

# **Bacterial Metabolites and Particle Size Determine Cerium Oxide Nanomaterial Biotransformation**

Blanche Collin, Mélanie F Auffan, Emmanuel Doelsch, Olivier Proux, Isabelle Kieffer, Philippe Ortet, Catherine Santaella

# **To cite this version:**

Blanche Collin, Mélanie F Auffan, Emmanuel Doelsch, Olivier Proux, Isabelle Kieffer, et al.. Bacterial Metabolites and Particle Size Determine Cerium Oxide Nanomaterial Biotransformation. Environmental Science and Technology, 2022, 56 (23), pp.16838-16847.  $10.1021/acs.est.2c05280$ . cirad-03938129

# **HAL Id: cirad-03938129 <https://hal.science/cirad-03938129>**

Submitted on 13 Apr 2023

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

This document is confidential and is proprietary to the American Chemical Society and its authors. Do not copy or disclose without written permission. If you have received this item in error, notify the sender and delete all copies.

### **Bacterial metabolites and particle size determine cerium oxide nanomaterial biotransformation**



**SCHOLARONE™** Manuscripts



 BM30/CRG-FAME, ESRF, Université Grenoble Alpes, CNRS, IRSTEa, Météo France, IRD, OSUG, Grenoble 38000, France Philippe Ortet Aix Marseille Univ, CEA, CNRS, BIAM, LEMIRE, Laboratory of Microbial Ecology of the

- Rhizosphere, ECCOREV FR 3098, F-13108, St-Paul-lez-Durance, France
- 

 **Abstract** 

 Soil is a major receptor of manufactured nanomaterials (NMs) following unintentional releases or intentional uses. Ceria NMs have been shown to undergo biotransformation in plant and soil organisms with a partial Ce(IV) reduction into Ce(III) but the influence of environmentally widespread soil bacteria is poorly understood. We used high-energy resolution fluorescence detected X-ray absorption spectroscopy (HERFD-XAS) with an unprecedented detection limit to assess Ce speciation in a model soil bacterium 58 (*Pseudomonas brassicacearum*) exposed to CeO<sub>2</sub> NMs of different sizes and shapes. The 59 findings revealed that the  $CeO<sub>2</sub>$  NMs size drives the biotransformation process. No 60 biotransformation was observed for the 31 nm CeO<sub>2</sub> NMs, contrary to 7 nm and 4 nm CeO<sub>2</sub> 61 NMs, with a Ce reduction of  $64 \pm 14\%$  and  $70 \pm 15\%$ , respectively. This major reduction appeared quickly, from the early exponential bacterial growth phase. Environmentally relevant organic acid metabolites secreted by *Pseudomonas*, especially in the rhizosphere, were investigated. The 2-keto-gluconic and citric acid metabolites alone were able to 65 induce a significant reduction in 4 nm CeO<sub>2</sub> NMs. The high biotransformation measured for < 7 nm NMs would affect the fate of Ce in the soil and biota. 

## **Synopsis**

 Soil bacteria are widespread in the environment but little is known on their influence on the fate of nanomaterials. The NM size, the bacterial interaction and bacterial metabolites drive NM bioreduction.

# **Introduction**

 The biological transformation of nanomaterials (NMs) is a key to understanding the 75 impact and fate of NMs in the environment.<sup>1, 2</sup> Soil is a major environmental collector for 76 NMs, along with sediment and landfill.<sup>3, 4</sup> Soil is also a complex ecological system hosting a diverse range of living organisms and millions of microorganisms that can induce environmental NM transformation at the scale of the organisms, tissues, cells and biomolecules. Plants and microorganisms secrete secondary metabolites, i.e. specialized molecules, that are essential for their biological activities while modulating their interactions with the environment and neighboring organisms. Despite soil biodiversity, NM biotransformation in terrestrial ecosystems has primarily focused on plant-83 nanomaterial interfaces and several studies have assessed the fate of  $CeO<sub>2</sub>$  NMs in the 84 vicinity of plant roots.<sup>5-7</sup> Due to the electronic structure of  $CeO<sub>2</sub> NMS$ , the  $Ce(IV)/Ce(III)$ .

85 ratio at the NM surface is variable, while the core atoms are mainly in a Ce(IV) state.<sup>8</sup> 86 Partial Ce reduction of CeO<sub>2</sub> NMs was measured in several plant species such as tomato<sup>9</sup>, 87 wheat<sup>5</sup> and cucumber<sup>10</sup>. Owing to this versatile redox reactivity,  $CeO<sub>2</sub>$  NM is a good candidate to assess the NM biotransformation in soil. 

89 The root surface<sup>6, 9, 11, 12</sup> and root-soil interface<sup>7</sup> are the sites of interaction between plants 90 and NMs. Root metabolites excreted at these interfaces induce a reduction of  $CeO<sub>2</sub>$  NMs in 91 plant systems. $6, 13$  Plenty of metabolites are candidates (saccharides, amino acids, organic 92 acid, phenols, etc.), yet only citric acid and ascorbic acid were shown to reduce  $CeO<sub>2</sub> NMs$ , 93 through incubation experiments simulating root exudates.<sup>11</sup> Soil and root associated bacteria also release a large set of secondary metabolites in their environment for their nutrition (dissolution and uptake of iron, phosphate, etc.), and for their communication with other microorganisms and plants. For instance, biodissolution of Au NMs has been induced by the associated microbiome of a freshwater macrophyte, 98 despite the high chemical stability of Au NMs.<sup>14</sup> The main driver of Au biodissolution is its oxidation and complexation with a secondary metabolite, i.e. cyanide, emitted by 100 microorganisms.<sup>14</sup> Regarding CeO<sub>2</sub> NMs, their transformation by bacteria has been little addressed in the literature and the mechanism is poorly documented. Up to 50% of 102 reduced Ce occur in *Escherichia coli* and *Synechocystis* sp. exposed to 7 nm CeO<sub>2</sub> NMs.<sup>15, 16</sup> However, the soil bacterium *Bacillus subtilis* has proven inefficient in reducing Ce in rod- ceria 8.5 nm x.168 nm NMs, without the addition of exogenous ascorbic acid.<sup>17</sup> The 105 bacterial cell surface is the nano-bio interface that promotes the reduction of  $CeO<sub>2</sub> NMs, <sup>16</sup>$ , 

106 <sup>17</sup> however the mechanisms and the potential bacterial metabolites allowing this reduction are still unknown.

 What are the differences in terms of cell interface and secondary metabolism between bacteria that do (*E. coli* and *Synechocystis* sp., gram-negative bacteria) or do not reduce 110 (*B. subtilis*, gram-positive bacteria)  $CeO<sub>2</sub> NMS$ ? Apart from the structural differences in their outer envelope, Gram-negative and Gram-positive bacteria differ in the way they metabolize glucose, i.e. a carbon source for bacteria. Both take up glucose and metabolize it by phosphorylation to generate energy-rich molecules (e.g. ATP). However, Gram- negative bacteria can shut down this pathway and enable rapid conversion of glucose to 115 gluconic acid, and then to 2-keto-gluconic acid instead.<sup>18, 19</sup> Gluconic acid and 2-keto- gluconic acid can both be secreted at the outer bacterial membrane surface, and later internalized to feed the cell metabolism. Gluconic acid is the preferred substrate, and is 118 metabolized to pyruvic acid and then citric acid through the Krebs cycle.<sup>20</sup> 2-keto-gluconic acid is a reserve substrate for *Pseudomonas* sp. and known to accumulate in the 120 rhizosphere.<sup>21</sup> This ability to rapidly convert and subtract glucose from other species, and to secrete 2-keto-gluconic acid is a key feature involved in the ecological fitness of Gram-negative bacteria, such as *Pseudomonas* sp., a diverse, ubiquist and ecologically significant soil bacteria. Citric acid and 2-keto-gluconic acid are important metabolites 124 secreted by *Pseudomonas* sp. to solubilize phosphate<sup>22</sup> and chelate iron<sup>23</sup>. As they are released at the cell membrane interface, these two organic acids stand out as two 126 potential metabolite candidates for  $CeO<sub>2</sub>$  NM biotransformation. 

 Key physicochemical properties of NMs, such as their shape, size and surface charge, drive 128 their fate and their biological impact in the environment.<sup>24</sup> Among octahedral, cubic, rod, 129 and irregularly shaped  $CeO<sub>2</sub>$ -NMs, rod-like particles are transformed to a greater degree 130 than other  $CeO<sub>2</sub>$ -NMs in solutions simulating plant exudates in hydroponic cucumber 131 plants.<sup>25</sup> According to the particle shape and surface chemistry,  $CeO<sub>2</sub> NMs$  exhibits both pro- and antioxidant activities, with the former being toxic to *E. coli* <sup>26</sup>. The surface charge 133 of CeO<sub>2</sub> NMs has been shown to affect Ce(IV) biotransformation with a partial reduction 134 into Ce(III) in activated sludge bioreactors<sup>27</sup>, in the nematode *Caenorhabditis elegans*<sup>28</sup>, 135 and in wheat<sup>5</sup> and tomato<sup>9</sup>. The size and aggregation of  $CeO<sub>2</sub>$  NMs drive the microbial 136 activity and bacterial community structure in the canola rhizosphere.<sup>29</sup> Therefore, the 137 biotransformation of  $CeO<sub>2</sub>$  NMs depends on the NM properties but also on the tested 138 organisms and environmental conditions.<sup>28, 30, 31</sup> No marked relation between the size of CeO2 NMs and the extent of Ce reduction in the organisms has been noted to date. 

140 While predicted concentration of  $CeO<sub>2</sub>$  NMs in soils ranges from ng.kg<sup>-1</sup> to mg.kg<sup>-1 32-36</sup>, most studies addressing Ce biotransformation used concentrations close to or higher than 142 100 mg.kg<sup>-1,37</sup> One of the reasons for this is the technical difficulty of detecting, 143 quantifying and characterizing  $CeO<sub>2</sub>$  NMs at lower concentrations in complex matrices. Recent advances in X-ray absorption spectroscopy (XAS) such as high-energy resolution fluorescence detected XAS (HERFD-XAS) have enabled measurement of XANES spectra with higher energy resolution, i.e. with an apparent core hole lifetime broadening around 0.70eV, smaller than the probed electronic level bandwidth (around 3.2eV) and a lower 

 detection limit (as low as 60 mg Ce/kg). 38,39 The better defined edge and pre-edge features in the HERFD-XAS spectra enable clear detection of the presence of Ce(III) in the 150 structure of  $CeO<sub>2</sub>$ -NMs.<sup>39, 40</sup> It is essential to determine the oxidation state of  $CeO<sub>2</sub>$  NMs so 151 as to be able to identify the potential biotransformation of  $CeO<sub>2</sub>$  NMs, a process which 152 influences their bioavailability and potential ecotoxicity.<sup>13, 41, 42</sup>

 This study aimed at identifying the biotransformation mechanisms of different sizes and shapes of CeO2 NMs by *P. brassicacearum* NFM421, a phytobenefic soil bacteria isolated 156 from *Arabidopsis thaliana* and *Brassica napus* plants.<sup>43</sup> We investigated the fate of CeO<sub>2</sub> NMs in contact with *P. brassicacearum* cells in planktonic conditions, and in cell-free incubations with specific metabolites of citric acid and 2-keto-gluconic acid. Ascorbic acid was also tested as a positive control to assess Ce reduction. The bacterial metabolism and cell growth cycle are interconnected: different sets of metabolites are produced 161 throughout the three growth phases in planktonic conditions. We therefore explored  $CeO<sub>2</sub>$  NM reduction by *P. brassicarearum* at different steps of the growth cycle (i.e. the latence phase, where cells adapt to their environment, the exponential phase, where cells divide, and the stationary phase, where cells adapt to the nutrient reduction conditions), and we tested the secretome, i.e. all metabolites excreted by the bacteria in the late exponential phase. We used HERFD-XAS and were thus able to assess changes in Ce speciation and 167 identify the drivers of  $CeO<sub>2</sub>$  NMs biotransformation.

### **Materials and Methods**

### **CeO2 nanomaterials**

 

173 Three different sizes of commercially available types of bare  $CeO<sub>2</sub>$ -NMs were used in this study, all of which are made of cerianite crystallites. Transmission electron microscopy 175 indicated average sizes of 4 nm  $(CeO<sub>2</sub>-4$  nm) diameter with a spherical shape (Rhodia 176 Chemicals), 7 nm (CeO<sub>2</sub>-7 nm) with an ellipsoidal shape (Rhodia Chemicals), and 31 nm 177 (CeO<sub>2</sub>-31 nm) diameter with a spherical shape (Nanograin® Umicore). These NMs have 178 been previously described<sup>16, 29, 30, 39, 44-46</sup> and their characteristics are summarized in the Supporting Information (Table S1). The 7 nm NMs were an ellipsoid of revolution with an 180 average radius of the semi-axes of the ellipsoid of  $R = 3.5$  and 0.91 nm. The specific surface 181 areas were estimated from their size and shape:  $231 \text{ m}^2 \text{.g}^{\text{-}1}$  for CeO<sub>2</sub>-4 nm, 290 m<sup>2</sup>.g<sup>-1</sup> for 182  $\text{CeO}_2$ -7 nm and 30 m<sup>2</sup>.g<sup>-1</sup> for CeO<sub>2</sub>-31 nm. The surface atom percentages were calculated using the formula given in Deshpande *et al*. <sup>47</sup> using different lattice parameters (0.557 184 nm 0.553 and 0.547 nm for 4, 7 and 31 nm  $CeO<sub>2</sub> NMs)$ . We hypothesized that the outer 185 shell of atoms has a finite thickness of half the lattice constant.<sup>47</sup> The percentages of surface atoms were approximately 36%, 41% and 5.2% for the 4 nm, 7 nm and 32 nm forms, respectively). 

 **Bacteria and incubation conditions.**  

 The *Pseudomonas brassicacearum* strain NFM421, a phytobenefic soil-bacteria, was 

 selected for this study*.* The *Pseudomonas* genus has a robust metabolism for different 191 carbon sources, and a three-pronged metabolic system to metabolize glucose,<sup>18</sup> a bacterial carbon source that can be released in the rhizosphere from plant roots.<sup>48</sup> The *P. brassicacearum* genome is sequenced, thereby allowing metabolic pathway prediction.<sup>49</sup> Bacterial growth involves three different main phases: i) a lag phase, i.e. a nonreplication period during which the bacterial cells adapt; ii) an exponential phase, during which cell division occurs at a constant rate, with a generation time close to 55 min for *P. brassicacearum*; and iii) the stationary phase, where the growth rate levels off at zero, while the bacteria stay metabolically active and compensate for those that die.

 *P. brassicacearum* NFM421 from frozen stock (−80 °C) was grown at 30°C in 10-fold diluted tryptic soy broth (TSB 1/10) solidified with 15 g L−1 agar (Sigma) under dark 201 conditions. After 48 h, a single colony was resuspended in TSB 1/10, incubated overnight 202 at 30°C and stirred at 150 rpm in an orbital shaker, until a concentration of  $1 \times 10^9$  cells per mL in stationary phase was reached, and this was used as a pre-culture. 

 Aliquots of the pre-culture were transferred to a new TSB 1/10 medium in order to reach 205 an optical density at 600 nm (OD<sub>600</sub>) of 0.05 ( $\sim 2.10^6$  CFU/mL Figure SI 2). Bacteria and TSB/10 medium were transferred into 50 mL tubes or 1 L erlens, stirred at 150 rpm, 207 incubated until the early exponential growth phase was reached at  $OD_{600} = 0.1$  ( $\sim$ 4.10<sup>7</sup>) CFU/mL), after approximately 60 min. Several exposures of the *P. brassicacearum* soil 209 bacteria to different sized and shaped  $CeO<sub>2</sub>$  NMs were then performed: 

- 210 1) Exposure to the three sizes of  $CeO<sub>2</sub> NMs: 4 nm, 7 nm and 31 nm at 5 mg  $CeO<sub>2</sub>/L$ , from$ 211 the early- to the late-exponential phase (t3:  $OD_{600}$  0.7 (2.8  $10^8$  CFU/mL)). (See SI Figure SI 3).
- 213 2) Exposure to several  $CeO<sub>2</sub>$ -4 nm concentrations: 0.01-0.1-1-5-10 mg  $CeO<sub>2</sub>/L$ , from 214 the early- to the late-exponential phase (t3). Note that the tested concentration did not affect the bacterial growth.
- 216 3) Exposure to  $CeO<sub>2</sub>$ -7 nm at 5 mg  $CeO<sub>2</sub>/L$  with bacterial sampling at different times 217 during the exponential phase: t1)  $OD_{600}$  0.15, after 45 min of contact, which was less 218 than the generation time (cell division) for *P. brassicacearum*, t2)  $OD_{600}$  0.3, 90 min 219 mid-exponential phase and t3)  $OD_{600}$  0.7, 165 min, late-exponential phase.

220 For each size of  $CeO<sub>2</sub> NMS$ , a stock solution at 0.3 g Ce/L was prepared prior to bacterial 221 culture dosing. The bacterial culture was incubated at 30°C and stirred at 150 rpm. After 222 the bacterial growth phase, the suspensions were centrifuged at g at  $4^{\circ}$ C to separate 223 the bacteria and supernatant. The supernatants were discarded, and the bacterial pellets 224 were either immediately frozen in 5 mm diameter pellets in liquid  $N_2$  and stored at -80°C for XAS measurements, or freeze-dried for ICP-MS analysis. 

 The effect of the culture medium, single metabolites and secretome (i.e. all metabolites 227 excreted by the bacteria exposed or not to  $CeO<sub>2</sub>$  NMs) on the Ce reduction were tested. 228 For these analyses, only  $CeO<sub>2</sub>$ -4 nm was tested due to time constraints for the Synchrotron analyses and because the extent of bioreduction after bacterial exposure was similar for 

 4 nm and 7 nm (Figure 1). The culture medium (TSB 1/10) and several commercialized metabolites (2 keto-gluconic acid, ascorbic acid, citric acid from Sigma Aldrich) were 232 incubated with  $CeO<sub>2</sub>$ -4 nm. The culture medium (TSB 1/10) and citric acid were also 233 incubated with  $CeO<sub>2</sub>$ -7 nm. Metabolite solutions were prepared at 50 mM concentration, 234 and  $CeO<sub>2</sub>$  NMs at 300 mg  $Ce/L$  (2 mM) concentration. The metabolites were in excess 235 compared to  $CeO<sub>2</sub>$  NMs. Solutions were adjusted to pH=7.2. The suspensions were incubated at 25°C under agitation for 24 h. After the incubation, droplets of the 237 suspensions were frozen in liquid  $N_2$  and kept at -80 $\degree$ C before XAS analyses.

 To test the influence on the Ce reduction of the secretome, the bacterial culture 239 supernatant was incubated with  $CeO<sub>2</sub>$  NMs. We hypothesized that the bacteria could respond to stress when exposed to NMs responsible for specialized metabolite excretion. We compared two different secretomes: the supernatant of *P. brassicacearum* not 242 exposed to NMs, after a bacterial growth to an  $OD_{600}$  of 0.7 (same conditions as explained 243 above), and the supernatant of *P. brassicaceae* exposed to 5 mg.L<sup>-1</sup> CeO<sub>2</sub>-4 nm. After the bacterial growth, the suspensions were centrifuged to separate the bacteria and 245 supernatant. The 4-nm  $CeO<sub>2</sub>$  NMs were added to these supernatants at 300 mg  $Ce/L$  concentration and incubated under agitation for 24h. After the incubation, droplets of the 247 suspensions were frozen in liquid  $N_2$  and kept at -80 $\degree$ C before the XAS analyses. 

#### **ICP MS measurement**

 The Ce concentration was measured using ICP-MS (PerkinElmer, NexIon 300X) in 250 bacteria exposed to several  $CeO<sub>2</sub>$ -4 nm concentrations: 0.01-0.1-1-5-10 mg  $CeO<sub>2</sub>/L$ , from early- to late- exponential phases (t3). Bacterial pellets of three suspensions in 1 L-erlens were pooled to obtain sufficient material for the analysis (30 mg dry weight [DW]), no replicates were performed. Before ICP-MS analysis, each sample was dried and solubilized in acid using a Milestone UltraWAVE® microwave. The acids used for plant 255 sample digestion were 2 mL  $HNO<sub>3</sub>$  and 1 mL  $H<sub>2</sub>O<sub>2</sub>$ . The measurement quality was controlled using certified reference materials (mussel tissue BCR-668 from IRMM). All Ce concentrations presented are expressed in mg per kg of dried matter. 

#### **XAS measurements**

 Ce speciation was measured by X-ray absorption near edge structure (XANES) spectroscopy at the European Synchrotron Radiation Facility (ESRF, Grenoble, France). The samples were analyzed in high energy resolution fluorescence detected mode (HERFD) on the FAME-UHD beamline (BM16) using a crystal analyzer spectrometer (CAS) equipped with 5 Ge(331) bent crystals (1 m radius of curvature). With CAS, photons are first optically selected in energy (Bragg diffraction) by the crystals and later collected by a detector, thereby generating a very small energy bandwidth (typically 1-2eV) and 266 thus a higher detection limit and a better spectral resolution.<sup>38, 50</sup> To avoid beam damage, the frozen pellets were analyzed at 10 K using a liquid He cryostat. No photoreduction effect was observed when comparing the different scans. All of the XANES spectra shown here were the sum of 2 to 12 scans, depending on the Ce concentration, in order to 

 improve the signal-to-noise ratio. Normalization and data reduction were performed 271 using the IFFEFIT software package.<sup>51</sup> Initial CeO<sub>2</sub>-NMs, Ce(III)-acetate, Ce(III)-oxalate, 272 Ce(III)-phosphate and Ce(III)-cysteine were used as Ce(IV) and Ce(III) reference samples.

273 Quantitative determination of the Ce oxidation state was performed through a peak fitting analysis of the pre-edge features of the normalized spectra. To analyze the pre-edge features, a baseline subtraction was first performed by fitting a Gaussian function on the 276 low energy tail of the edge using data a few eV before and after the pre-edge peak.<sup>52, 53</sup> The pre-edges were then deconvoluted into Gaussian components (Pseudo-Voigt was also tested but did not improve the deconvolution performance). As a first step, the Gaussian 279 parameters were allowed to vary for the standard compounds. The Gaussian parameters were then fixed and applied to all spectra (see for example Figure 2C for a deconvolution 281 involving both Ce(III) and Ce(IV)). Two Gaussians functions were used to fit the Ce(III) contribution, with center max positions set at 5718.25 and 5720.87 eV and full widths at half maximum (FWHM) of 2.09 and 1.06 eV, respectively, while the area of each Gaussian function was set at 87% and 13% of the total area, respectively. The Ce(IV) contribution was fitted with one Gaussian with a center max position at 5720.12 eV and an FWHM of 286 1.77 eV. The pre-edge information was obtained by calculating the "integrated intensity" (sum of the integrated intensities of each component). We hypothesized that the intensity 288 of the pre-edge peak reflects the proportion of the Ce(III) and Ce(IV), however this implies several assumptions. We put forward the hypotheses that: (i) there is a linear relation between the peak integrated area and the Ce oxidation state, (ii) that the local symmetry does not markedly influence the intensity and position of the pre-edge features. It would be difficult to conclude on this last point as the Ce(III) species formed in this study were not identified. To assess the uncertainty of the pre-edge fitting procedure, we calculated theoretical pre-edge spectra for mixtures of Ce(III) and Ce(IV) reference compounds (Ce(III)+Ce(IV)=100% and Ce(III) ranged from 5 to 95%). The peak fitting analysis described above was applied to these mixtures and the relative uncertainty was 297 calculated. The uncertainty ranged from 3 to 22%. The results are thus all expressed with 22% uncertainty and presented as mean ± 0.22\*mean. 

 

# **Results and Discussion**

### **Input of high-energy resolution fluorescence-detected X-ray absorption spectroscopy (HERFD-XAS) for the study of Ce speciation in highly diluted biological samples**

# **Unequivocal identification of the Ce redox state in CeO2 NMs**

 XANES at the Ce L3-edge was sensitive to the Ce oxidation state with one intense white line typical of Ce(III) (feature A in Ce(III)-acetate, Figure 1) and two peaks for Ce(IV) 309 (features B and C in CeO<sub>2</sub> NMs, Figure 1) related to the Ce 5d electron density of states.<sup>40</sup> However, feature A, located at 5729 eV, could also be attributed to crystal field splitting 

311 (into  $e_g$  and  $t_{2g}$ ) for Ce(IV) in CeO<sub>2</sub> nanoparticles.<sup>38</sup> With HERFD-XANES, the pre-edge features clearly distinguished the contributions of Ce(III) and Ce(IV) to the signal (Figure 2b). The Ce(III) compound pre-edges showed two features, i.e. D and F, with maximum peaks at 5718 and 5721 eV, respectively, while the Ce(IV) compound pre-edges showed a single peak E with a maximum at 5720 eV. Peaks D and E were easily distinguishable with HERFD-XAS as they were separated by approx. 2eV. Regardless of the NM size, HERFD- XANES analysis solely revealed the presence of the Ce(IV) oxidation state (SI Table 2 ) in the CeO<sub>2</sub> NMs tested before interaction with bacteria. This result is in agreement with the 319 previously observed absence of Ce(III) inside CeO<sub>2</sub> NMs.<sup>40, 54</sup> Therefore we confirmed the 320 absence of Ce(III) for several sizes of  $CeO<sub>2</sub> NMs$ . 

 

 

### **Major advance in the speciation of highly diluted Ce bacteria**

324 In this study, we analyzed Ce in bacteria exposed to concentrations as low as 10 µg Ce/L. XAS measurement using CAS allows the study of diluted chemical elements and therefore realistic NM concentrations. The predicted Ce concentration in soil ranged from ng/kg to mg/kg, depending on the area and the soil usage, such as sludge-amended soil or urban soil<sup>32, 36, 55</sup>, but no predictions of environmental Ce concentrations in soil solution have been published to date. In aquatic media, the Ce predicted environmental concentration 330 ranged from 1 ng/L to 1 µg/L.<sup>36</sup> With bacterial cell concentrations ranging from 10<sup>4</sup> to 10<sup>6</sup> 331 CFU/mL, thereby representing an NM/cell ratio ranging from 4.8 to 4.8  $10^5$ . In this study, 332 the bacterial density ranged from 4.  $10^7$  to 2.8  $10^8$  CFU/mL, and the NM/cell ratio from 333 1.7 10<sup>2</sup> to 1.2 10<sup>6</sup> depending on the CeO<sub>2</sub> NM concentration exposure. Therefore, the NM/cell ratio tested in this study was very plausible compared to the environmental prediction. 

 

337 The Ce L3-edge XANES of bacteria exposed to  $CeO<sub>2</sub>$ -4nm ranged from 0.01 to 10 mg/L, indicating a dominant single absorption edge peak located at 5729 eV, which is typical of Ce(III) (Figure 3). The pre-edge peak analysis confirmed a major Ce(III) contribution of > 340 70  $\pm$  15% for 5 and 10 mg CeO<sub>2</sub>/L concentrations (Table SI 3). The Ce concentrations of these bacteria ranged from 6.2 to 1870 mg Ce/kg DW (Table 1). To date, the lowest Ce concentration analyzed by bulk Ce L3 edge XANES in environmental samples was 35 mg Ce/kg DW.<sup>11</sup> HERFD-XANES analysis therefore improved the detection limit, since XANES were recorded for concentrations as low as 6.2 mg Ce/kg DW. To our knowledge, this is the first time that Ce speciation has been detected at concentrations below 10 mg/kg DW in biological samples. Note that the bacterial samples were analyzed frozen and hydrated (> 90 % water content), therefore with lower exposure Ce concentrations of around 1 mg/kg dried or freeze-dried samples could be used for the analyses. 

### **Size-dependent biotransformation of Ce NMs**

 The bacterial growth cycle comprises three main phases, each with a specific physiological profile. In the lag phase, just after introduction in the fresh medium, cells sense the environment and retool the machinery in preparation for the exponential phase, 

 during which they actively divide, depending on the energy sources and nutrients. At the late-exponential phase, as the energy sources are depleted and metabolic waste products accumulate, cells enter in the stationary phase, where growth apparently ceases, with equivalent rates of bacterial cell growth and death. The late exponential phase is an interesting step in the bacterial growth cycle. Secondary metabolism involved in microbial and ecological interactions tends to occur at these late exponential and stationary bacterial growth phases.<sup>56</sup> 

 The speciation of the Ce accumulated in bacteria after exposure to three different sizes of CeO<sub>2</sub> NMs was measured in the late-exponential phase (Figure 1). We obtained contrasted 364 results between bacteria exposed to 31 nm-spherical CeO<sub>2</sub> NMs and to the two small sizes: 365 4 nm-spherical and 7 nm-ellipsoidal. XANES spectra of bacteria exposed to 31 nm  $CeO<sub>2</sub>$ 366 NMs overlapped the spectra of initial  $CeO<sub>2</sub>$  NMs, indicating that the atomic structure of 367 the CeO<sub>2</sub> NMs was not markedly affected. XANES spectra of bacteria exposed to CeO<sub>2</sub>-31 nm were fitted with 100% Ce(IV) using Ref 31 nm\_CeO2 (see Table SI 2). XANES spectra of bacteria exposed to 4 nm and 7 nm showed a dominant absorption edge (feature A, Figure 1) typical of the presence of Ce(III). The reduction was similar between the two 371 sizes: pre-edge peak fitting indicated a contribution of Ce(III) of 70  $\pm$  15% for 4 nm CeO<sub>2</sub> 372 NMs and  $64 \pm 14\%$  for 7 nm CeO<sub>2</sub> NMs at 5 mg Ce/L (Table SI 3). To avoid a potential Ce 373 reduction due to the composition of the culture medium with 1.4 mM  $K_2HPO_4$ , CeO<sub>2</sub>-4 nm was incubated in sterilized TSB/10 medium for 24 h and no Ce reduction was detected by 375 XAS (Figure 4B). Our results showed that the 4 and 7 nm  $CeO<sub>2</sub>$  NMs were biotransformed whereas 31 nm were not modified after bacterial exposure. 

 Several physicochemical mechanisms could explain the differences in Ce reduction for 379 different  $CeO<sub>2</sub> NMs$ , such as aggregation, specific surface area and reactivity.<sup>1, 8, 28, 57</sup> The aggregation state in the media could influence the amount of surface-exposed NMs to bacterial cells. Yan *et al.* suggested that the reduction rates of iron oxide nanoparticles of different sizes (10, 30 and 50 nm) by *Geobacter sulfurreducens* was directly correlated with the bacteria-hematite contact area, while taking NM aggregation into account, not 384 with the total oxide surface area.<sup>58</sup> The isoelectric points (IEP) of  $CeO<sub>2</sub>$  NMs, and the pH and ionic strength of the culture media suggested that aggregation could also be expected 386 for the three types of  $\text{CeO}_2$  NMs. Since only the 4 and 7 nm  $\text{CeO}_2$  NMs were bioreduced, no simple correlation was noted in this study between the aggregation state and the reduction rate. 

 

 The Ce(IV) reduction could also be attributed to the redox cycle occurring at the surface 391 of the solid and/or to reductive dissolution.<sup>59</sup> The percentage of surface atoms for the 31- nm NMs was around 5% (Table SI1), whereas the uncertainty of the HERDF-XAS method 393 was  $\pm$  22%. This could explain why Ce(III) was not detected at the surface of 31 nm CeO<sub>2</sub> 394 NMs. For the 4 nm and 7 nm CeO<sub>2</sub> NMs, the percentage of Ce(IV) reduction (>  $64\pm14\%$ , Table SI3) was higher than that of surface atoms (>36 %, Table SI1). We suggest that this reduction could be explained by reductive dissolution. Indeed, Ce(III) has high solubility, 

 with logKs = -21.5 for Ce(III) oxide (Smith, 2004), and its release rate could be expected 398 to exponentially increase as the grain size decreases.<sup>31, 60</sup> To conclude, we hypothesized that the absence, or low Ce reduction, for the 31 nm NMs could be explained by a low reductive dissolution rate due to a lower percentage of surface atoms compared to NMs of under 7 nm.

 

#### **Biotransformation kinetics**

 Metabolism is dependent on the phase of the growth cycle. Different sets of metabolites are produced in the primary metabolism, related to cell growth and division *vs.* the secondary metabolism, in the late-exponential and the stationary phases, related to ecological function.<sup>56</sup> To determine the metabolic context in which Ce reduction takes place, we investigated the NM biotransformation kinetics by sampling cells after different 409 exposure times to  $CeO<sub>2</sub>$ -4 nm and  $CeO<sub>2</sub>$ -7 nm, from the early-exponential to the stationary phase. Figure 2 shows that for each tested exposure time, a dominant post-absorption edge peak at 5729 eV, characteristic of Ce(III), was observed on all XANES, after exposure 412 to  $CeO<sub>2</sub>$ -7 nm. The pre-edge peak fitting analysis indicated similar results for all exposure times: 68 ± 15% Ce(III) after 45 and 90 min and 64 ± 14% Ce(III) after 320 min in bacteria 414 exposed to 7 nm CeO<sub>2</sub> (Table SI3). Similar results were obtained after exposure to CeO<sub>2</sub>- 4 nm: 76 ± 17% Ce(III) after 45 min and 70 ± 15% after 90 and 320 min (Table SI3). 

 Therefore, these results indicated that a partial reduction occurred rapidly (in less than 45 min) in the early-exponential phase and did not vary up to the stationary phase. Interestingly, the conversion of glucose to gluconic acid and to 2-keto-gluconic acid by *Pseudomonas* occurred from the lag to the early exponential phases, which suggests a 420 potential role of these metabolites in the reduction of  $CeO<sub>2</sub> NMs$ . 

 

#### **Secretome and interaction between bacteria and NMs**

 The secretome, i.e. the culture medium sampled after bacterial growth, including all the metabolites excreted by the bacteria, was isolated from the bacterial culture. Figure 4B 425 shows XANES spectra of  $CeO<sub>2</sub>$ -4 nm incubated in the bacterial secretome, sampled from a 426 control or from a  $CeO<sub>2</sub>$ -4 nm exposure sample. The spectra were similar between the NM suspension and after incubation in the two scenarios. The pre-edge analysis confirmed the absence of a Ce(III) contribution (Table SI 3). The secretome therefore did not induce a Ce reduction after incubation with NMs. This indicates that the NM/bacterial interaction 430 is needed to induce a Ce reduction.  $CeO<sub>2</sub>$  NMs can be adsorbed at the surface of the bacteria, as shown with *Escherischia coli* in the studies of Thill et al. <sup>16</sup> and Zeyons et al.<sup>15</sup> . In the close vicinity of the bacteria, the chemical composition of the media was highly influenced by the bacterial exudation. Locally, the metabolite concentration was much higher than in the sampled secretome. Moreover, the metabolites could also be rapidly degraded in the media. Therefore, the bacterial metabolites in the secretome may not have induced a Ce reduction due to the low concentrations in the media and probable biodegradation. 

#### **Reduction of CeO2 NMs is driven by 2-keto-gluconic acid and citric acid bacterial metabolites**

 To test the involvement of individual metabolites secreted in the extracellular medium in the Ce reduction, we focused on environmentally relevant metabolites that are known to be secreted in soil, especially in the rhizosphere, as *P. brassicacearum* NFN 421 is a good plant root colonizer.<sup>43</sup> We selected 2-keto-gluconic acid and citric acid bacterial metabolites as potential sources of Ce reduction. As mentioned, these organic acids are produced by *Pseudomonas* sp., and more generally by Gram-negative bacteria, in their environment so as to retrieve phosphate, iron and interact with plants. *P. brassicacearum* 448 contains all the genes necessary for the production and secretion of these metabolites <sup>49</sup>. 2-keto-gluconic acid can be detected in the culture medium of *P. brassicacearum* in the early exponential phase (C. Santaella, personal communication). 

 CeO<sub>2</sub> NMs were incubated with 2-keto-gluconic acid, citric acid and ascorbic acid (which is not produced by bacteria), as ceria NMs can undergo a reduction with the assistance of 453 these organic acids.<sup>11, 17</sup> All of the tested metabolites induced a partial reduction of  $CeO<sub>2</sub>$ - 4 nm after 24h exposure (Figure 4A). Peak fitting of the pre-edge confirmed the presence 455 of Ce(III) in these samples with  $23 \pm 5\%$  for citric acid,  $30 \pm 7\%$  for 2-keto-gluconic acid 456 and 26  $\pm$  6% for ascorbic acid for CeO<sub>2</sub>-4 nm, and 31  $\pm$  7 % for ascorbic acid for CeO<sub>2</sub>-7 nm (Table SI 4). Note that the bacterial metabolite 2-keto-gluconic acid proved to be as 458 efficient as ascorbic acid, the main plant metabolite currently found to be involved in  $CeO<sub>2</sub>$  NMs reduction. As *Pseudomonas* species are ubiquituous bacteria that are frequently present in hydroponic plant cultures, and in the rhizosphere, this suggests the 461 involvement of 2-keto-gluconic acid in  $CeO<sub>2</sub>$  NM reduction at the plant root interface. 462 Around 36% of Ce atoms are localized at the surface of CeO<sub>2</sub>-4 nm, and 41% for the CeO<sub>2</sub>-7 nm. Assuming that the Ce reduction only occurred at the surface, the reductions induced 464 by 2-keto-gluconic acid, citric acid and ascorbic acid for CeO<sub>2</sub>-4 nm were 83  $\pm$  19%, 65  $\pm$ 465 14% and 72  $\pm$  17% of the surface atoms. For the CeO<sub>2</sub>-7 nm, ascorbic acid induced a 466 reduction of  $75 \pm 17\%$  in the surface atoms. The metabolites alone, when present in excess 467 in the suspension, were thus able to induce a marked reduction in  $CeO<sub>2</sub> NMS$  of the surface atoms. The reduction measured with the tested metabolites remained lower than in the biotic interaction. This could be explained by the implication of other metabolites not tested in this study and the effect of the colocalization of NMs and metabolites at the nano- cell interface. 

 Overall, our results showed that the biotransformation and fate of NMs was size- dependent, and metabolites such as citric acid and 2-keto-gluconic acid proved efficient in reducing Ce from CeO2-4 nm. Many soil bacteria other than *Pseudomonas* sp. can secrete 2-keto-gluconic acid,<sup>61</sup> e.g. *Gluconobacter oxidans.* This suggests that in environmentally relevant conditions, the presence of soil microbiota associated with plant roots could 478 contribute to the reduction of  $CeO<sub>2</sub> NMS$ . This biotransformation will affect the fate of Ce in the soil and the effect of Ce on soil biota (e.g. bacteria, fungi, macroorganisms and plants). 

 Beyond the fate of NMs in soil, this is an important result to take into consideration for 482 diverse NM applications because the main properties of  $CeO<sub>2</sub>$  NMs will be impacted by 483 biotransformation. For example, it is often speculated that  $CeO<sub>2</sub>$  NMs present an autoregenerative mechanisms of valence reversion, thereby explaining the catalytic 485 properties.<sup>41</sup> Interesting, their ability to self-regenerate their surface has biomedical applications.<sup>42</sup> However, if NMs are transformed or if the reduction is total, as observed 487 in our study, it is more than likely that Ce(III) would be released from  $CeO<sub>2</sub> NMs$ , thereby making the valence reversion impossible in NMs. Note that *Pseudomonas* sp. also include 489 human pathogens such as *P. aeruginosa* or *P. putida*. Studies using CeO<sub>2</sub> NMs should therefore investigate possible contact of NMs with bacteria and adapt the NM size to take both the modification of NM properties and their environmental fate into account. 

 **Supporting Information:** Physicochemical characteristics of nanomaterials; Correlation between the optical density at 600 nm and colony forming units; *Pseudomonas*  <sup>495</sup> *brassicacearum* growth curve; Normalized Ce L<sub>3</sub>-edge HERFD-XANES of CeO<sub>2</sub>-4 nm, CeO<sub>2</sub>- $\,$  7 nm, CeO<sub>2</sub>-31 nm and zoom on the pre-edge region; Pre-peak fitting parameters of the standard compounds; Results of the pre-peak fitting of the samples.

### **Acknowledgements**

 The authors thank ESRF for providing beam time access on FAME-UHD/BM16 (experiment EV-202). The FAME-UHD project is financially supported by the French *Grand Emprunt* EquipEx (EcoX, ANR-10-EQPX-27-01), the CEA-CNRS CRG consortium and the INSU CNRS Institute. This work is also a contribution to the OSU-Institut Pythéas. The authors thank CNRS for the IRP iNOVE funding.

- 
- 

- 
- 





 Transformation, and Impacts of Ceria Nanomaterials in Aquatic Mesocosms Simulating a Pond Ecosystem. *Environ. Sci. Technol.* **2014,** *48*, (16), 9004-9013. 31. Misra, S. K.; Dybowska, A.; Berhanu, D.; Luoma, S. N.; Valsami-Jones, E., The complexity of nanoparticle dissolution and its importance in nanotoxicological studies. *Science of The Total Environment* **2012,** *438*, 225-232. 32. Park, B.; Donaldson, K.; Duffin, R.; Tran, L.; Kelly, F.; Mudway, I.; Morin, J. P.; Guest, R.; Jenkinson, P.; Samaras, Z.; Giannouli, M.; Kouridis, H.; Martin, P., Hazard and risk assessment of a nanoparticulate cerium oxide-based diesel fuel additive - A case study. *Inhal. Toxicol.* **2008,** *20*, (6), 547-566. 33. Johnson, A. C.; Park, B., Predicting contamination by the fuel additive cerium oxide engineered nanoparticles within the United Kingdom and the associated risks. *Environ Toxicol Chem* **2012,** *31*, (11), 2582-7. 34. Gottschalk, F.; Lassen, C.; Kjoelholt, J.; Christensen, F.; Nowack, B., Modeling Flows and Concentrations of Nine Engineered Nanomaterials in the Danish Environment. *International Journal of Environmental Research and Public Health* **2015,** *12*, (5), 5581. 35. Meesters, J. A. J.; Quik, J. T. K.; Koelmans, A. A.; Hendriks, A. J.; van de Meent, D., Multimedia environmental fate and speciation of engineered nanoparticles: a probabilistic modeling approach. *Environmental Science: Nano* **2016,** *3*, (4), 715-727. 36. Giese, B.; Klaessig, F.; Park, B.; Kaegi, R.; Steinfeldt, M.; Wigger, H.; von Gleich, A.; Gottschalk, F., Risks, Release and Concentrations of Engineered Nanomaterial in the Environment. *Scientific Reports* **2018,** *8*, (1), 1565. 37. Layet, C.; Auffan, M.; Santaella, C.; Chevassus-Rosset, C.; Montes, M.; Ortet, P.; Barakat, M.; Collin, B.; Legros, S.; Bravin, M. N.; Angeletti, B.; Kieffer, I.; Proux, O.; Hazemann, J.-L.; Doelsch, E., Evidence that Soil Properties and Organic Coating Drive the Phytoavailability of Cerium Oxide Nanoparticles. *Environ. Sci. Technol.* **2017,** *51*, (17), 9756-9764. 38. Proux, O.; Lahera, E.; Del Net, W.; Kieffer, I.; Rovezzi, M.; Testemale, D.; Irar, M.; Thomas, S.; Aguilar-Tapia, A.; Bazarkina, E. F.; Prat, A.; Tella, M.; Auffan, M.; Rose, J.; Hazemann, J.-L., High-Energy Resolution Fluorescence Detected X-Ray Absorption Spectroscopy: A Powerful New Structural Tool in Environmental Biogeochemistry Sciences. *J. Environ. Qual.* **2017,** *46*, (6), 1146-1157. 39. Tella, M.; Auffan, M.; Brousset, L.; Morel, E.; Proux, O.; Chaneac, C.; Angeletti, B.; Pailles, C.; Artells, E.; Santaella, C.; Rose, J.; Thiery, A.; Bottero, J. Y., Chronic dosing of a simulated pond ecosystem in indoor aquatic mesocosms: fate and transport of CeO2 nanoparticles. *Environmental Science: Nano* **2015,** *2*, (6), 653-663. 40. Cafun, J. D.; Kvashnina, K. O.; Casals, E.; Puntes, V. F.; Glatzel, P., Absence of Ce3+ Sites in Chemically Active Colloidal Ceria Nanoparticles. *ACS Nano* **2013,** *7*, (12), 10726- 10732. 41. Pulido-Reyes, G.; Rodea-Palomares, I.; Das, S.; Sakthivel, T. S.; Leganes, F.; Rosal, R.; Seal, S.; Fernandez-Pinas, F., Untangling the biological effects of cerium oxide nanoparticles: the role of surface valence states. *Sci Rep* **2015,** *5*, 15613. 42. Rajeshkumar, S.; Naik, P., Synthesis and biomedical applications of cerium oxide nanoparticles–a review. *Biotechnology Reports* **2018,** *17*, 1-5. 43. Achouak, W.; Sutra, L.; Heulin, T.; Meyer, J.-M.; Fromin, N.; Degraeve, S.; Christen, R.; Gardan, L., Pseudomonas brassicacearum sp. nov. and Pseudomonas thivervalensis sp. nov., two root-associated bacteria isolated from Brassica napus and Arabidopsis thaliana. *International Journal of Systematic and Evolutionary Microbiology* **2000,** *50*, (1), 9-18. 

 44. Auffan, M.; Tella, M.; Santaella, C.; Brousset, L.; Paillès, C.; Barakat, M.; Espinasse, B.; Artells, E.; Issartel, J.; Masion, A., An adaptable mesocosm platform for performing integrated assessments of nanomaterial risk in complex environmental systems. *Scientific reports* **2014,** *4*, (1), 5608. 45. Auffan, M.; Rose, J.; Orsiere, T.; De Meo, M.; Thill, A.; Zeyons, O.; Proux, O.; Masion, A.; Chaurand, P.; Spalla, O.; Botta, A.; Wiesner, M. R.; Bottero, J. Y., CeO2 nanoparticles induce DNA damage towards human dermal fibroblasts in vitro. *Nanotoxicology* **2009,** *3*, (2), 161-U115. 46. Nabavi, M.; Spalla, O.; Cabane, B., Surface Chemistry of Nanometric Ceria Particles in Aqueous Dispersions. *Journal of Colloid and Interface Science* **1993,** *160*, (2), 459-471. 47. Deshpande, S.; Patil, S.; Kuchibhatla, S.; Seal, S., Size dependency variation in lattice parameter and valency states in nanocrystalline cerium oxide. *Applied Physics Letters* **2005,** *87*, (13), 3. 48. Stubbs, V.; Standing, D.; Knox, O.; Killham, K.; Bengough, A.; Griffiths, B., Root border cells take up and release glucose-C. *Annals of Botany* **2004,** *93*, (2), 221-224. 49. Ortet, P.; Barakat, M.; Lalaouna, D.; Fochesato, S.; Barbe, V.; Vacherie, B.; Santaella, C.; Heulin, T.; Achouak, W., Complete genome sequence of a beneficial plant root- associated bacterium, Pseudomonas brassicacearum. **2011,** *193*, (12), 3146-3146. 50. Llorens, I.; Lahera, E.; Delnet, W.; Proux, O.; Braillard, A.; Hazemann, J. L.; Prat, A.; Testemale, D.; Dermigny, Q.; Gelebart, F.; Morand, M.; Shukla, A.; Bardou, N.; Ulrich, O.; Arnaud, S.; Berar, J. F.; Boudet, N.; Caillot, B.; Chaurand, P.; Rose, J.; Doelsch, E.; Martin, P.; Solari, P. L., High energy resolution five-crystal spectrometer for high quality fluorescence and absorption measurements on an x-ray absorption spectroscopy beamline. *Review of Scientific Instruments* **2012,** *83*, (6), 9. 51. Ravel, B.; Newville, M., ATHENA, ARTEMIS, HEPHAESTUS: data analysis for X-ray absorption spectroscopy using IFEFFIT. *J Synchrotron Radiat* **2005,** *12*, (Pt 4), 537-41. 52. Bordage, A.; Balan, E.; de Villiers, J. P.; Cromarty, R.; Juhin, A.; Carvallo, C.; Calas, G.; Raju, P. S.; Glatzel, P., V oxidation state in Fe–Ti oxides by high-energy resolution fluorescence-detected X-ray absorption spectroscopy. *Physics and Chemistry of Minerals*  **2011,** *38*, (6), 449-458. 53. Wilke, M.; Farges, F.; Petit, P.-E.; Brown Jr, G. E.; Martin, F., Oxidation state and coordination of Fe in minerals: An Fe K-XANES spectroscopic study. *American Mineralogist* **2001,** *86*, (5-6), 714-730. 54. Plakhova, T. V.; Romanchuk, A. Y.; Butorin, S. M.; Konyukhova, A. D.; Egorov, A. V.; Shiryaev, A. A.; Baranchikov, A. E.; Dorovatovskii, P. V.; Huthwelker, T.; Gerber, E.; Bauters, S.; Sozarukova, M. M.; Scheinost, A. C.; Ivanov, V. K.; Kalmykov, S. N.; Kvashnina, K. O., Towards the surface hydroxyl species in CeO2 nanoparticles. *Nanoscale* **2019,** *11*, (39), 18142-18149. 55. Liu, H. H.; Cohen, Y., Multimedia Environmental Distribution of Engineered Nanomaterials. *Environ. Sci. Technol.* **2014,** *48*, (6), 3281-3292. 56. O'Brien, J.; Wright, G. D., An ecological perspective of microbial secondary metabolism. *Current opinion in biotechnology* **2011,** *22*, (4), 552-558. 57. Collin, B.; Auffan, M.; Johnson, A. C.; Kaur, I.; Keller, A. A.; Lazareva, A.; Lead, J. R.; Ma, X.; Merrifield, R. C.; Svendsen, C.; White, J. C.; Unrine, J. M., Environmental release, fate and ecotoxicological effects of manufactured ceria nanomaterials. *Environmental Science: Nano* **2014,** *1*, (6), 533-548. 58. Yan, B.; Wrenn, B. A.; Basak, S.; Biswas, P.; Giammar, D. E., Microbial Reduction of Fe(III) in Hematite Nanoparticles by Geobacter sulfurreducens. *Environ. Sci. Technol.*  **2008,** *42*, (17), 6526-6531. 





**Figure 2 – A) Normalized Ce L3-edge high energy resolution fluorescence detected (HERFD)** 

 **XANES of** *Pseudomonas brassicacearum* **exposed to CeO2-7 nm at 5 mg Ce.L-1 for several**  726 incubation times: t1 ( $OD_{600}$  0.15) 45 min, t2 ( $OD_{600}$  0.3) 90 min, t3 ( $OD_{600}$  0.7) 320 min .<br>727 Ce(III)-acetate, and the CeO<sub>2</sub>-7 nm are shown as reference compounds (BM13, ESRF), Spectra **Ce(III)−acetate, and the CeO2-7 nm are shown as reference compounds (BM13, ESRF). Spectra t3 (OD600 0.7) are the same as the spectra shown in Figure 1 of bacteria exposed to CeO2-7 nm. B) Zoom of the pre-edge region for the reference compound and** *Pseudomonas brassicacearum* **exposed to CeO2-7 nmat 5 mg Ce.L-1 for 45 min t1 (OD600 0.15). The Ce(III) compound pre-edges showed two features, i.e. D and F, with maximum peaks at 5718 and 5721 eV, respectively, while the Ce(IV) compound pre-edges showed a single peak E with a maximum at 5720 eV. C) Extracted pre-edge of** *Pseudomonas brassicacearum* **exposed to CeO2-7 nmat 5 mg Ce.L-1 for 45 min t1 (OD600 0.15) and the corresponding pre-edge peak fitting after baseline subtraction. Two Gaussian functions fitted the Ce(III) and one fitted the Ce(IV).**

- 
- 



737 **Figure 3 - Normalized Ce L<sub>3</sub>-edge high energy resolution fluorescence detected (HERFD)**<br>738 **XANES of** *Pseudomonas brassicacearum* **exposed to CeO<sub>2</sub>-4 nm at several concentrations XANES of** *Pseudomonas brassicacearum* **exposed to CeO2-4 nm at several concentrations from 0.01 to 10 mg Ce .L-1 . Ce(III)−acetate, and the initial CeO<sup>2</sup> NMs are shown as reference compounds (BM13, ESRF).** 

 



 **XANES of CeO2-4 nm and different incubations (A) incubations of CeO2-4 nm with the growth culture medium (TSB1/10) and several metabolites: ascorbic acid, citric acid, 2-keto-gluconic acid (B) incubations of CeO2-4 nm NMs with the secretome: culture medium (TSB1/10) after 24 h of** *Pseudomonas brassicacearum* **growth, exposed or not to CeO2-4 nm during growth.**

 

 

 

- 
- 



#### **Table 1: Ce concentration (mg.kg-1 dry weigh) in bacteria** *Pseudomonas brassicacearum* **exposed at several CeO2-4 nm concentrations from the early- to late- exponential phase nanoparticles OD600 0.7 (2.8 10<sup>8</sup> CFU/mL).**



