005). The E and NE air-mass sectors have the highest average O_3 , NO_3 (with the E sector having $[\text{NO}_3]$ of 11.7 pptv, compared to less than 6 pptv in all the other sectors) and total alkene concentrations as seen in Table 3. The highest night-time peroxy radical concentrations are observed in the SE and E sectors (cf. Allan et al., 2000).

Sommariva et al. (2006b) found that the model had a tendency to underestimate night time peroxy radical levels except on 31 August and 1 September. Closer agreement between the model and measurements was achieved when moving from a “clean” model with only CO and CH_4 to the full model with more complex hydrocarbons. Short-term NO_x spikes during the night are often matched with elevated peroxy radical concentrations as high as 10 pptv, or even 25 pptv in the polluted E period on 2 and 3 August. On 16 August, when $[\text{NO}_x]$ suddenly increased at about 20.00, a significant rise in $[\text{HO}_2 + \sum\text{RO}_2]$ followed closely, as shown in the case day in Fig. 7b. The same is not observed for the 8th d August (Fig. 7a). $[\text{HO}_2 + \sum\text{RO}_2]$ vs. $[\text{NO}_x]$ for nights with NO_3 data are shown in Fig. 8a. As $[\text{NO}_x]$ increases, $[\text{HO}_2 + \text{RO}_2]$ increases at $[\text{NO}_x] > 0.1$ ppbv. This peroxy radical increase with NO_x is suggestive of NO_3 radicals (formed by $\text{NO}_2 + \text{O}_3$) reacting with hydrocarbons to form peroxy radicals.

Figure 11a shows NO_3 and $\text{HO}_2 + \sum\text{RO}_2$ from a typical overnight period. Figure 11b is a plot of average $[\text{HO}_2 + \sum\text{RO}_2]$ vs. $[\text{NO}_3]$ (Saiz-Lopez et al., 2006) for all the air-mass sectors. The $[\text{HO}_2 + \sum\text{RO}_2]$ were separated into six

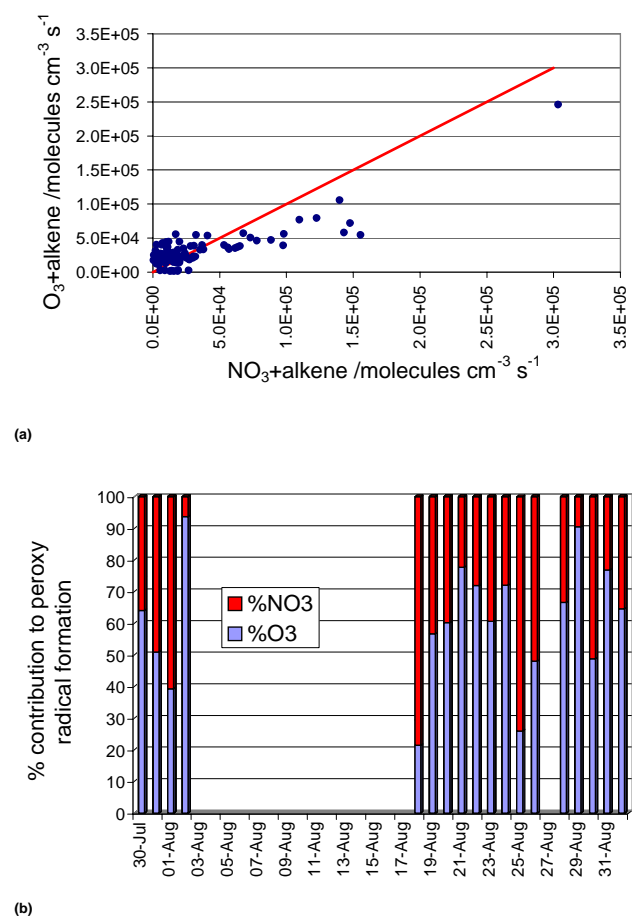


Fig. 12. (a) Flux of peroxy radicals formed from alkene- NO_3 and alkene- O_3 channels: Hourly-averaged night-time (19:00 to 06:00 UT) fluxes. The red-line is the 1:1 ratio; (b) Percentage contribution of O_3 and NO_3 to peroxy radical formation from alkene night-time reactions.

$[\text{NO}_3]$ bins and plotted on the same graph with error bars showing their standard deviation. The E sector, even though $[\text{NO}_3]$ varies widely, always has higher $[\text{HO}_2 + \sum\text{RO}_2]$ than the other sectors, irrespective of $[\text{NO}_3]$. Figure 11c shows the night-time profiles of $[\text{NO}_3]$ and $[\text{HO}_2 + \sum\text{RO}_2]$ for the entire marine (N, NW, SW and W combined) and continental (NE, E combined) air-mass sectors. The NO_3 concentration was always higher in the continental sector. The peroxy radical concentration was also always higher in the continental sector. There does not appear to be a significant peroxy radical pattern throughout the night.

Rate constants for the reaction of NO_3 with the measured alkenes were used to calculate the rate of the NO_3 loss and the O_3 reactions with alkenes (cf. Salisbury et al., 2001). The flux of peroxy radicals formed from the alkene reactions from the NO_3 and O_3 channels were compared by correlating all night-time hours of the campaign as shown in Fig. 12a. At low peroxy radical-forming fluxes, the ozone-alkene re-

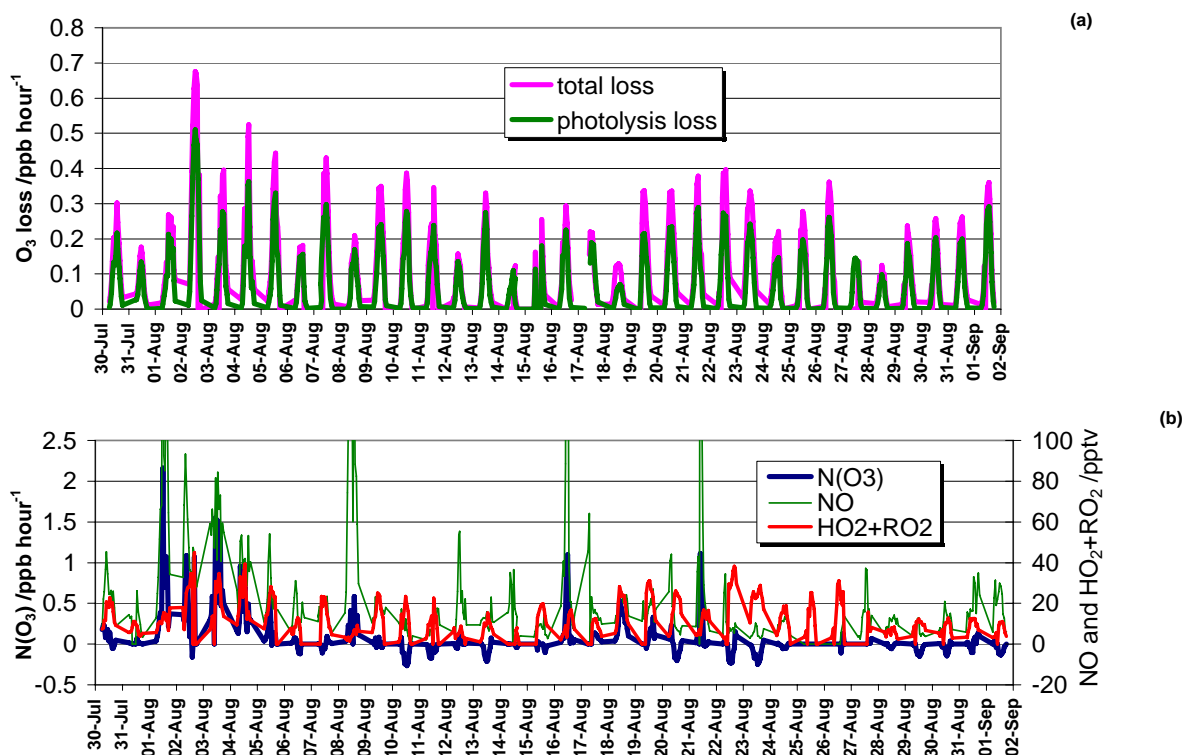


Fig. 13. (a) Total ozone loss and ozone loss via photolysis. (b) Net ozone production and corresponding $[NO]$ and $[HO_2 + \sum RO_2]$ used to calculate $P(O_3)$.

actions tended to dominate over the NO_3 -alkene reactions. When NO_3 levels were high, the fluxes from NO_3 -alkene reactions were far higher than the ozone-alkene fluxes. At NO_3 -alkene fluxes above 5×10^4 molecules $cm^{-3} s^{-1}$, the ozone-alkene flux was always lower than the NO_3 -alkene flux. Figure 12b shows the percentage contribution to peroxy radical formation from alkene night-time reactions. This varies strongly from night to night, with high NO_3 contributions on 18 and 25 August, receiving W and SW air-masses respectively. For the nights for which full data is available the overall contribution of ozone-alkene chemistry to peroxy radical production was 59% compared to 41% for NO_3 -alkene.

Peroxy radical levels were seen to decrease throughout the night in EASE 97 (Salisbury et al., 2001), with more polluted conditions experiencing less of a decrease throughout the night. Analysis to determine the percentage contribution of the ozone – alkene and NO_3 reactions to form peroxy radicals was carried out for EASE 97 (Salisbury et al., 2001). The contribution of both was found to vary between 30 and 70%, for each wind sector, but on the whole as with this study the ozone-alkene reaction was the dominant production mechanism. Carslaw et al. (1997) found a positive correlation between $HO_2 + \sum RO_2$ and NO_3 at the Weybourne Atmospheric Observatory (see also Geyer et al., 2003), while Mihelcic et al. (1993) found a negative correlation between per-

oxy radicals and NO_3 (presumably owing to highly variable reactive hydrocarbon fluxes) at Schauinsland. Any lack of correlation is not surprising, as NO_3 is both a source (Wayne et al., 1991) and a sink (Biggs et al., 1994) of peroxy radicals.

3.6 Photochemical production of ozone

Net photochemical ozone formation, $N(O_3)$ (or ozone tendency) was calculated for each hour of the campaign between 06:00 and 19:00, using Eq. (3) (For assumptions inherent in this form of calculation see Salisbury et al., 2002). The production term represents NO_2 formation and subsequent photolysis to form ozone (Reactions 1 to 4). k_p is the combined rate coefficient for the oxidation of NO to NO_2 by all peroxy radicals (Reactions 1 and 2). The loss term represents the reaction of ozone with OH and HO_2 and ozone photolysis (where f represents the fraction of $O(^1D)$ that reacts with H_2O to form OH).

$$N(O_3) = P(O_3) - L(O_3) \quad (2)$$

$$N(O_3) = k_p [NO] [HO_2 + \sum RO_2] - \{f \cdot j(O^1D) + k_{21}[OH] + k_6[HO_2]\} [O_3] \quad (3)$$



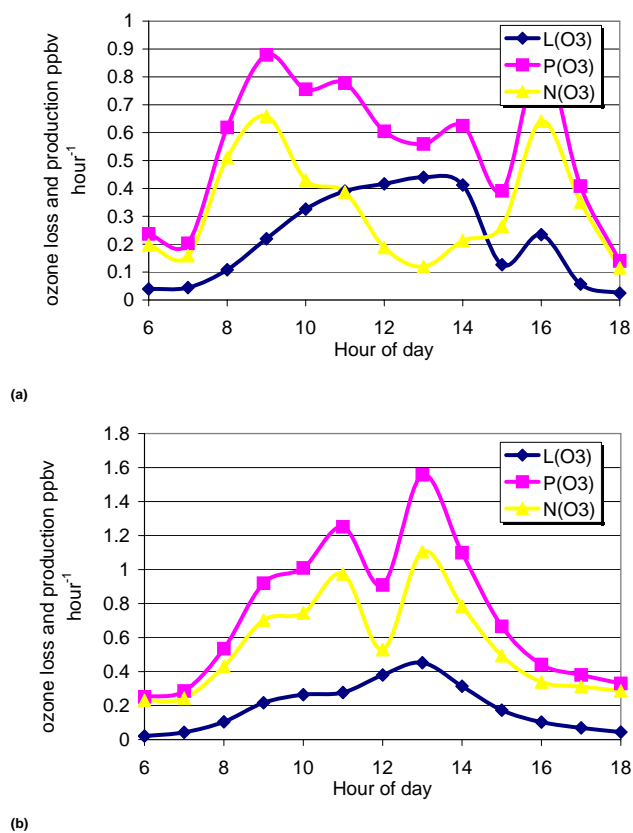


Fig. 14. Hourly-averaged loss, production and net ozone production for (a) Clean conditions (full days with SW/W wind direction) and (b) Polluted conditions (full days with NE/E wind direction).

Figure 13a shows a time series of calculated ozone loss for all campaign daylight hours. The largest contribution to the calculated loss is that of ozone photolysis. The average ozone loss chemistry was calculated to be 64% from ozone photolysis, 8% from the OH+O₃ reaction and 24% from the HO₂+O₃ reaction. The contributions from the three loss reactions vary from day to day, with total ozone loss varying between 0.1 and 0.7 ppbv h⁻¹ at the solar maximum.

Figure 13b is a plot of net ozone production, N(O₃) throughout the campaign with [HO₂ + ΣRO₂] and [NO] plotted on the right-hand axis. The ozone production term, P(O₃) is dependent on [NO] and [HO₂ + ΣRO₂], the ratio of which varies greatly from day to day, showing an inverse relationship during the daylight hours. [NO] was more variable than [HO₂ + ΣRO₂] during NAMBLEX. Lower [NO] leads to a smaller P(O₃) term, which means that ozone loss becomes nearly as great as ozone production, leading to a few hours and days where N(O₃) was negative.

Table 2 shows the average mixing ratios for a range of tracers and N(O₃) for daylight (06:00–19:00 UT) hours with differing air mass sectors. Figure 14 shows the hourly-averaged ozone loss and production rates for NAMBLEX. The loss term follows $j(\text{O}^1\text{D})$, peaking at solar noon, and does not

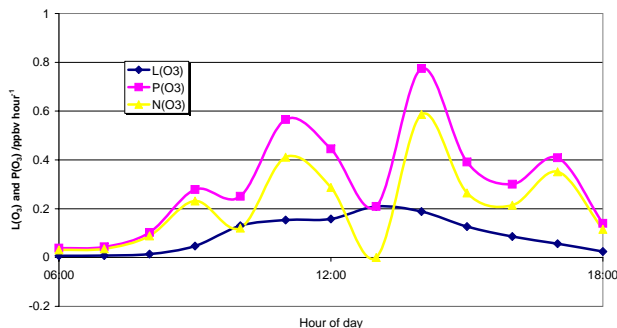


Fig. 15. Case day 8 August: Ozone production, loss and net production.

vary widely from day to day. However, ozone production values show a high degree of variation between days, with midday values varying from 0.1 to 2.5 ppbv h⁻¹. The shift of the maximum ozone loss term towards the afternoon results in the net ozone production being lower in the afternoon than the morning. The rise in P(O₃) in the late afternoon caused by high peroxy radical levels leads to an increase in net ozone production at 16:00. This averaged diurnal cycle appears to show overall ozone production but the high P(O₃) during the polluted E air-mass sector period of 1–5 August shifts the balance to positive N(O₃), despite the many periods of net ozone destruction.

Figure 15 shows N(O₃) for 8 August. This was a day where high [NO_x] reduced peroxy radical levels and the elevated [NO] led to higher net ozone production than on the days preceding and following it. A high NO_x episode in the morning delayed peroxy radical production until around 14:00 (Fig. 7a) and produced high P(O₃). At 13:00 P(O₃) was low because [NO_x] dropped away, and the peroxy radical levels had not yet recovered. The build-up of peroxy radical levels in the afternoon led to a boost in P(O₃) and another boost between 17:00 and 18:00 when night-time peroxy radical-forming reactions become important.

Figure 16 shows N(O₃) plotted against [NO_x]. N(O₃) rises sharply with increasing [NO_x] until around 1 ppbv [NO_x], when the increase in N(O₃) levels off. The increase in N(O₃), with increasing NO_x, during the SOAPEX 2 campaign at Cape Grim, Tasmania showed very similar characteristics until [NO_x] of 0.5 ppbv. Investigations of the effect of NO_x on N(O₃) at Mace Head during the spring campaign of EASE 97 showed a much steeper increase in N(O₃) at similar NO_x levels to those seen during NAMBLEX. Table 6 shows the ozone production values for Mace Head – EASE 97 (Salisbury et al., 2002), Cape Grim – SOAPEX 2 (Monks et al., 2005) and NAMBLEX, demonstrating the much higher range during the spring EASE 97 campaign.

Mace Head has experienced a positive trend in background ozone of 0.49±0.19 ppbv year⁻¹ since 1987 (Simmonds et al., 2004), the largest trend being during the winter season. The behaviour of this trend may be attributed to the sensitivity of

Table 5. Sensitivity of derived $P(O_3)$ and $L(O_3)$ to NO from a series of marine boundary layer campaigns.

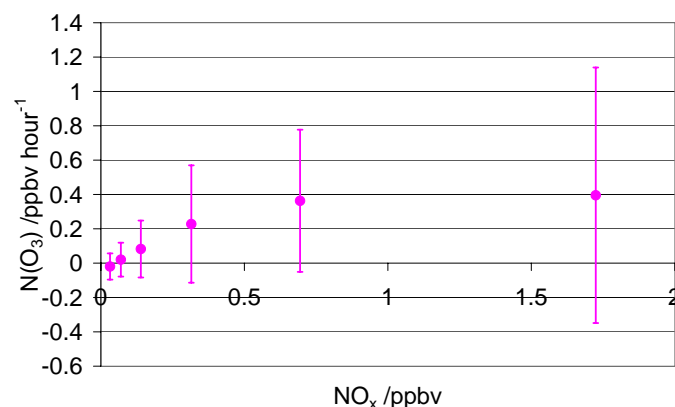
	$d\ln(P(O_3))/d\ln(NO)$	$d\ln(L(O_3))/d\ln(NO)$
Cape Grim(SOAPEX 2 ^a)	0.90 ± 0.03	0.06 ± 0.05
Mace Head (EASE 97 ^b)	1.10 ± 0.05	0.27 ± 0.05
Weybourne (winter ^c)	0.92 ± 0.12	-0.02 ± 0.14
Weybourne (summer ^c)	0.95 ± 0.02	0.28 ± 0.05
Mace Head (NAMBLEX ^d)	1.0 ± 0.04	0.32 ± 0.07

The uncertainties are given as 1σ standard deviations; ^a) data from Monks et al. (2006); ^b) data from Salisbury et al. (2002); ^c) data from Fleming et al. (2006); ^d) this work.

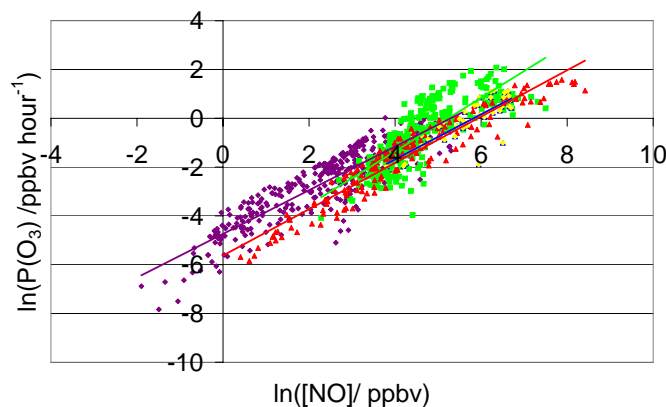
Table 6. Daily $N(O_3)$ campaign average in EASE 97, SOAPEX 2 and NAMBLEX (all in $ppbv\ h^{-1}$).

Campaign and Season	Season	Mean
Mace Head (EASE 96 ^a)	Summer	0.3
Mace Head (EASE 97 ^a)	Spring	1.0
Cape Grim (SOAPEX 2 ^b)	Summer	-0.01
N. Pacific (PHOBEA ^c)	Spring	-0.1
E. Pacific (ORION99 ^d)	Summer	0.2–3.4
Mace Head (NAMBLEX ^e)	Summer	0.11

^a) see Salisbury et al. (2002), ^b) see Monks et al. (2006), ^c) Kotchenruther et al. (2001), ^d) Kanaya et al. (2002), ^e) this work.

**Fig. 16.** Net ozone production for daylight hours binned according to $[NO_x]$.

the background ozone level to changing European emissions of NO_x and VOC (Derwent et al., 2003; Monks, 2003). Following the methodology of Stroud et al. (2004) the sensitivity of $P(O_3)$ to NO was calculated as $d\ln P(O_3)/d\ln(NO)$, as shown in Fig. 17, for a series of marine boundary layer campaigns with differing continental influences. Table 5 summarises the derived sensitivity values of the ozone production term to NO. Both the Mace Head data sets have ozone production with linear sensitivity (i.e. $d\ln P(O_3)/d\ln(NO)=1$)

**Fig. 17.** $\ln(P(O_3))$ vs. $\ln(NO)$ for Weybourne summer (red) and winter (blue), Mace Head; NAMBLEX (yellow), EASE 97 (green) and Cape Grim; SOAPEX 2 (purple). See Table 5 for campaign details.

to NO as compared to Cape Grim and Weybourne, that have values of around 0.9. The Mace Head values imply that the ozone production rate is strongly dependent on the $[NO]$. The equivalent derived values of $d\ln L(O_3)/d\ln(NO)$ are also given in Table 5 the bulk of these values range from ca. 0 to 0.3, unsurprisingly this suggests that $L(O_3)$ is generally independent of small changes in $[NO]$. In tandem, these results imply that the $N(O_3)$ will be strongly sensitive in the marine boundary layer to small changes in $[NO]$.

4 Conclusions

During NAMBLEX, the Mace Head Atmospheric Research Station received a substantial mix of air-masses from both the Atlantic and from Britain and Ireland. 80% of the air-masses were from the clean N, NW, W and SW sectors. The marine air-mass sectors had peroxy radical levels below 10 pptv, whereas the other sectors experienced levels above 13 pptv. The higher peroxy radical concentrations in the air-mass sectors with a continental influence were accom-

panied by over twice as high NO_x levels and much higher anthropogenic hydrocarbon mixing ratios.

Peroxy radical diurnal cycle maxima were typically shifted towards the afternoon, with daily maximum levels between 10 and 40 pptv. MCM modelling of peroxy radical levels provided a good model-measurement comparison, with occasional slight over-estimations by the box model.

Correlations of peroxy radicals with $j(\text{O}^1\text{D})$ were often disturbed by NO_x episodes that temporarily lowered peroxy radical levels. No significant reliable linear or square root dependence with $j(\text{O}^1\text{D})$ was observed to make a clear separation between clean and polluted conditions. Photolysis of compounds other than ozone led to a broader peroxy radical diurnal cycle than would be seen from production via ozone photolysis alone, especially in continentally-influenced air-masses. Correlations with $j(\text{HCHO})$ in the afternoon and a definite shift in the HCHO diurnal cycle towards the afternoon suggests the high potential for HCHO photolysis at this time. A sudden increase in photolysis rates (i.e. a rise in $j(\text{O}^1\text{D})$) in the early evening was seen to cause a large pulse in peroxy radicals.

Short-term large NO_x mixing ratios, termed “ NO_x spikes”, reaching the site caused a marked drop in peroxy radical levels. Plotting the overall effect of NO_x on peroxy radical levels during the whole campaign demonstrated a decrease in peroxy radicals with increasing NO_x . A break in the linear decrease around 0.1 and 0.2 ppbv NO_x was accompanied by an increase in VOCs which led to a short period where peroxy radicals actually increased with NO_x . This VOC influence on peroxy radicals could be thought of as the break between NO_x - and VOC-limited ozone producing regimes.

Comparisons with FAGE HO_2 measurements have shown that $\text{HO}_2/(\text{HO}_2+\sum\text{RO}_2)$ ratios are dependent on $[\text{NO}_x]$ and ranged between 0.2 and 0.6. $\text{HO}_2/(\text{HO}_2+\sum\text{RO}_2)$ ratios increase remarkably linearly with increasing NO_x . The MCM model did not replicate this NO_x -dependence with the model in general over-estimating $\text{HO}_2/(\text{HO}_2+\sum\text{RO}_2)$ ratios. The addition of halogen oxide chemistry improved the level of agreement. Calculating the expected $\text{HO}_2/(\text{HO}_2+\sum\text{RO}_2)$ ratios from OH oxidation reactions of VOC, HCHO and CO did not show any significant NO_x -dependence.

Night-time peroxy radical concentrations rose to as high as 25 pptv in continental air-masses with high NO_x . Sudden NO_3 spikes definitely caused an increase in peroxy radicals, but constant higher NO_3 levels did not necessarily lead to higher peroxy radical concentrations. Peroxy radical and NO_3 mixing ratios were higher in continental compared to marine air-masses. The contribution of ozone-alkene and NO_3 -alkene reactions to peroxy radical formation varies from night to night and there are variations as to which one predominates. At low NO_3 , ozone-alkene reactions are always predominant

Net photochemical ozone production reached as high as 1.5 ppbv h^{-1} with the lowest values being negative at -0.5 ppbv h^{-1} . Highest net ozone production was observed

during high NO_x periods, demonstrating a clear increase in ozone production at higher NO_x . The ozone production rate is strongly dependent on $[\text{NO}]$ having a linear sensitivity ($\text{dln}(\text{P}(\text{O}_3))/\text{dln}(\text{NO})=1.0$). The results imply that the $\text{N}(\text{O}_3)$ will be strongly sensitive in the marine boundary layer to small changes in $[\text{NO}]$ which has ramifications for changing NO_x loadings in the European continental boundary layer.

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